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## **FINAL REPORT**

**TITLE:** Feasability studies for composting paper mill waste

**SPONSOR:** Weyerhaeuser

**CONTRACT NO.** FR37270-DJS1

**DURATION:** June 1, 1996 to October 31, 1997

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## TABLE OF CONTENTS

Subject	Page
Abstract	1
A. Bench-Top Feasibility Study	2
Introduction	2
Materials and Methods	2
Results	4
Composting Studies	4
Anaerobic Studies	9
Table 1A-Description of Aerobic Preparations	14
Table 1B-Description of Anaerobic Preparations	17
Table 2-Data Summary of Aerobic Composting Studies	19
Plots of %CO <sub>2</sub> Produced in Aerobic Reactors	20-30
Table 3-Summary of Results of Anaerobic Reactors	31
Table 4-Calculations of Results of Anaerobic Reactors	32
Table 5-CHN Analyses	33
Table 6- Organism Counts	34
Plots of Pressure Produced Daily	35-46
Plots of Moles Gas Produced	47-65
Plots of Moles-% of the Gases	66-80
Appendix	81
B. Testing the Composting of Paper Waste (to be submitted separately by K. C. Das)	



## ABSTRACT

Paper mill waste generated at the Weyerhaeuser, Flint River, plant can be effectively composted under controlled conditions. The mix of paper mill waste requires a moisture content of 40-55%, a pH below 7.5 to allow for maximum mold formation, relatively fresh primary sludge from the primary clarifier which provides active biological cultures, a mixture content with sufficient bulking to allow for air filtration, and a supplement of nitrogen compounds.  $\text{KNO}_3$  and  $\text{NH}_4\text{NO}_3$  were the most effective supplements for composting the waste with the least effective being Urea. The addition of other supplements such as Chicken Litter,  $\text{NH}_4\text{H}_2\text{PO}_4$ , and  $\text{NH}_4\text{Cl}$  and combinations were less effective. The additions of mineral salts and commercial additives advertised or reported to stimulate composting were ineffective in this study. It was concluded from the bench-top experiments that under suitable conditions 20% of the paper mill waste can be mineralized within a 30 day cycle.

The potential for anaerobically converting paper mill waste to  $\text{CH}_4$  and  $\text{CO}_2$  was demonstrated with as much as 60% of the total raw material mineralized. The supplements required for this conversion in a 40-60 day cycle was  $\text{NH}_4\text{NO}_3$  or  $\text{NH}_4\text{Cl}$  or a combination of  $\text{KNO}_3$  and  $\text{NH}_4\text{Cl}$  in sequence. A starter culture was essential for rapid start-up.  $\text{NH}_4\text{H}_2\text{PO}_4$ ,  $\text{KNO}_3$ , Urea or Chicken Litter supplement stimulated  $\text{CO}_2$  production with relatively small quantities of  $\text{CH}_4$ . The addition of  $\text{NH}_4\text{SO}_4$  resulted in sulfur being reduced.  $\text{N}_2\text{O}$  and traces of  $\text{NO}$  were observed in all systems containing  $\text{NO}_3^-$  as a supplement.



## FINAL REPORT

### A. BENCH TOP FEASIBILITY STUDY

#### INTRODUCTION

The thrust of the research study had two goals. One of them was to reduce the quantity of waste products accumulated daily at the Flint River site. The other was to test the feasibility of converting waste into a product that could be returned to the earth as a soil amender. The principal method to be utilized was traditional composting. The principal target was to measure the degree of mineralization that could occur in a reasonable time period and to evaluate the quality of the product.

During the study, it was decided that anaerobic methods should also be tested for the potential for mineralization of the wood waste as well as for methane gas production. These experiments were initiated six months after the start of the program.

#### MATERIALS AND METHODS

Materials: The waste (raw) materials available for this study are summarized in Appendix 1. Screening Knots were included in pilot studies but are no longer a major waste product at the Plant and therefore was excluded. Fly Ash was included but decidedly limited in its applicability for supporting microbial growth because of high pH values of around 12. Also, there was the concern of too high of an ash content causing metal shock to microbial cells. Therefore, boiler fly ash was used as a 1% enrichment and usually required pH adjustment with acid prior to its addition. Lime mud was not used in the study because of its high pH property and the nature of its composition that would be detrimental to microbial activity unless greatly diluted.

Primary sludge, bark and grit were chosen as the test materials because of its high organic carbon content. Primary sludge contained significant quantities of sand. Depending on the lot of Primary sludge collected at the Flint River site, the sand content varied significantly; accordingly, the ash contents recorded in the results section reflects this variation. The bark and grit were used as both the bulking agents and reactant but with the primary sludge representing the principal reactant. The volume ratios of the principal reactants used in preparing the mixtures were determined volumetrically in 1 and 5 gallon buckets and then weighed. All other additives were measured gravimetrically.

Since the starting materials are largely woody in nature (cellulosic, hemicellulosic), the nitrogen contents were virtually nonexistent. The nitrogen supplements used in the study were chicken litter from chicken ranches in the Flint River vicinity,  $\text{NH}_4\text{NO}_3$ ,  $\text{NH}_4\text{Cl}$ ,  $\text{NH}_4\text{H}_2\text{PO}_4$ , Urea,  $\text{KNO}_3$ ,  $\text{NH}_4\text{SO}_4$  or combinations thereof. Although the starting materials were relatively high in



inorganics (Appendix 2), complex mineral salts commonly used to support microbial growth was tested as supplements. The Mineral Media consisted of  $\text{KH}_2\text{PO}_4$  - 4.5g;  $\text{Na}_2\text{HPO}_4$  - 3g;  $\text{KNO}_3$  - 5g;  $\text{MgSO}_4$  - 0.25g;  $\text{H}_3\text{BO}_3$  - 2.9g;  $\text{MnCl}_2$ -4aq - 1.8mg;  $\text{ZnSO}_4$ -7aq - 0.22mg;  $\text{CuSO}_4$ -5aq - 0.08mg;  $\text{H}_2\text{MoO}_4$ -1aq - 0.015mg; and  $\text{FeCl}_3$  - 0.0025mg per L volume of reactants.

**Reactor designs:** The aerobic reactors were made of clear plexiglass configured rectangularly with a removal lid that was bolted to the assemblage to create a water and air tight reactors. The reactors were equipped with sampling ports, air injection system, exhaust ports, air flow rate meters and temperature probes. Two sizes were used in this study. The largest had a sixty-liter capacity (18 cm x 18 cm x 153 cm) and the others were of 20 liter capacities (18 cm x 18 cm x 65 cm). Each reactor contained a metal grid positioned 10 cm above the base plate to create a plenum for air delivery in an upward flow through the reactor. The aeration rate was maintained between 5 - 7.2  $\text{m}^3/\text{lb}/\text{day}$  of compost mix. Since the air flow fluctuated and the volume of the reactant reduced in volume relative to biological activity, the quantity of  $\text{CO}_2$  produced in each reactor was normalized by multiplying the amount of  $\text{CO}_2$  in the gas sample by 0.1 of the flow rate volume.

The anaerobic reactors consisted of 2, 4 and 6 L wide-mouth Erlenmeyer-type flasks and 3 L Fernbach flasks with stoppers containing a septum for gas sampling and ports for connection to an open-ended Hg tube for recording gas pressures and for the daily release of the gas pressures and/or a graded inverted cylinder filled with water to measure gas displacement of the water in the cylinder. Each of the reactor vessels contained water with the compost mix in a slurry state. Two of the 20 L reactors used in the aerobic study were also employed in the anaerobic study but as a solid phase. In these modes, the inflow and outflow air ports were plugged with serological septa. In some cases, the reactors were filled and sealed whereas in others they were flushed with  $\text{N}_2$  before plugging the ports.

**Inoculation:** No extrinsic inocula were used. All microbial activity was by *in situ* cultures. For the anaerobic studies, however, the standard mix of reactants were slurried without nitrogen additive and incubated at room temperature. These slurries were used as seed cultures. The ideal seed cultures are those obtained from the vessels actively producing  $\text{CO}_2$  and  $\text{CH}_4$ ; however, the contents of those reactors were used to quantify the percentages of the bioconversion processes, and were not available as seed cultures.

**Microbial Counts:** Enumerations were determined by modified method described by Craft and Nelson (Appl. Environ. Microbiol. 62:1550-1557(1996)). Enumerations for bacteria, fungi and actinomycetes from the compost mixtures were done by taking 1g of compost and suspending it in 99 mL of 0.1% water agar (1g agar/L). The suspensions were then blended for 45s in a Waring blender. Tenfold dilutions were prepared with sterile distilled water and 0.1 ml of the suspensions spread on 4 replicate plates of the following media:

- for fungi - 0.3X potato dextrose agar with 50  $\mu\text{g}/\text{ml}$  each of streptomycin and ampicillin.
- for bacteria and actinomycetes - 1/10 & 1/50-strength trypticase soy agar, respectively.

All plates were incubated for 72 hrs at 24 C. Colonies were enumerated and populations were



expressed as CFU (colony forming units) per gram of compost.

**Microbial Cultivations:** Molds from the aerobic reactors were transferred by inoculating loops to agar plates containing 1% cellulose or cellobiose. The plates were then incubated at 24°C for up to 2 weeks. The plates were evaluated as positive or negative for supporting growth.

**Analytical Procedure:** Nitrogen was determined by the standard Kjeldahl method and for CHN by combustion analyses. The total volatile solids (TVS), ash and moisture contents were determined by drying the samples overnight in an oven at 100 C, followed by ashing in a muffle oven at 600 C for 45 minutes with the differences in weights recorded gravimetrically. No chemical marker analyses were performed to attempt to determine the amount of microbial biomass that may have accumulated during the bioconversion process and which would contribute to the TVS value. Cautions were taken to not overly shake or mix the dried compost samples during this procedure since the sand readily separated from the dried particles and filtered to the bottom of the container (it stuck firmly to wet particles), otherwise making it impossible to obtain consistent measurements. Even with care in sampling and in selecting samples, representative samples were difficult to obtain because of the diverse nature of the starting materials being fibrous clumps (primary sludge) and a wide range in sizes of bark and grit chips. Even with a ball-mill grinder, representative samples were flawed by the separation of the sand from the particles.

For quantifying the bioconversion rates in the anaerobic reactors, the total volume was removed from the reactor at the end of the run and dried in an oven and gravimetrically measured. Gas analyses were performed with a HP 5890A Gas Chromatograph equipped with a thermoconductivity detector and a 3 m x 0.6 cm SS column packed with 100/120 carbosieve S-II. Identities of the gases were determined by comparing RT values to those of authentic standards of H<sub>2</sub>, CO<sub>2</sub>, CH<sub>4</sub>, CO, N<sub>2</sub>O, NO, O<sub>2</sub>, H<sub>2</sub>O, SO<sub>2</sub>, and NH<sub>3</sub>. The methods used to calculate the volume of gases emitted from the anaerobic reactors was accomplished by Hg pressure gauges according to standard gas laws (Appendix 3) followed by measuring the amount of gas displacement of acidified water in a graded inverted cylinders. The error from CO<sub>2</sub> solubility in water was estimated to be only 1%. The pH of the samples was determined by placing the compost mix into a beaker with a minimal amount of distilled water required to suspend the mix. Samples were recorded directly after a 30 min soaking period as well as by measuring the pH of its filtrate.

## RESULTS

**Aerobic Compost Reactors:** The results from the first three reactors assembled in this study are not included in the tables and figures in this final report and are only described in this paragraph. The reactor mixtures were the same as those described in Table 1 (reactor 1) but with no pH adjustments or additives. The pH values at the start (t<sub>0</sub>) were 8.5 - 9.2. After fourteen days of aeration, no CO<sub>2</sub> or temperature changes were observed and they were discontinued.

The compositions of the mixtures used in the composting bioconversion studies are summarized



in Table 1. The reactor numbers in Table 1 correspond to the reactor numbers as well as the figure numbers used throughout the text. Accordingly, the CO<sub>2</sub> emission profiles that describe the relative biological activities of each of the compost reactors over the course of their operations are listed as Reactors 1-26. The moisture, pH, ash and TVS values for each reactor operated with forced-air in a compost mode are summarized in Table 2. The data in Table 2 and the CO<sub>2</sub> profiles in Reactors 1-26 may not directly correspond since the TVSs include the total solid fraction (biomass, fermentation products, etc.) accumulated in the reactor.

Reactors 1-3 (Table 1) differ from each other particularly in terms of the ratio and contents of the compost substrates, the method of pH adjustment and that Urea as 1% of the mix was added to reactor 2 that lacked an addition of fly-ash. The reason they are linked together is because of the low percentage of CO<sub>2</sub> yields, indicating low biological activity. The use of Urea as a nitrogen source in composting was repeated with similar results as well as used later as a mixture with other nitrogen sources (Reactor 15 to be discussed later). The results from Reactor 1-3 are similar in that the adjustment of pH clearly showed improved biological activity but the increase in CO<sub>2</sub> was still judged as too low for expected productive bioconversion rates. The effect of pH adjustment is most evident in Reactor 1 after pH was adjusted after 10 days of operation. Generally, no pH adjustments were necessary if the concentration of fly-ash was kept below 0.1% of the composition. Nonetheless, fly-ash was not used in further mixes unless adjusted to near neutrality prior to its addition to the compost mix. The addition of Urea (Reactor 2) had relatively little if any effect on the compost mix. This result was attributed to the oligotrophic environment that exist, or to the location of the various nutrients in the mix that are not easily reached by the microorganisms. While the basic data summarizing the reactions of Reactors 1-3 show only 1.7 - 3.6% conversion of the compost to humic type material (Table 2) within 16 days, the volume of the compost mix in the reactor reduced some 30%. This volume reduction was seen in all reactors in which biological activity occurred. This reduction in volume was attributed to biological alterations of components of the mix and to biological activities in general. It was not attributed to settling and compaction over-time due to gravity because reactors without biological activities did not reduce in volume under the same environmental conditions and incubation times. Later on it will be noted that highly biological reactions gave reduction rates approaching 40%.

Reactors 4-7 are compared because each of them was constructed with the same mix and fortified with NH<sub>4</sub>NO<sub>3</sub>. Reactor 5 contained twice the amount of salt as the other three (Table 1). Reactors 4 and 5 contained no additional supplements whereas Reactor 6 received 200 mL of a product advertized as an effective enhancer of composting, sold as Ecosystem Plus by Neozyme Inc., Newport Beach, CA 92663. Although the exact content of Neozyme was not made public, it is clearly a Vinasse solution (yeast fermentation broth) enriched with a surfactant (soap). Assumably, the vinasse is rich in B-vitamins, salts and fermentation by-products that are good supplements to most media for growing microorganisms. Further, the surfactant should increase the potential for interaction between the substrate and microorganisms. Reactor 7 contained "Formula # 615 - special formulation for composting". This product was sold by Deutrel Laboratories, Inc., Palmdale, CA. Warning label states that it contains H<sub>2</sub>SO<sub>4</sub> (the product had a



pH of 1 and appeared to contain little more than trace mineral salts). Reactors 4-7 (and others thereafter) were particularly distinguished by white mycelium of the fungi that completely engulfed the compost mix, giving the appearance of snow. The mold appeared around the 12th day and lasted to the 27-30th day. The appearance of the white mold was consistently observed in all of the additional reactors receiving supplements of  $\text{NH}_4\text{NO}_3$  as the only ammonium salt. Other notable features were the increases in temperature to 35 - 40°C corresponding to  $\text{CO}_2$  emission levels of 15 to 62% of the effluent gases (figures of Reactors 4-7). Temperatures were not expected to get higher because of the relatively large surface area of the reactors that were not insulated. The highest  $\text{CO}_2$  levels were recorded for Reactor 6 which contained Neozyme supplement. This increase was attributed to microbial activity feeding off of the Neozyme products since the TVS and ash contents (Table 2) were not significantly different. In fact, the highest bioconversion of substrates was observed in Reactor 5 which contained 100 gm of  $\text{NH}_4\text{NO}_3$ . The  $\text{CO}_2$  effluent levels were not the highest recorded for this set but they remained at an elevated level for a longer time period (figures of Reactors 4-7). In fact, the  $\text{CO}_2$  profiles demonstrate a possible diauxic curve for those containing  $\text{NH}_4\text{NO}_3$ ; the  $\text{NH}_4^+$  of  $\text{NH}_4\text{NO}_3$  is perhaps used first with the  $\text{NO}_3^-$  utilization occurring afterwards. Thus, a demonstration of two different culture enrichments. The basic conclusions were that neither Ecosystem Plus nor Formula #615 were effective in enhancing composting and that  $\text{NH}_4\text{NO}_3$  had a significant impact on bioconversion rates.

In pursuit of demonstrating the importance of inorganic nitrogen, Reactors 8-12 were assembled to measure the concentration range of  $\text{NH}_4\text{NO}_3$  that affected the bioconversion rate. The amounts of  $\text{NH}_4\text{NO}_3$  added ranged from 30 g to 120 g (Table 1). The results are not as clear as one would expect. This was attributed, in part, to the difficulty in solubilizing more than 60 g of the salt into a minimal amount of  $\text{H}_2\text{O}$  to be sprayed onto the compost mix while keeping the moisture content of the compost mix below 55%. Moisture contents more than 55%, with the added  $\text{H}_2\text{O}$  coming from the metabolic activity, resulted in nonuniformed air diffusion through the matrix. The data in Table 2 indicate that the poorest conversion rate occurred in Reactor 8 when the C:N content was about 123:1. The bioconversion results on the basis of TVS was only 2.6%, similar to that for Reactor 7 with higher  $\text{NH}_4\text{NO}_3$  plus Formula #615. The best conversion rate of 7.9% (based on TVS) occurred in Reactor 11 with a C:N ratio of 43:1. Yet the  $\text{CO}_2$  profile of Reactor 11 (operated for 30 days) showed less  $\text{CO}_2$  output than reactor 9 which had a C:N of 77:1 and a TVS conversion of 6.0%. The amount of gas recorded daily over the life of the reactor does not proportionately reflect the differences in the bioconversion rates (TVS) reported in Table 2. Reactor 12 with the C:N ratio of 30:1 was comparable in  $\text{CO}_2$  profiles and TVS conversion rates to Reactor 9 (C:N of 77:1). In summary, it appears that the highest bioconversion rate expected when only  $\text{NH}_4\text{NO}_3$  is added to the raw compost mix is 8% (Reactor 11, Tables 1 & 2).

To further test the effects of inorganic nitrogen on composting of paper mill waste, Reactors 13-16 were assembled (Table 1), each containing a different nitrogen salt. Reactor 13 contained  $\text{NH}_4\text{Cl}$ , Reactor 14 contained  $\text{NH}_4\text{H}_2\text{PO}_4$  and Reactor 15 contained both of the previous salts plus  $\text{NH}_4\text{NO}_3$  and Urea (Table 1). All of the salts were added to give a calculated final concentration



of C:N of 30:1. The gas flow emission profiles (figures of Reactors 13-15) and final TVSS (Table 2) indicate that the more diverse the nitrogen source the more simultaneous (rather than succession) enrichment of the diverse population of microorganisms which exist in the compost mix. This conclusion was based on the relatively high gas evolution that occurred in Reactor 15. The data is inconclusive as to the preferred nitrogen source with the exception that when Urea was added separately (Reactor 2, Table 2) a positive effect was noted but with relatively small bioconversion rates.

One of the limitations of the compost reactor is the localization of nutrient and microbial events because of the solid nature of the composition and absence of a diffusion medium. Therefore, Reactor 16 (Table 1) was assembled but in a "pond-like" configuration, based on the rationale that nutrients and microorganisms could be mixed and distributed throughout the matrix. An aliquot of the compost mix used to construct Reactor 9 (Table 1) was suspended in 4 Liters of H<sub>2</sub>O and aerated. The percentages of CO<sub>2</sub> in the effluent air (figure of Reactor 16) was almost at background levels and, if present, diluted in the effluent air stream. After 43 days there was no reactivity that was measurable in Reactor 16. It was terminated with the conclusion that the aerated system (lagoon style) was either ineffective in biodegradation of the pulp and fiber, with the system employed for the test, or all of the CO<sub>2</sub> dissolved in the water. A total weight determination was not none, thus no decision can be made about the results.

Another test was conducted with the aerated Reactor 17 containing KNO<sub>3</sub> as the nitrogen supplement. Relatively good gas production occurred in terms of the percentage of CO<sub>2</sub> (Reactor 17, page 25). The CO<sub>2</sub> produced was comparable to other nitrogen supplements with a percentage conversion, in terms of mineralization, of 8.4% (Table 2).

All of the remaining experiments conducted aerobically were those receiving chicken litter as the primary source of nitrogen. Reactors 18, 19 and 20 contained a ratio of compost:chicken litter of 1:1, 2:1 and 1:2, respectively (Table 1). The best conversion rate of 7.8 (Table 2), in terms of TVSS, occurred in the reactor with the 1:2 ratio (Reactor 20, Table 1). In all three Reactors, ammonification occurred as determined by a distinct odor of ammonia in the effluent gas and by the increase of pH from 7.4 to 8.1 - 8.8 (Table 2). No visible molds were detected in these reactors. It was established from the cultivation of the molds on both Sabouraud agar and the reactor mix that the visualization of white mycelia growth does not occur when pH values are above 7.5. In Reactor 20 with 1:2 ratio (Table 1), the biological activity occurred over a longer time-period (CO<sub>2</sub> emission profiles) than Reactors 18 and 19. Most of the activity, however, was attributed to the chicken litter and not biological conversion of the pulp and paper waste. Interestingly, the highest temperature of 45°C was reached in Reactor 18 in the first 4 days of operation, dropping to 25-33°C range for the second week and falling to room temperatures thereafter. The temperatures in Reactors 19 and 20 were more consistent with values of 29-38°C for the first 21 days and then cooling-off to room temperature of about 22°C. These were considered reasonably high temperatures since the reactors were uninsulated and had relatively high surface areas. The percentages of CO<sub>2</sub> were clearly the highest in these reactors than any of those previously studied without chicken litter (see figures of reactors 18-20).



The compost to chicken litter mix was again restructured at ratios of 2:1 (reactor 23, a repeat of reactor 19), 1:1 (Reactor 24, a repeat of Reactor 18), 1:2 (Reactor 25, a repeat of Reactor 20) and 3:1 (Reactor 22). Reasonably good correlations were observed for the ratios of 2:1 (Reactors 19 and 23, Table 2) and 1:2 (Reactors 20 and 25, Table 2) with differences in TVS values of 5.6/7.6 and 7.8/9.0, respectively. A larger difference was observed with the duplicates of the ratio of 1:1 in Reactors 18 and 24 with differences in TVS values of 4.7 and 10.7, respectively. The duplicates of each reactor mix were set up some three months apart. One must assume, however, when relatively small samples are used, the variation in the sample may be significant. This problem should disappear on scale-up. Expectedly, the highest CO<sub>2</sub> output occurred with the reactors containing the highest proportion of chicken litter in Reactors 18 & 24 (1:1) and Reactors 20 & 25 (1:2). Nonetheless, the CO<sub>2</sub> profiles are different (see corresponding figures).

Reactor 22 containing a ratio of 3:1 of compost mix and chicken litter produced a broad and continuous output of CO<sub>2</sub> resulting in a mineralization of 12% of the total compost mixture (Table 2). The ratio was repeated twice with very similar results. While the level of CO<sub>2</sub> was less than those with higher concentrations of chicken litter (TVS of 5.6-9.0, Reactors 19, 20, 23, 25 and Table 2) the degree of mineralization was greater in Reactor 22 (Table 2) with its 3:1 ratio.

The highest percentage of mineralization (21.5%) was recorded in Reactor 26 which contained only primary sludge and chicken litter in a ratio of 3:1. This was expected since the bark and grit were excluded with the sawdust of the chicken-litter being sufficient for bulking the compost.

Reactor 21 was established with compost mix to chicken litter in a ratio of 4:1 (Table 1). As expected, the biological activity, measured by CO<sub>2</sub> emissions, was initially high but relatively brief, completing the cycle in 10 days. About 16% of the compost mix was degraded but over a 56 day period as determined by combustion of aliquot samples and quantitating the ash and TVS contents by gravimetric methods (Table 2). The volume of compost was also reduced by 35%. At the end of the 56th day, the compost mix was sprayed with 200 ml of 75 g of NH<sub>4</sub>NO<sub>3</sub>. The matrix was rebulked with grit and bark (1:1) and added to the compost mix to replace the approximately 35% volume lost. The additional supplement had little effect on the bioconversion rate, increasing the CO<sub>2</sub> emission only slightly (Reactor 21A). At 78 days the CO<sub>2</sub> emission had fallen to near zero. At the 78th day the matrix was removed and sprayed with a 100ml of 60g of KH<sub>2</sub>PO<sub>4</sub> (Table 1). A modest increase in CO<sub>2</sub> resulted and appeared to stabilize for the next 20 days (Reactor 21A). At the 100th day, the forced air flow to the reactor was stopped, the reactor flushed with N<sub>2</sub> gas and the reactor sealed. The reactor was then monitored daily for CO<sub>2</sub> and CH<sub>4</sub> production under anaerobic conditions. The CO<sub>2</sub> level remained stable in the closed reactor for an additional 56 days (156th day) with no detectable CH<sub>4</sub> production and only a brief and small gas pressure buildup on three brief occasions. These results support a proposal that addition to the reactants with salt supplements after extended incubation periods are ineffective. This conclusion is further supported by the results presented in the anaerobic digestion study described in the following section.



In none of the reactors was the moisture content of the mixes particularly critical. In all cases where biological activity occurred additional "wetness" occurred. In general, It was found that when the moisture content approached the lower limit of 35% microbial activity was slow and poor reaction rates were obtained. Thus, in all experiments, care was taken to have the starting moisture contents at no less than 40%. When the moisture content at the start of the composting was at 65% and higher, the additional water from the composting caused the mix to be too wet and the porosity diminished. In such cases, air diffusion was limited and in a few trial runs, CH<sub>4</sub> gas was observed in relatively trace quantities in the outflow gas stream. Thus, moisture contents of 45 to 55% were considered as the ideal operating range.

The temperature for optimal composting rates must vary. Most microorganisms with diverse metabolic properties are mesophiles (10- 40°C). These are the major consumers of complex organics. The higher temperatures are also needed for the heat shocking of fungal and bacterial spores and for enhancing activities of thermotolerant organisms that are known to exist. Thus, the microbial activity is one of succession and interspecies activities. No significant thermophilic microbial activity per se was expected because of the nature of the *in situ* cultures adapted to psychrophilic and predominantly mesophilic environments. For optimal activities, the temperature range should naturally be allowed to vary from 20°C to 60°C. Temperatures above 60°C diminish biological conversion rates and if allowed to persist, lengthen the recovery of the microbial activities. The temperatures can be controlled by aeration rates and moisture content. The temperature cycle up to 60°C should last for only 2-3 days if the compost is properly bulked and air diffusion is not limited.

It is felt that much of the bioconversion data (TVS/ash) presented in Table 2 represent values below those that actually existed. The basis for this lies in the difficulty to obtain representative samples on a small scale for measuring ash and TVS. The diverse sizes of wood and bark chips and the fluffy nature of the primary sludge and the high content of sand created technical difficulties. The sand adhered to the particles when wet but easily separated from the particles when dry. Also, ball-milling the particle enhanced the separation of sand from the moderately dense primary sludge and the less dense bark chips. Unfortunately, not all of the sand or consistent amounts of it could be removed by this method. In the first lot of primary sludge obtained from the plant in Olgethorpe, GA, the ash content was between 20 and 30%. For the second lot of primary sludge, the ash content was between 40 and 68%. Significant quantities of sand separated from the samples during the collection, drying and ashing process. When added back to the samples, the conversion rates are significantly higher. In due process, however, such corrections were not made in Table 2.

#### Anaerobic Compost Reactors.

The mix of pulp and paper waste used in the aerobic composting studies described above was also used in most of the anaerobic experiments. Seven different types of nitrogen compounds and mixes thereof were tested as supplements. All seven sources of nitrogen stimulated microbial activity, but with mixed results. The experimental description of the supplement added to each



reactor, the concentrations, and sequences of additions are summarized in Table 1 and are designated as A1-A21. The performances of these reactors are summarized in Tables 3 and 4 (pp31 & 32). The amount of gas pressure produced daily (pp35-46), the total moles of gas produced daily (pp47-65) and the mole percentages of each gas (pp66-80) are given in figure form. All reactors started out under microaerobic conditions but quickly went to anaerobic condition within the first week of operation as detected by oxygen sensitive strips and the loss of  $O_2$  detection by gas chromatography. Nonetheless, the  $N_2/O_2$  detections are not differentiated in the figures. The addition of  $NH_4NO_3$  and  $NH_4Cl$ , resulted in predominantly  $CH_4$  and  $CO_2$  production whereas the addition of  $KNO_3$ ,  $NH_4H_2PO_4$ , Urea and chicken litter produced principally  $CO_2$ . The amount of gas pressure produced are summarized in Table 3 with the daily productions recorded in the corresponding figures. In all reactors containing an addition of  $NO_3^-$ , nitrous oxide ( $N_2O$ ) was produced. The most abundant  $N_2O$  was produced in Reactor A11 with only  $KNO_3$  added in the beginning of the incubation period. The results are expressed in the figures designated A11 (p38,39,56, & 71). For  $NH_4NO_3$  containing mixes in Reactors A1- A5 (Table 1 and corresponding figures),  $N_2O$  appeared in relatively trace quantities during the first two weeks of the reaction and disappeared from detection prior to the appearance of  $CH_4$ . The detection of  $N_2$  occurred throughout the reaction periods, in all reactors, with considerable variation in relative quantity. Some of the  $N_2$  variation was attributed to contamination from air during sampling, however, the variation was somewhat rhythmical suggesting a combined nitrification-denitrification system in operation like that previously described by Pel et al (Appl. Microbiol. 63:474-481, 1997) under limited oxygen conditions.

The highest total gas pressures were recorded with those reactors containing  $NH_4NO_3$  and  $NH_4Cl$ . These reactors also produced gas pressures over the longest time period and with the highest metabolism of substrates (Table 3) and with the largest percentage of  $CH_4$  (Table 4). In Reactors 1-7, 33 to 59% of the starting materials were mineralized during the incubation period (Table 3) as determined by dry weight differences. Interestingly, the odor of the final product of A1 and A2 smelled like fresh barnyard waste whereas the odor of A3 and A4 were more earthy. Perhaps the differences are due to the addition of  $NH_4H_2PO_4$  to Reactor A3 and  $NH_4Cl$  to Reactor A4 on day 119 of the incubation periods (Table 1, p17). The addition of these salts to Reactor A3 and A4, late in the incubation period, however, had no significant impact on the bioconversion of the existing reactants.

The effects from the addition of  $NH_4Cl$  (Reactor A5, Table 1) and  $NH_4H_2PO_4$  (Reactor A6, Table 1) and combined  $NH_4Cl$  and  $NH_4H_2PO_4$  (Reactor A7, Table 1), at a calculated C:N concentration of about 30:1, are summarized in Tables 3 and 4 and the corresponding figures. Gas production in Reactor A6 responded most quickly but dropped off after only 20 days and didn't start generating significant amounts of gas until the 65th day (Fig. A6, p36). Reactor A5 with  $NH_4Cl$  started to produce after 35 days and continued for another 35 days but at a relatively low rate. The addition of  $NH_4H_2PO_4$  to Reactor A5 (Fig. A5, p36) on the 80th day had no effect on the bioconversion rate. The amount of  $CH_4$  produced was less than 36% of the total gas produced. The dry weight differences between the starting material and the end product, however, showed that a respectable 39.3% of the material was lost during the incubation period. Reactor A6,



however, received  $\text{NH}_4\text{Cl}$  on the 80th day resulting in a significant simulation of gas production (Fig. A6, p36). Its percent  $\text{CH}_4$  was similar to Reactor A5 and its difference in total weight of the reactants after incubation showed that 44% was lost, only slightly greater than Reactor A5. Reactor A7 with both salts had better gas production (Fig. A7, p36), methane production of 44% and greater utilization of reactants than the reactors containing the individual salts (Tables 3 and 4). The required incubation times for the reactors was the shortest for Reactor A7 with only 90 days as compared to 120 days for Reactors A3, A4, A5 and A6 and 150 days for Reactors A1 and A2. The results suggest that mixed salts added initially provide better enrichment conditions. However, the results from Reactor A8 which contained all of the salts plus urea in a C:N ratio of 30:1 gave poorest results. High gas production started to occur on the 7th day in Reactor A8, peaking at the 17th day but soon dropping to no pressure and then neg. pressure on the 22nd day. The pH of the reactants was 6.06 on day 23. Thereafter, the gas pressure oscillated in five day cycles (Fig A8, p69) with gas compositions of principally  $\text{CO}_2$  with only detectable amounts of  $\text{H}_2$ ,  $\text{N}_2\text{O}$ , and  $\text{CH}_4$ . The bioconversion reduction of the reactants was only 15%. The final product smelled strongly of butyrates indicating that one of the predominant microbial enrichments was possibly the fermentative Clostridia. The same mixture with the nitrogen sources was also placed in a 20 L reactor in a solid form and sealed. This reactor labeled A9 reflected what one may find in a landfill if diverse nitrogen salts were added. Essentially, nothing had happened after 85 days of incubation.  $\text{CO}_2$  was the major head-space gas but with no gas pressure or detectable changes occurring. These results are comparable to a similar set up of reactants taken from aerobic compost Reactor 21 (Table 1), spiked with nitrogen supplements and sealed after flushing with nitrogen gas to remove air from the chamber. This Reactor produced no indication of further metabolic activity.

Reactor A10 was supplemented with only urea. No gas pressure changes were detected until the 7th day when a negative pressure was recorded (fig. A10, p37). The reactor was unsealed and pH recorded at 8.9. With a strong smell of ammonia coming from the flask, it was assumed that ammonification occurred with the  $\text{NH}_4^+$  dissolving in the aqueous mixture and creating the negative pressure. The pH was adjusted to 7.64 with  $\text{H}_2\text{SO}_4$  and the reactor resealed. Thereafter, a positive gas pressure was recorded for the next 25 days. No  $\text{CH}_4$  was detected (Table 3), however, a respectable reduction (28%) of reactants was recorded (Table 3). The odor of the final product was once again that of specific fermentative bacteria with a strong smell of butyrates.

Reactors A-11 supplemented with  $\text{KNO}_3$  and A12 reactants mixed with Chicken litter are described in the corresponding figures on pp39,40,56,57,71 and 72. The only gases detected in A12 were  $\text{N}_2$  and  $\text{CO}_2$  with  $\text{CO}_2$  comprising 80% of the total gas emitted for the first 50 days and then a moderate amount of  $\text{CH}_4$  produced thereafter. Reactor A11 differs in that a significant fraction of the emitted gases was  $\text{N}_2\text{O}$  for an extended period. After the  $\text{NH}_4\text{Cl}$  spike, the  $\text{CH}_4$  has gradually increased with the inverse production of  $\text{CO}_2$ . The final results are reported in Table 3.

Reactors A13 and A14 were established with the standard mixture of pulp, fiber, bark, grit and



ash with no exogenous nitrogen added. These represented controls. The two preparations differ, however, in that the primary sludge used in bioreactor A13 was taken from a barrel containing high moisture content (loaded in the barrel 12 months previously) and demonstrated fermentation processes. The primary sludge used in bioreactor A14 was taken from a more recently loaded Barrel that had little moisture and no evidence of biological activities (see Table 1B). Bioreactor A14 showed no biological activity for the first 35 days and it was subsequently terminated. Bioreactor A13, however, showed biological activity within the first 2 days of incubation. The amount of biological activity in A13 was relatively good in comparison to other Reactors supplemented with exogenous nitrogen. Most of the gases produced was  $\text{CO}_2$  with low amounts of  $\text{CH}_4$  occurring at 45 days. At the end of 45 days the majority of the contents of Reactor A13 were used as the starter cultures for reactors (A19-A21). The remnants of A13 as well as several others established just like it are being held for seed cultures for future reactors.

Reactors A16 and A18 were structured like A13 with primary sludge taken from the barrel with biological activity but different from A13 in that these two reactors received  $\text{NH}_4\text{Cl}$  in A18 and  $\text{NH}_4\text{H}_2\text{PO}_4$  in A16. For comparison to the effects from starting with an active biological culture vs. one not biologically active, Reactor 17 was established with dried primary (nonactive) sludge but with  $\text{NH}_4\text{Cl}$  supplement like A18. A17 was terminated after 30 days as it was just starting to ferment (results not included in this report). This experiment was performed only to establish the differences in lag time of seeded and unseeded cultures. A18 and A16 started to produce gas within hours of the start of its incubation period. A18 differed from A16 in that  $\text{CH}_4$  was produced almost immediately whereas no  $\text{CH}_4$  occurred in A16 until much later and even then at relatively low quantities. A16 continues to produce high gas levels at day 45 but still with relatively low  $\text{CH}_4$  yields (Fig. 16A, pp42, 61 and 76). A18's level of  $\text{CH}_4$  reached 50% of the total gas within 20 days of incubation, however, the gas pressures began to drop shortly thereafter. A small supplement of  $\text{NH}_4\text{Cl}$  stimulated methanogenesis. The periodicity of the supplements are given in Table 1, with results described in Figures on pp 43, 62 & 76. The summary of the results of experiment A18 is given in Tables 2 & 3 (pp33 & 34).

Three addition reactors have been recently established with nonbiologically active mix (A19, A20, A21) with different  $\text{NH}_4\text{Cl}$  concentration and seeded with active cultures from bioreactor A13. The nitrogen supplement was determined to be at a C:N ratio 45:1, 60:1 and 75:1 for Reactors A19, A20 and A21, respectively. The biological activity was almost immediate with highest gas pressures produced first in Reactor A21 (Fig. A21, p44), the reactor with the least amount of  $\text{NH}_4\text{Cl}$ . The original intent was to supplement the reactors with additional salts through the course of their incubation periods. The results, however, showed excellent methanogenic activity for A21 with activity equal to or better than Reactors A20 and A19 with higher nitrogen supplements (Fig. A20 & A21, pp63-64 and pp78-80). The summary of the results are given in Tables 3 & 4. At 80 days, A21 has just about stabilized while A20 and A19 have moderate gas production, attributed to the additional supply of nitrogen. A19, however, is currently producing principally  $\text{CO}_2$ . The dry weight conversions values of these three reactors has not been completed. Further experiments will have to be established to determine the concentration of nitrogen needed to stimulate methanogenesis and the subsequent feeding strategy.



Nitrogen, carbon and hydrogen analyses were performed for 7 of the anaerobic reactors. The results are summarized in Table 5. The results are mixed. The results are suspect since only a few milligrams were analyzed from each preparation and questions remain about the representation of the samples analyzed. The data support the data presented in Table 3 that most of the nitrogen stayed in the reactor. The CHN analyses (Table 5) support the end-product results reported in Table 3 in terms of the high ash content and the percent converted in the reaction.

The market analyses of the value of  $\text{CH}_4$  produced was calculated for each of the reactors operated anaerobically (Table 4). The BTU values of the higher producers of  $\text{CH}_4$  were also calculated on a dry weight basis and organic content basis. The highest percentages of methane gas (>55%) was observed in Reactors A1-A4 and A21 (Table 3) which contained  $\text{NH}_4\text{NO}_3$  (A1-A4) or  $\text{NH}_4\text{Cl}$  (A21) at the start ( $t_0$ ) of the incubation period (Table 4), but in total liters of  $\text{CH}_4$ , the best producers were A4 and A18-21. Additions of  $\text{NH}_4\text{NO}_3$  later in the incubation periods were ineffective. Reactors A1-A5 were essentially replicates with the exception that increasingly more water/unit weight of reactants was added with A1 receiving the least proportionate amount of water (Table 1). The data clearly shows a relationship between the greatest amount of gas being produced and the more liquid properties of A3 and A4. Better total reduction in reactants, however, occurred in Reactors A1 and A2. Obviously there is a difference in biomass and fermentation product accumulation relative to the dynamics of the system some of which are volatilized during the ashing process. An important piece of information learned at this point in the study is the stimulation of the methanogens at higher C:N ratios, concentrations significantly removed from the reported ideal ratio of 30:1.

As for the nitrogen sources, the general conclusions are:

- Chicken Litter stimulates organisms that produce  $\text{CO}_2$ ,
- Urea stimulates organisms that produce  $\text{CO}_2$ ,
- $\text{KNO}_3$  stimulates organisms that produce  $\text{CO}_2$  and  $\text{N}_2\text{O}$  (in decreasing order)
- $\text{NH}_4\text{Cl}$  stimulates organisms that produce  $\text{CH}_4$  and  $\text{CO}_2$  (in decreasing order)
- $\text{NH}_4\text{H}_2\text{PO}_4$  stimulates organisms that produce  $\text{CO}_2$  and low levels of  $\text{CH}_4$
- $\text{NH}_4\text{NO}_3$  stimulates organisms that produce  $\text{CH}_4$ ,  $\text{CO}_2$  and  $\text{N}_2\text{O}$  (in decreasing order)
- $\text{NH}_4\text{SO}_4$  stimulates organisms that reduce sulfur

All of the nitrogen sources used to supplement the reactions resulted in significant mineralization and reduction in volume of the pulp and paper mill waste. The lowest bioconversions occurred with the additions of Urea, chicken litter and  $\text{KNO}_3$ .

#### Measurement of Microbial Populations.

A general enumeration of the microbial populations was conducted with the reactants of Reactors 8, 10, 11, and 12 containing different quantities of  $\text{NH}_4\text{NO}_3$ . The cultivation media used was chosen on the basis of testing the relative fastidiousness of bacteria on 1/10 and 1/50 Trypticase Soy Agar (Table 6) and for fungi on potato dextrose agar containing the antibiotic streptomycin



positive bacteria (Table 6). Samples were withdrawn from the reactors for cultivations at the start of the incubation period ( $t_0$ ), at day 12 ( $t_{12}$ ) of the incubation and when the incubation was stopped at 31 days ( $t_{31}$ ). At  $t_0$  the number of colony forming units (CFU) of microorganisms was relatively low in comparison to the number of colonies at  $t_{12}$  and  $t_{31}$ . As expected the highest CFU was at  $t_{12}$  since numerous protozoa and insects appeared in the reactors after  $t_{12}$  samples were taken and which normally feed on the microbial populations.

Media was also prepared with agar and cellulose or biocellulose as the only carbon source to test the ability of the population to produce cellulase. Two different molds were isolated that had the ability to grow on cellulose while numerous CFUs were obtained from the biocellulose medium. Mycelia cultures were also observed in the anaerobic reactors growing on "chips" of wood and bark attached to the sides of the cultivation flasks. No identities of the organisms were attempted at this time.



REACTOR	INGREDIENTS	ADDITIVES
1	Primary Sludge 3 Fly Ash 1 old Grit 2 Old Bark 1	pH adjusted to 7.2-7.5 on day 10 with 200 ml H <sub>2</sub> SO <sub>4</sub> and 100 ml HNO <sub>3</sub>
2	Primary Sludge 3 Old Grit 1 Old Bark 1	1% Urea added on day 6
3	Primary Sludge . 3 Old Grit 2 Fly Ash 0.01 Old Bark 1	Fly Ash was adjusted to pH 7.2 with 1 M H <sub>2</sub> SO <sub>4</sub>
4	Same as 3	50 g of NH <sub>4</sub> NO <sub>3</sub> to 5.46 Kg of mix, added as spray in 100 ml H <sub>2</sub> O
5	Same as 3	100 g of NH <sub>4</sub> NO <sub>3</sub> to 5.5 kg of mix, added as spray in 200 ml H <sub>2</sub> O
6	Same as 3	50 g of NH <sub>4</sub> NO <sub>3</sub> plus the addition of 200 ml of undiluted Ecosystem Plus <sup>R</sup>
7	Same as 3	50 g of NH <sub>4</sub> NO <sub>3</sub> plus 20 ml of Formula #615 Composting <sup>R</sup> in 100 ml of water sprayed on 5.56 Kg of mix.
8	Same as 3	30 g of NH <sub>4</sub> NO <sub>3</sub> calculated to give a C:N ratio of 123:1
9	Same as 3	50 g of NH <sub>4</sub> NO <sub>3</sub> in 100 ml H <sub>2</sub> O to 11.5 kg to give a C:N ratio of 77:1
10	Same as 3	60 g of NH <sub>4</sub> NO <sub>3</sub> to 11.5 Kg to give a C:N ratio of 64:1



11	Same as 3	90 g of $\text{NH}_4\text{NO}_3$ to 11.5 Kg give a C:N ratio of 43:1
12	Same as 3	120 g of $\text{NH}_4\text{NO}_3$ to 11.5 Kg to give a C:N Ratio of 32:1
13	Same as 3	72.73 g of $\text{NH}_4\text{Cl}$ in 200 ml $\text{H}_2\text{O}$ to 5.5 kg of mix
14	Same as 3	100 g of $\text{NH}_4\text{H}_2\text{PO}_4$ in 400 ml $\text{H}_2\text{O}$ on 5.5 kg of mix
15	Same as 3	170 gm $\text{NH}_4\text{Cl}$ + 30 gm $\text{NH}_3\text{H}_2\text{PO}_4$ + 100 gm $\text{NH}_4\text{NO}_3$ + 50 gm Urea in 600ml $\text{H}_2\text{O}$ sprayed on 18.2 kg of mix
16	Same as 3	36 gm $\text{NH}_4\text{NO}_3$ was added to 4L $\text{H}_2\text{O}$ and used to suspend 1.9 kg of mix at 40% moisture
17	Same as 3	400 gm $\text{KNO}_3$ to 1.9 kg of mix
18	Same as 3	Mixed with chicken litter at a ratio of 1:1
19	Same as 3	mixed with chicken litter at a ratio of 2:1
20	Same as 3	mix with chicken litter at a ratio of 1:2
21	Same as 3	mix to Chicken litter ratio 4:1. After 56 days, mix was bulked with old bark and grit (1:1) at 3:1 ratio and sprayed with 75 gm $\text{NH}_4\text{NO}_3$ . After 78 days, sprayed with 60 gm $\text{KH}_2\text{PO}_4$ . (see 21A)
21A	Composted ingredients of mixture 21	After 100 days, the reactor was flushed with $\text{N}_2$ and sealed for anaerobic growth.



**Description of Sample Preparation****Table 1A**

22	Same as 3	mix to chicken litter ratio 3:1
23	Same as 3	mix to chicken litter ratio 2:1
24	Same as 3	mix to chicken litter ratio 1:1
25	Same as 3	mix to chicken litter ratio 1:2
26	Primary sludge	mix to chicken litter ratio of 3:1



Table 1B

SAMPLE	INGREDIENTS	ADDITIVES
A1	Same as 3	1,874.95 g of mix with 30.28 g $\text{NH}_4\text{NO}_3$ in 2 L $\text{H}_2\text{O}$ .
A2	Same as 3	953.11 g mix plus 23.17 g $\text{NH}_4\text{NO}_3$ in 2 L $\text{H}_2\text{O}$ . 20 ml of Spizizen mineral medium was added at day 141.
A3	Same as 3	630.06 g of mix with 17.36 g $\text{NH}_4\text{NO}_3$ in 2 L $\text{H}_2\text{O}$ . 5 gm $\text{NH}_4\text{H}_2\text{PO}_4$ in 20 ml $\text{H}_2\text{O}$ was added at day 119.
A4	Same as 3	625.49 g of mix with 16.43 g $\text{NH}_4\text{NO}_3$ in 2 L $\text{H}_2\text{O}$ . 5 gm $\text{NH}_4\text{Cl}$ in 20 ml $\text{H}_2\text{O}$ was added at day 119.
A5	Aliquot of sample 13	612 g aliquot of #13 with an additional 0.5% $\text{NH}_4\text{Cl}$ in 1 L $\text{H}_2\text{O}$ . 5 gm $\text{NH}_4\text{H}_2\text{PO}_4$ in 20 ml $\text{H}_2\text{O}$ was added on day 81.
A6	Aliquot of sample 14	415.23 g aliquot of #14 with an additional 0.5% $\text{NH}_4\text{H}_2\text{PO}_4$ in 1 L $\text{H}_2\text{O}$ . 5 gm $\text{NH}_4\text{Cl}$ in 20 ml $\text{H}_2\text{O}$ added on day 81.
A7	mixed aliquots of samples 13 & 14	204.5 g #14 + 211.75 g #13 with an additional 0.25 % each of $\text{NH}_4\text{Cl}$ & $\text{NH}_4\text{H}_2\text{PO}_4$ in 2 L $\text{H}_2\text{O}$ . 0.5 gm $\text{NH}_4\text{NO}_3$ in 20 ml $\text{H}_2\text{O}$ added on day 81.
A8	Aliquot of sample 15	663 gm suspended in 1.5 L $\text{H}_2\text{O}$
A9	Aliquot of sample 15	6.4 kg loaded and sealed with no $\text{H}_2\text{O}$ addition



# Description of Sample Preparation

Table 1A

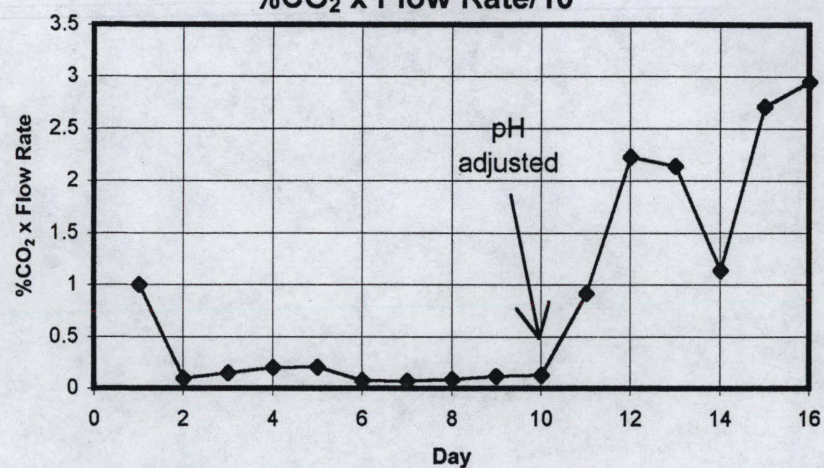
A10	Same as 3	647 gm + 10 gm urea in 1.5L H <sub>2</sub> O
A11	Same as 3	636.1 gm + 25.2 gm KNO <sub>3</sub> in 1.8L H <sub>2</sub> O + 5 gm NH <sub>4</sub> Cl on day 46 and 2 gm on day 85
A12	Same as 3	488 gm + 184 gm Chicken Litter in 1.5 L H <sub>2</sub> O
A13	Same as 3	698.4 gm (no nitrogen added), wet and biologically active
A14	Same as 3	654 gm of dried mix with no nitrogen added
A15	Same as 3	558.8 gm + 14.6 gm (NH <sub>4</sub> ) <sub>2</sub> SO <sub>4</sub> in 1.5 L H <sub>2</sub> O
A16	Primary sludge only	1940.1 gm of wet active sludge + 50 gm NH <sub>4</sub> H <sub>2</sub> PO <sub>4</sub> in 1.3 L H <sub>2</sub> O
A17	Primary sludge only	1104.1 gm dried sludge + 20 gm NH <sub>4</sub> Cl in 1.7 L H <sub>2</sub> O
A18	Primary sludge only	2179.5 gm active sludge + 25 gm NH <sub>4</sub> Cl + 1 L H <sub>2</sub> O + 2 gm NH <sub>4</sub> Cl on day 35
A19	Same as 3	938.4 gm + 500 mL from reactor 13 + 12 gm NH <sub>4</sub> Cl + 1.8 L H <sub>2</sub> O
A20	Same as 3	927.4 gm + 500 mL from reactor 13 + 9 gm NH <sub>4</sub> Cl + 1.8 L H <sub>2</sub> O
A21	Same as 3	945.7gm + 500 mL from reactor 13 + 7.2 gm NH <sub>4</sub> Cl + 1.8 L H <sub>2</sub> O



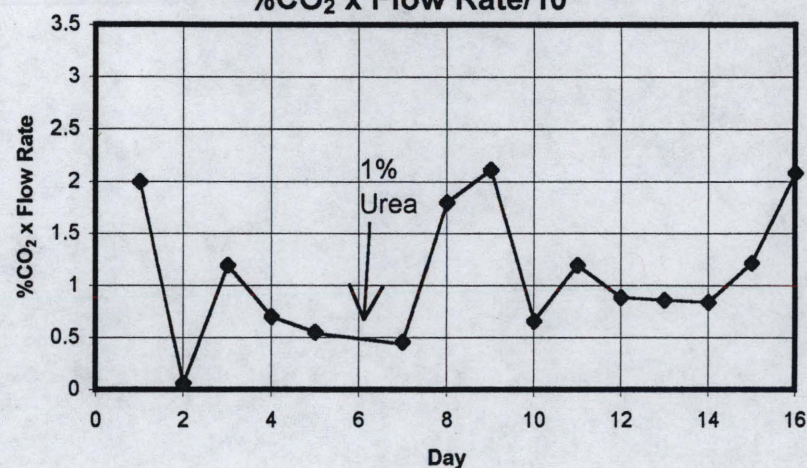
Sample	Moisture %		pH		Ash		TVS		Converted
	$t_0$	$t_f$	$t_0$	$t_f$	$t_0$	$t_f$	$t_0$	$t_f$	%
1	44.0	47.0	8.9	7.3	30.1	31.8	69.9	68.2	1.7
2	47.0	50.1	7.4	8.6	28.8	32.4	71.2	67.6	3.6
3	45.0	46.0	7.2	7.4	28.1	30.1	71.9	69.8	3.8
4	43.9	63.2	7.7	7.5	23.1	26.6	76.8	73.4	3.4
5	41.5	64.2	7.4	7.4	20.0	24.4	80.0	75.6	4.4
6	58.1	56.8	7.9	7.5	21.4	25.2	78.6	74.8	3.8
7	58.5	53.9	7.6	7.0	21.1	23.7	78.9	76.1	2.8
8	63.3	65.5	7.0	7.6	27.3	29.9	72.7	70.1	2.6
9	40.3	46.3	7.5	7.6	22.7	28.7	77.3	71.3	6.0
10	63.3	65.9	7.0	7.4	27.3	31.6	72.7	68.4	4.3
11	63.3	66.1	7.0	7.5	27.3	35.2	72.7	64.8	7.9
12	63.3	64.0	7.0	7.3	27.3	32.4	72.7	67.6	5.1
13	59.6	61.0	7.4	6.8	37.8	43.6	62.2	56.4	5.8
14	46.9	61.8	6.8	7.2	39.2	47.4	60.8	52.6	8.2
15	49.1	67.5	7.0	7.2	45.6	55.8	54.4	44.2	10.2
16	NA	NA	7.5	NA	45.6	NA	54.4	NA	NA
17	52.0	53.8	7.4	7.1	45.6	54.0	54.4	46.0	8.4
18	35.0	38.9	7.5	8.3	26.6	31.3	73.4	68.7	4.7
19	53.1	37.7	7.8	8.1	26.6	29.2	73.4	70.8	5.6
20	41.9	49.0	7.6	8.8	19.4	27.2	80.6	72.8	7.8
21	45.6	64.4	7.8	7.3	24.9	40.9	75.1	59.1	16.0
22	52.0	51.6	7.6	9.1	21.7	33.7	78.3	66.3	12.0
23	53.8	65.0	7.9	8.9	21.6	29.2	78.4	70.8	7.6
24	50.3	60.4	8.0	8.9	21.0	31.7	79.0	68.3	10.7
25	42.5	67.8	8.2	8.9	19.4	28.4	80.6	71.6	9.0
26	60.5	72.9	7.8	6.7	19.4	40.9	80.6	59.1	21.5
*Sample numbers correspond to Reactor numbers identified in Table 1 $t_0$ and $t_f$ represent time zero and final time, respectively									



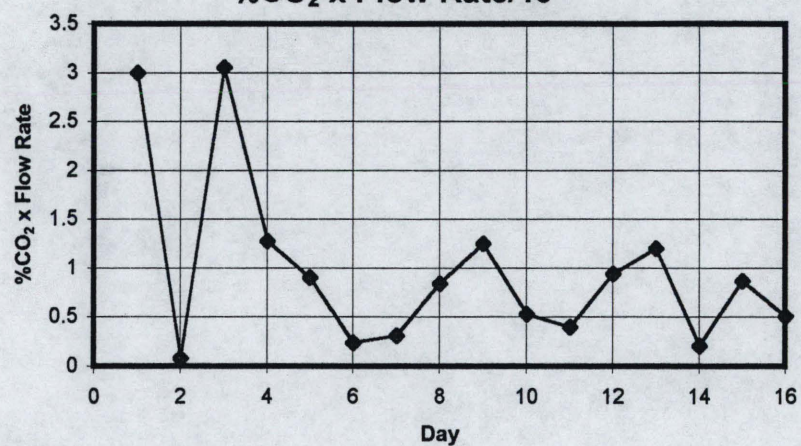
Reactor 1  
%CO<sub>2</sub> x Flow Rate/10



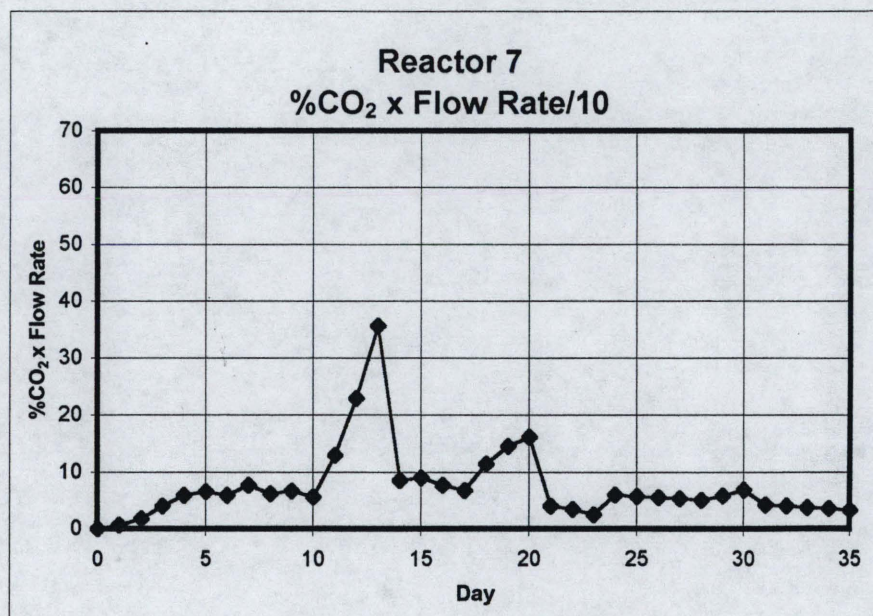
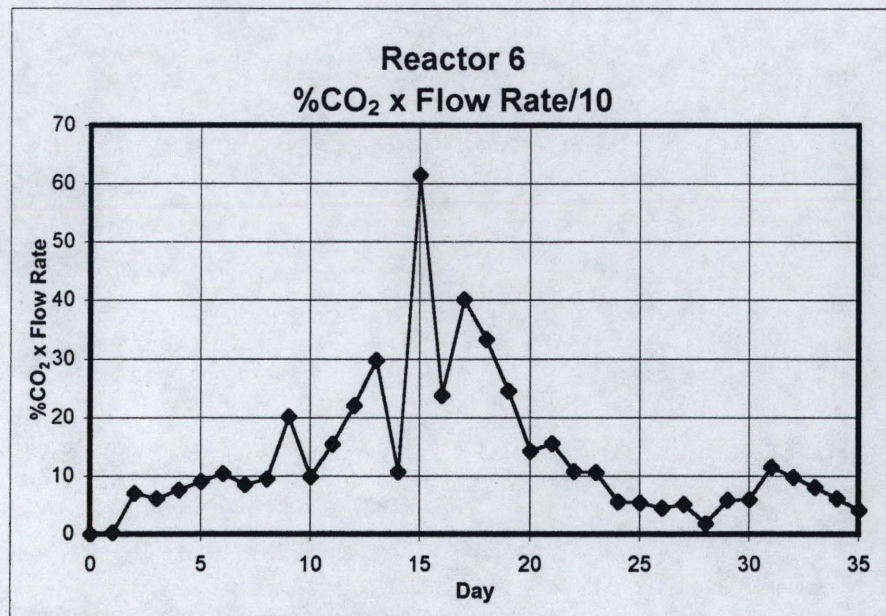
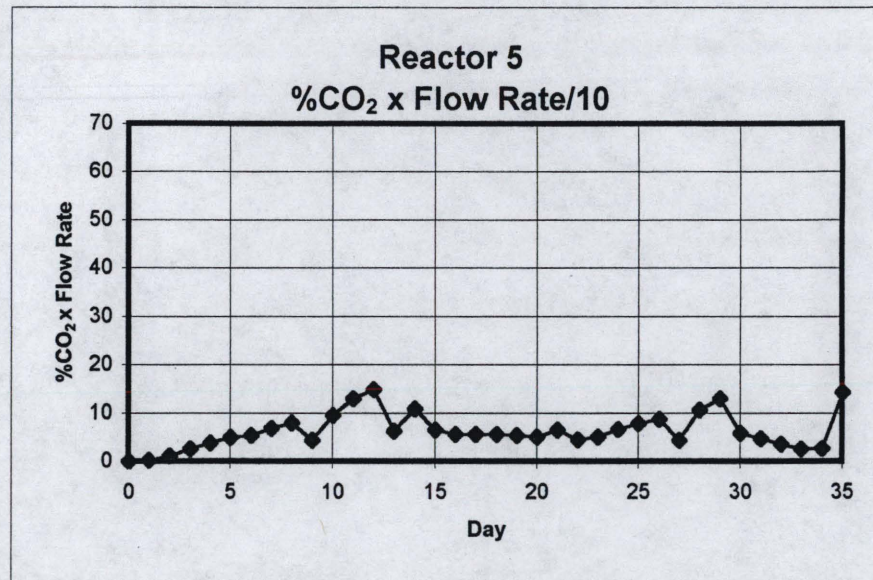
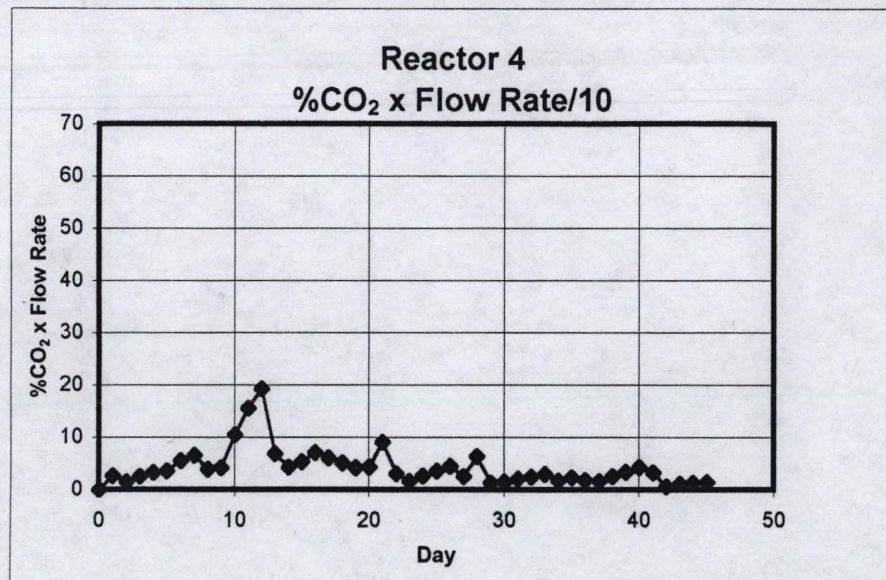
Reactor 2  
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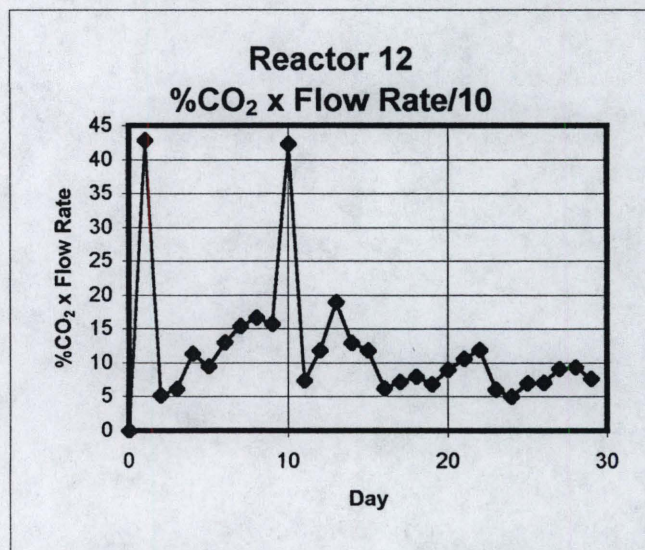
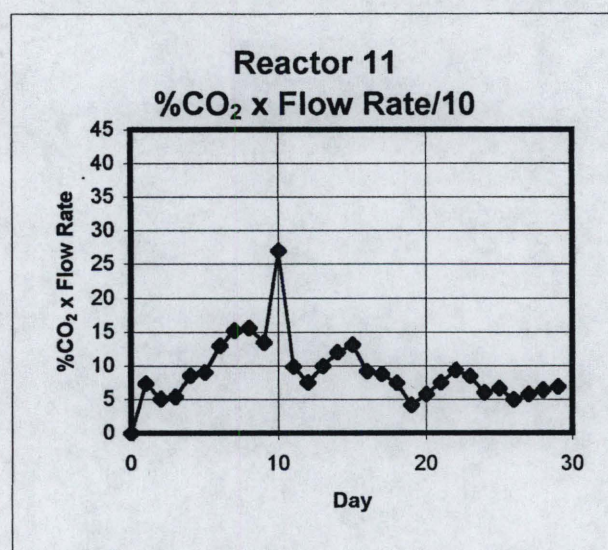
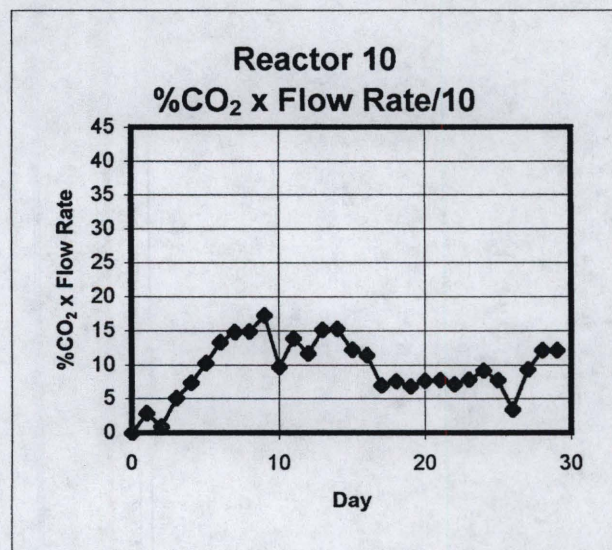
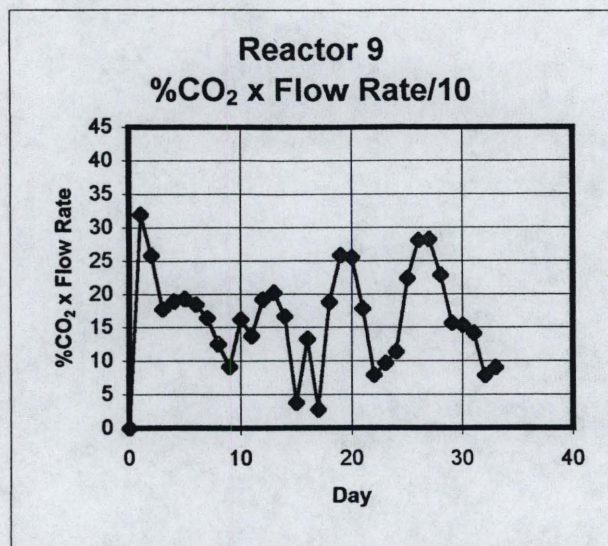
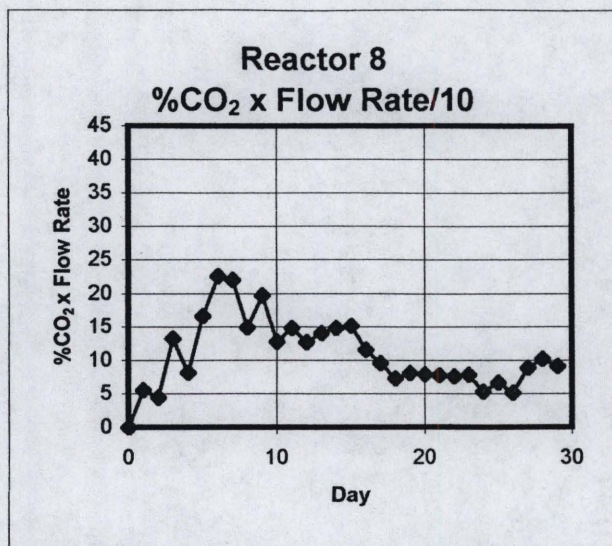
Reactor 3  
%CO<sub>2</sub> x Flow Rate/10



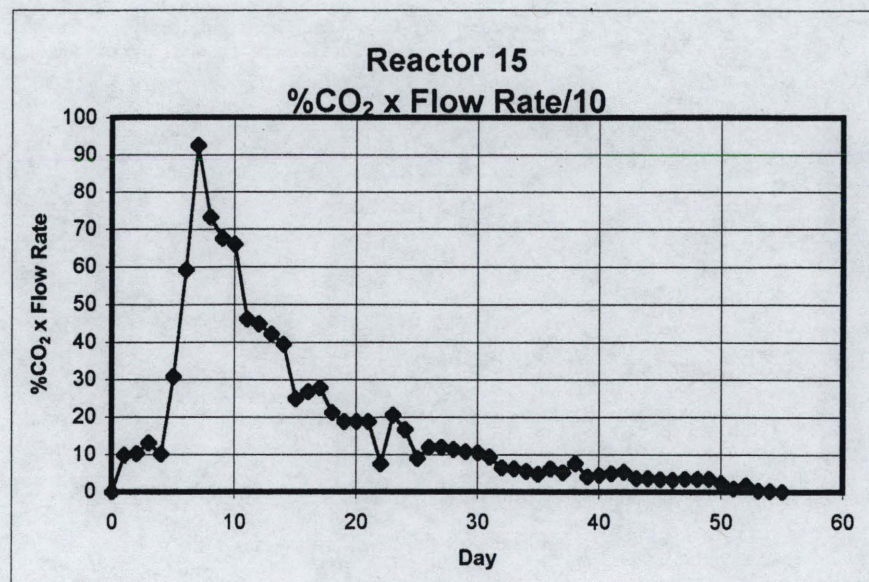
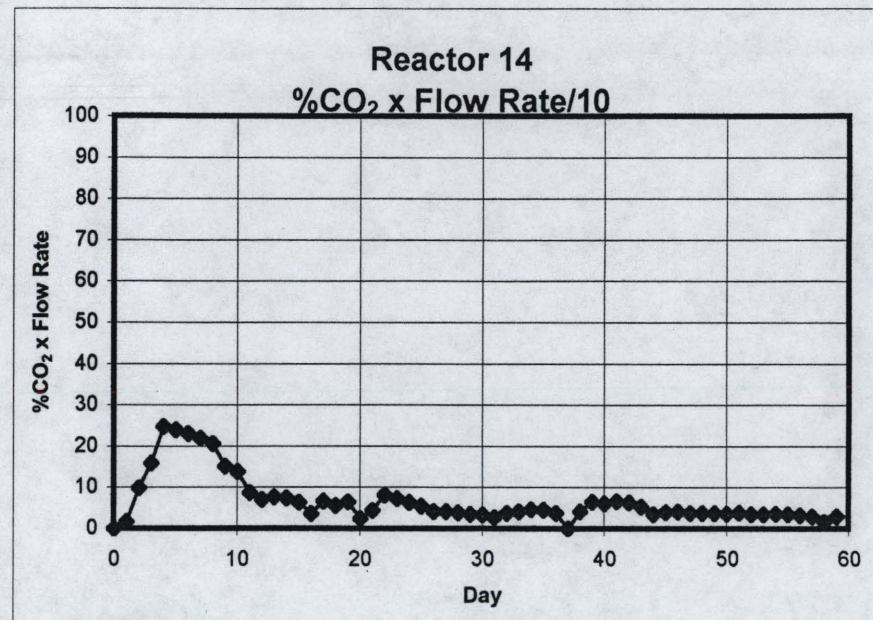
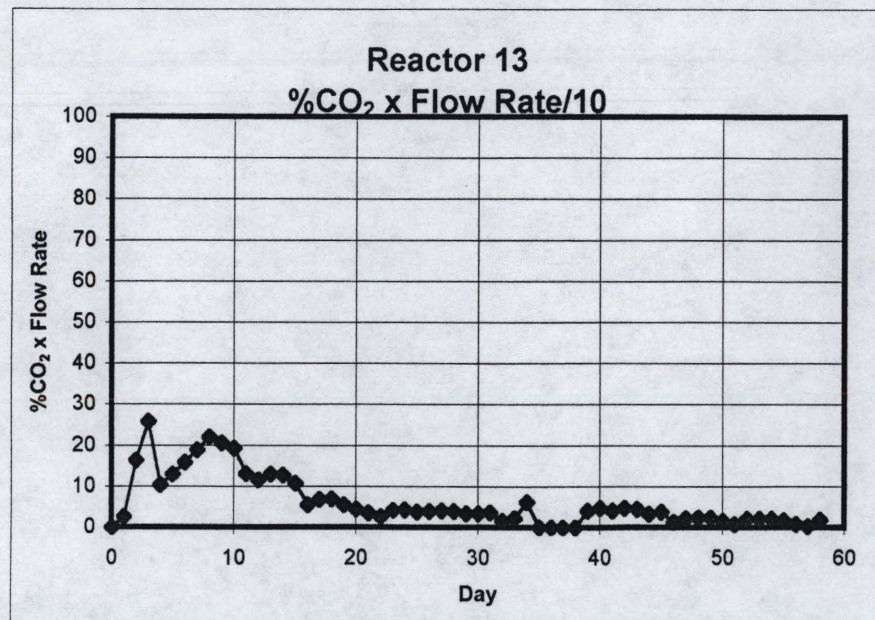




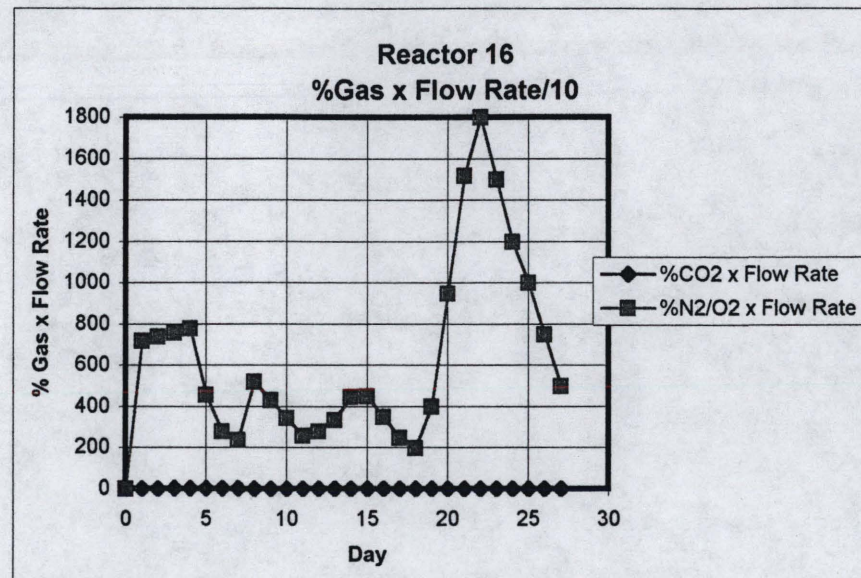






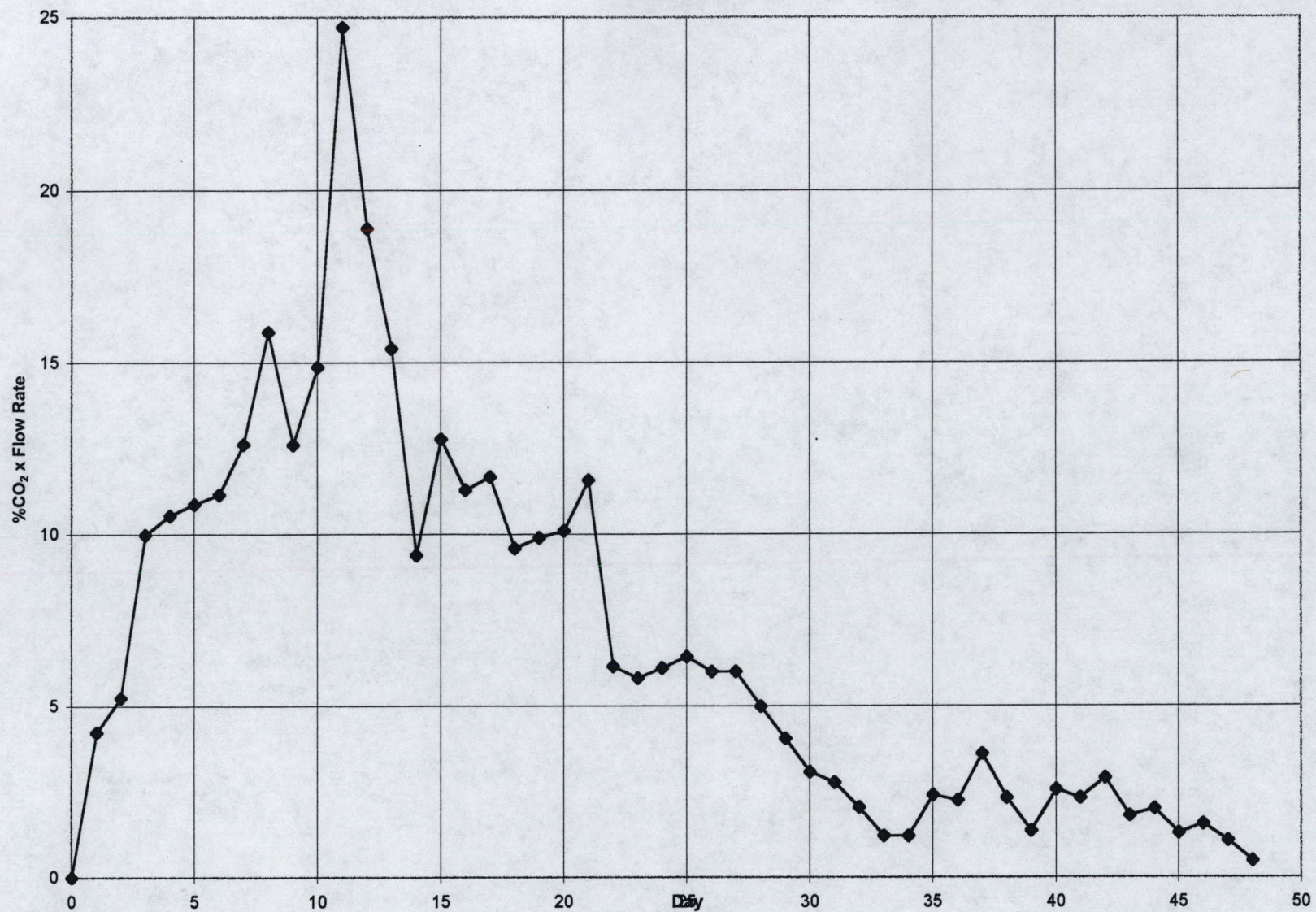




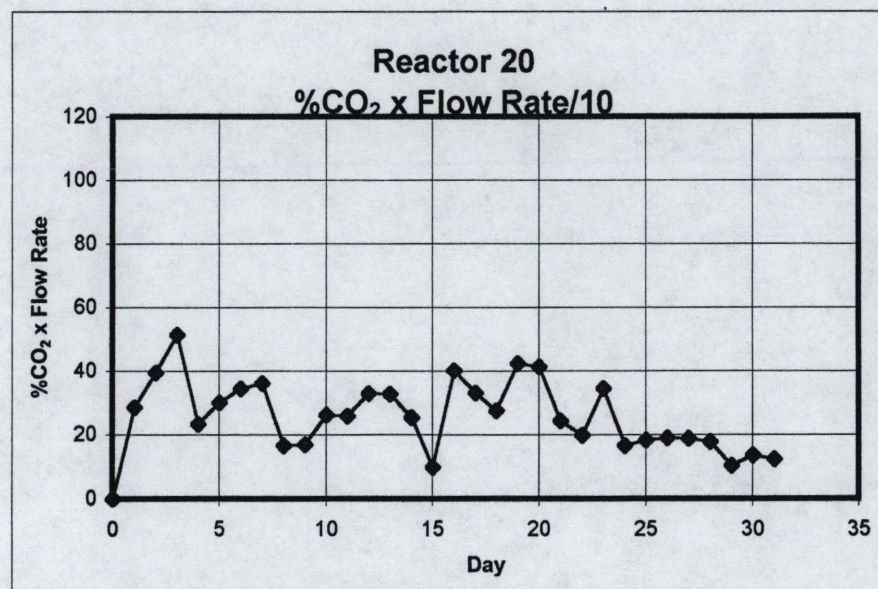
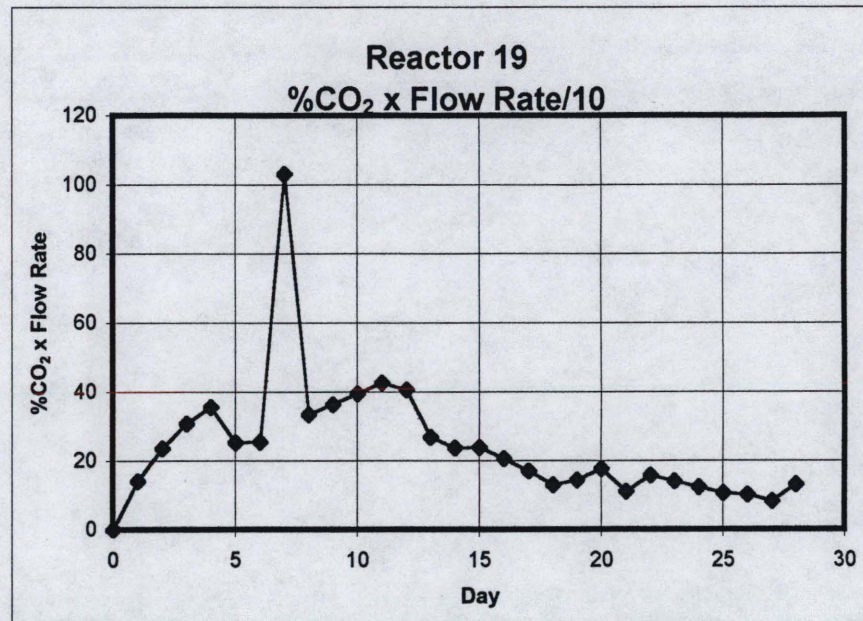
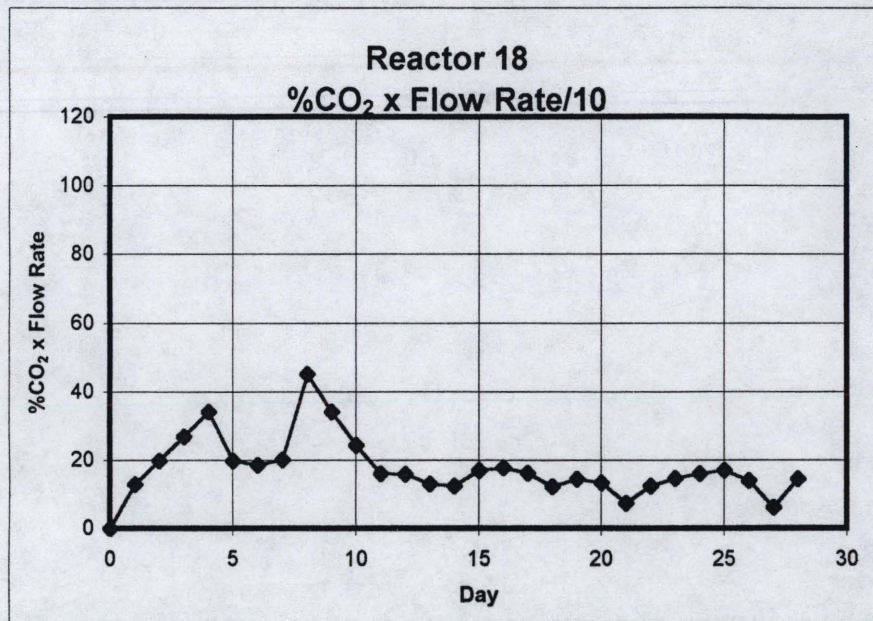




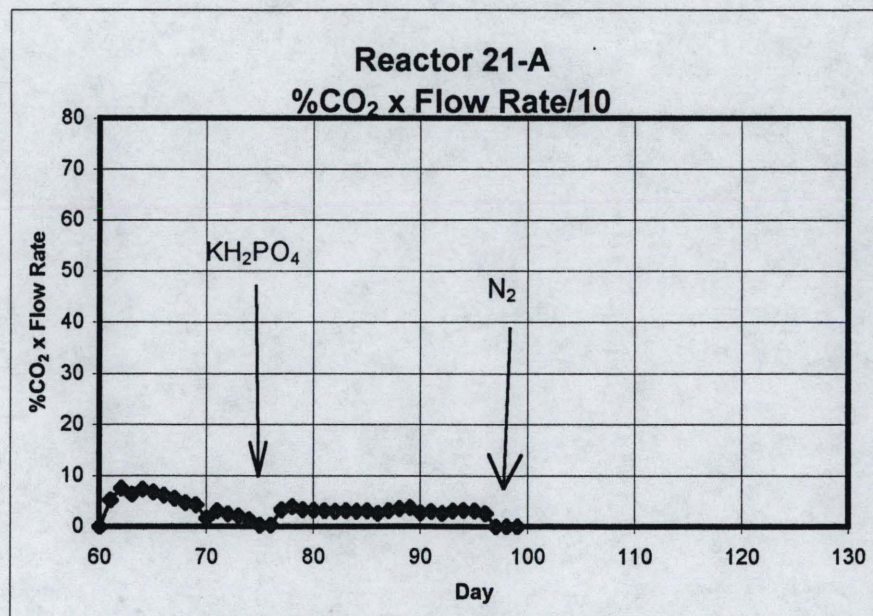
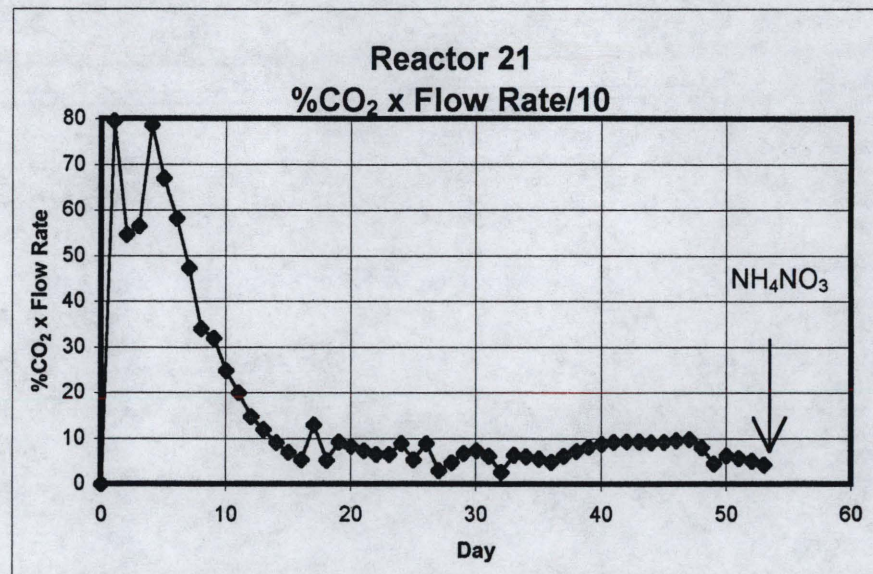
Reactor 17  
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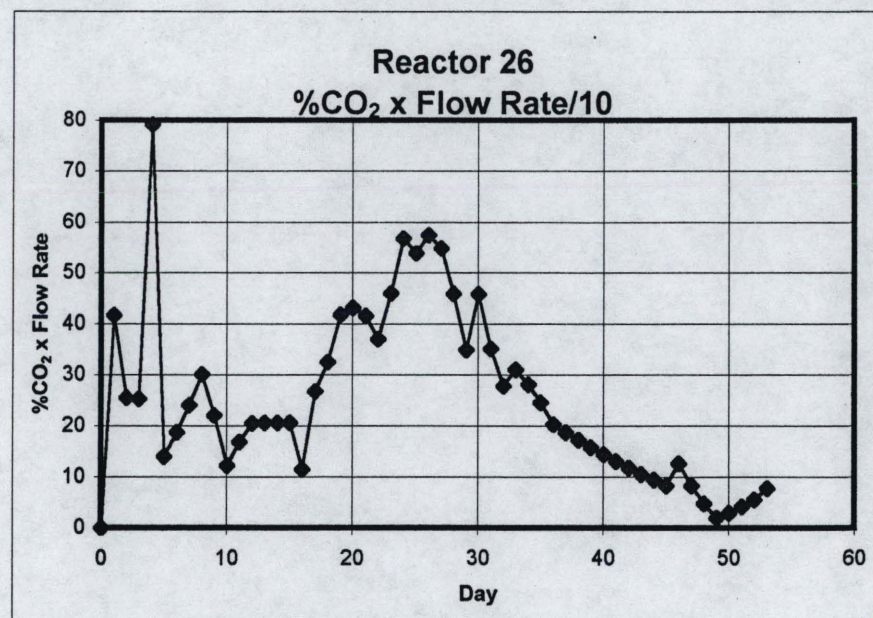
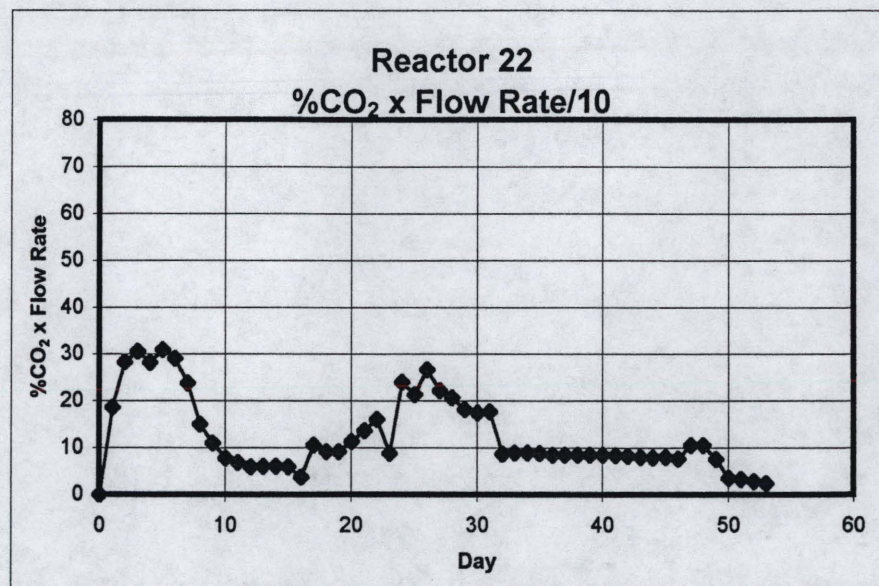




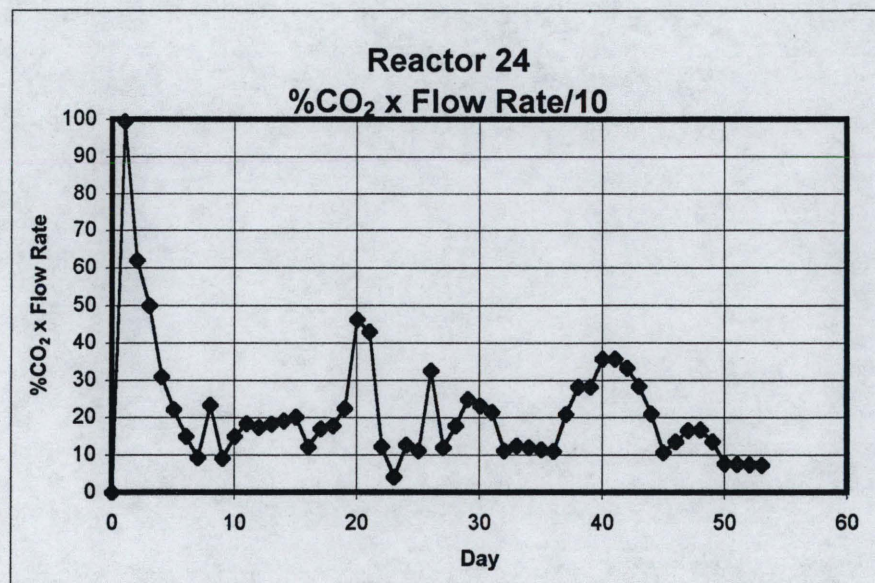
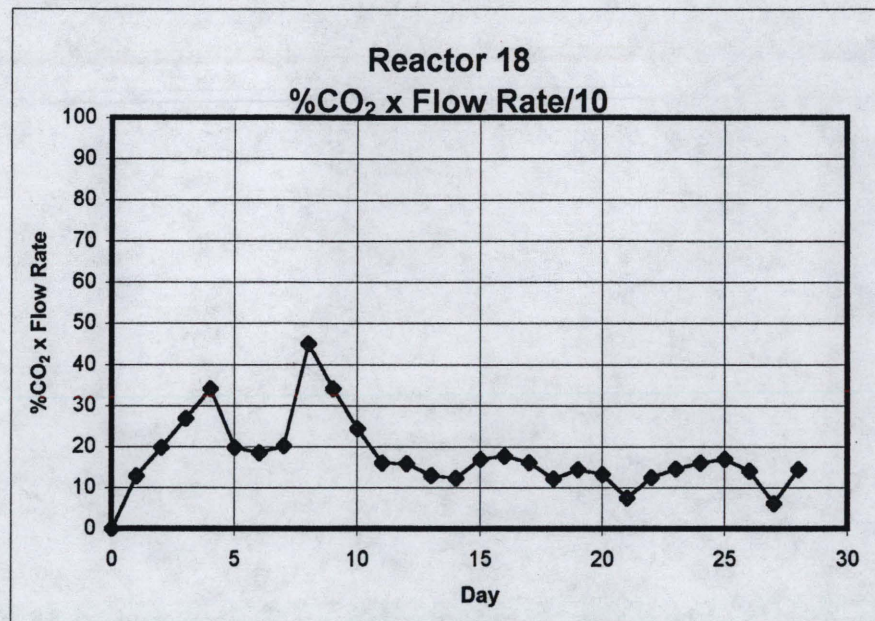




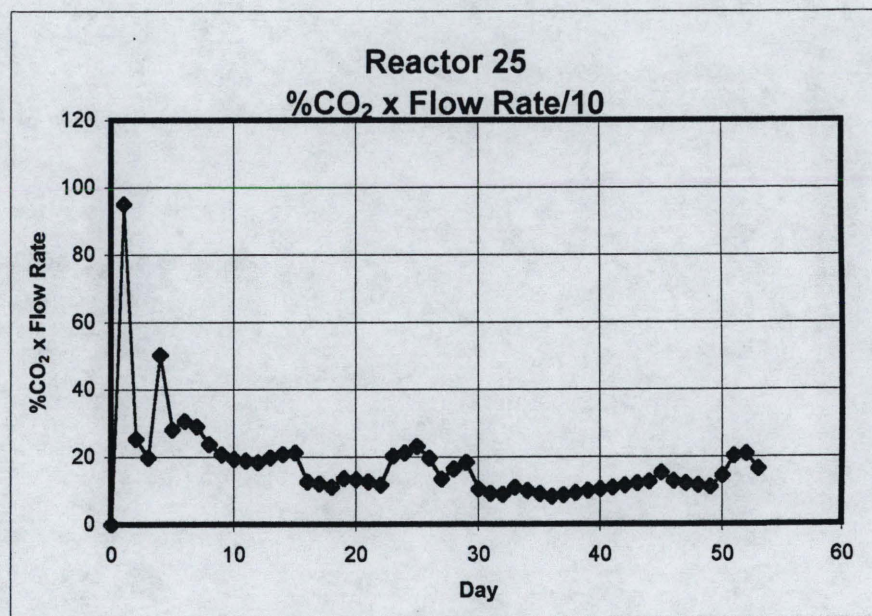
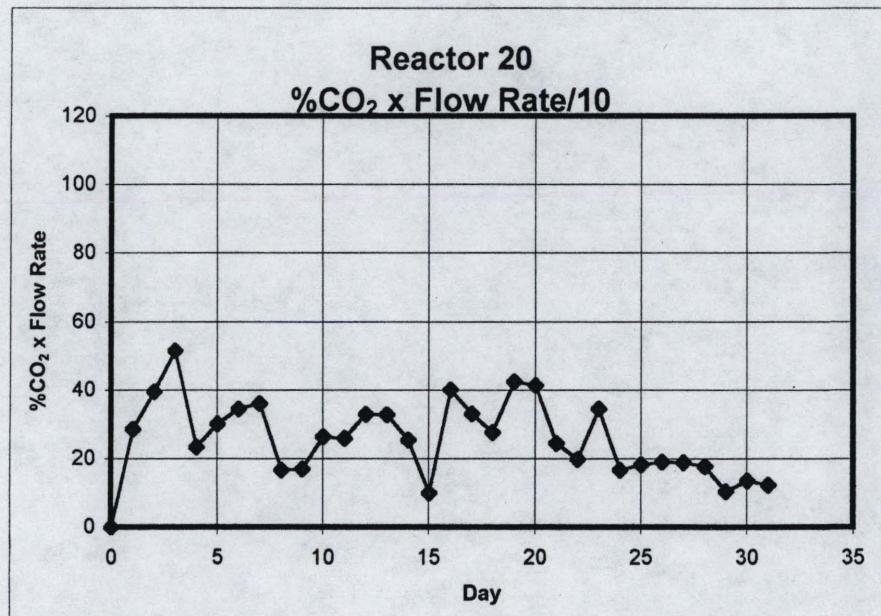
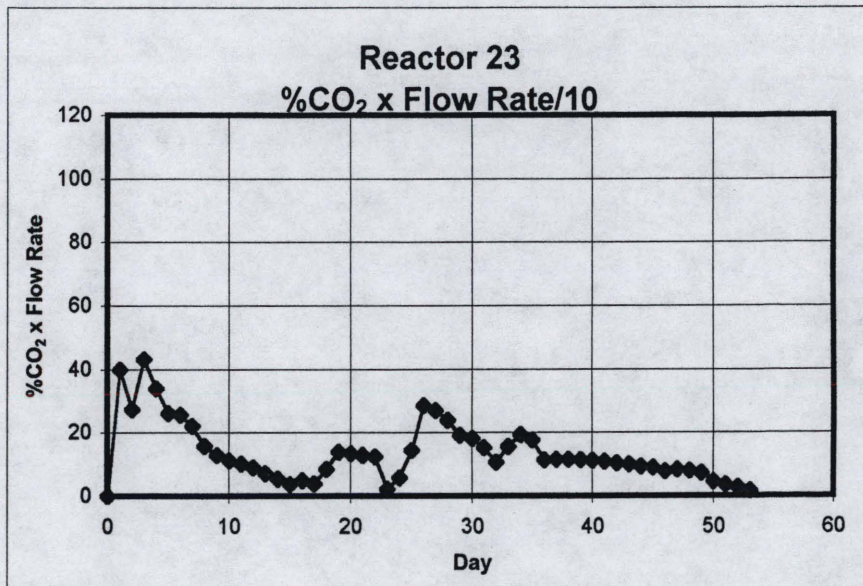
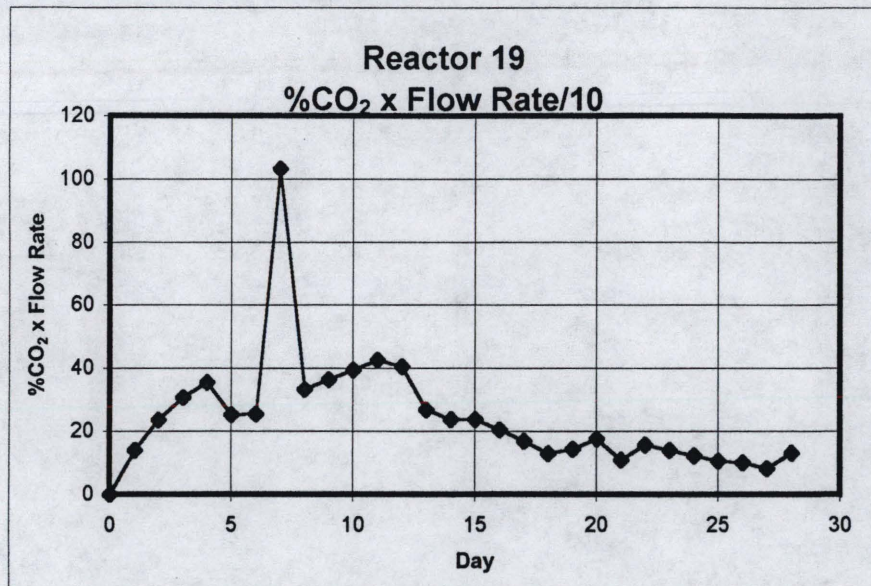














Reactor Numbers <sup>1</sup>	Gas Produced mmHg (x1000)	Methane %	Ash %		% Converted Dry weight <sup>2</sup>	Total Days	Odors of Dried Reactants
			t <sub>o</sub>	t <sub>f</sub>			
A1	8.30	64.83	29.6	43.3	59.12	175	Fresh Barn Yard
A2	6.10	55.06	29.6	37.5	55.78	175	Fresh Barn Yard
A3	10.00	59.22	29.6	41.8	36.12	160	Fresh Dirt
A4	9.90	61.04	29.6	49.3	32.96	160	Fresh Dirt
A5	0.60	35.45	43.7	63.3	39.34	130	Fresh Barn Yard
A6	2.10	32.93	43.7	56.7	43.7	130	Fresh Barn Yard
A7	2.40	44.13	49.2	58.5	50.1	115	Foul/Sour Fermentation Smell
A8	0.90	0.01	45.6	51.9	14.56	53	Sharp Cheese/ Butyrates
A9	0.00	0.00	45.6	46.8	0	67	Fresh Dirt
A10	0.40	0.00	45.6	49.7	28.35	53	Sharp Cheese /Butyrates
A11	3.05	40.38	48.7	65.9	31.64	132	Earthy Smell
A12	1.67	0.08	53.4	41.3	28.11	132	Sharp Cheese/Butyrates
A13	0.87	3.95	ND	ND	ND	45**	Fermentation Smell
A14	0.08	0.00	52.5	52.4	0	47	Native Smell
A15	0.03	0.00	51.8	51.7	0	41	Foul Smell
A16	5.38	5.98	67.6	76.7	10.87	41	Fresh Dirt
A17	0.00	0.00	58.8	58.8	0	30	
A18	8.19	38.77	66.58	76.4	10.65	90	Earthy Smell/Fermentation
A19	9.96	43.92	54.45	TBD	TBD	74*	Earthy Smell/Fermentation
A20	9.80	46.79	54.45	TBD	TBD	74*	Earthy Smell/Fermentation
A21	10.25	66.59	54.45	TBD	TBD	74*	Earthy Smell/Fermentation

<sup>1</sup>See Table 1 and corresponding figures for gas profiles.

<sup>2</sup>Percentage of t<sub>o</sub> and t<sub>f</sub> dry weights of starting material and finished product, respectively.

\* Still in operation; \*\* Stopped and used for seed culture; TBD to be determined; ND not determined



Calculation of Results of Anaerobic Reactors Table 4

Reactor Number <sup>a</sup>	Gas Produced in Liters <sup>b</sup>	Methane Liters <sup>d</sup>	Methane ft <sup>3</sup> kg <sup>-1</sup> <sup>c</sup>	BTU Kg <sup>-1</sup> wet wt <sup>e</sup>	BTU Kg <sup>-1</sup> dry wt	BTU Kg <sup>-1</sup> TVS <sup>h</sup>	\$Value per metric wet ton <sup>g</sup>
A1	49.8	32.3	0.72	704			3.31
A2	36.3	20	0.84	821			4.00
A3	60.0	35.5	2.22	2,152	4,304	6,114	10.00
A4	59.4	36.3	2.30	2,249	4,498	6,335	11.00
A5	31.0	1.1	0.07	69			0.32
A6	12.7	4.2	0.41	401			1.88
A7	14.3	6.3	0.61	597			2.81
A8	5.1	0	0	0			0.00
A9	<0.01	0	0	0			0.00
A10	2.4	0	0	0			0.00
A11	18.3	7.39	0.41	398			1.87
A12	10.0	0.79	NA	NA			NA
A13	5.2	0.2	NA	NA			NA
A14	0.5	0	NA	NA			NA
A15	0.2	0	NA	NA			NA
A16	32.3	1.9	0.03	34			0.16
A17	0	0	NA	NA			NA
A18	41	22.13	0.36	348	1,015	3,026	1.64
A19	39.8	17.48	0.65	637	1,490	3,276	2.99
A20	39.2	18.34	0.69	675	1,584	3,433	3.17
A21	46.3	30.83	1.08	1,056	2,396	5,281	4.96

\*Reactors still in operation

<sup>a</sup>See Table 1 for identification of reaction mixtures

<sup>b</sup>Gas measured by displacement volume in graduated cylinders

<sup>c</sup>moles calculated from mm Hg, see corresponding reactor figures

<sup>d</sup>Percentage determined by Gas Chromatographic profile, see reactor figures of mole-%

<sup>e</sup>One Liter = 0.035 ft<sup>3</sup>; Kg represent wet weight of 35-50% water and ash content of 20-68%

<sup>f</sup>978Btu/ft<sup>3</sup>

<sup>g</sup>100,000 Btu's = \$0.47, approximately.

<sup>h</sup>TVS = Total Volatile Solids



**Nitrogen, Carbon and Hydrogen Analyses****Table 5**

Sample	ug N	%N	ug C	% C	ug H	% H	g N added	g Final N	% Remaining
A1	38.08	0.64	1968.09	33.17	205.91	3.47	10.6	12.08	114
A2	51.36	0.57	3057.69	33.99	346.9	3.86	8.11	5.45	67.2
A3	60.72	0.74	2829.09	34.28	316.5	3.83	6.69	4.78	71.5
A4	75.87	0.89	2805.21	32.84	320.15	3.75	7.05	5.55	78.7
A5	86.34	1.3	1773.45	26.76	203.08	3.06	8	7.94	99.2
A6	67.88	0.92	2105.18	28.38	236.83	3.19	8	8	100
A8	31.95	0.58	1583.45	28.95	170.85	3.12	8	8	100



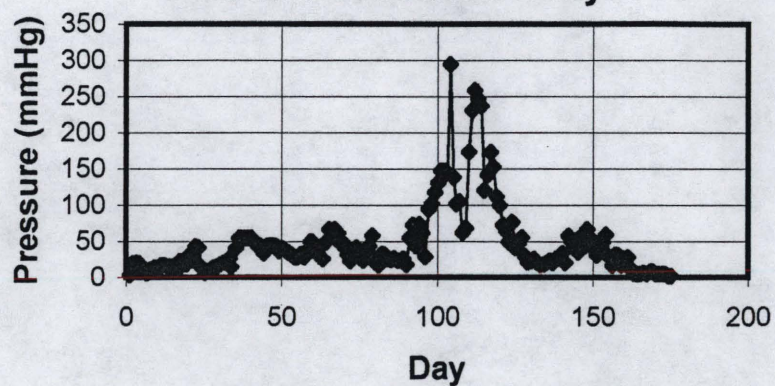
# Organism Counts

Table 6

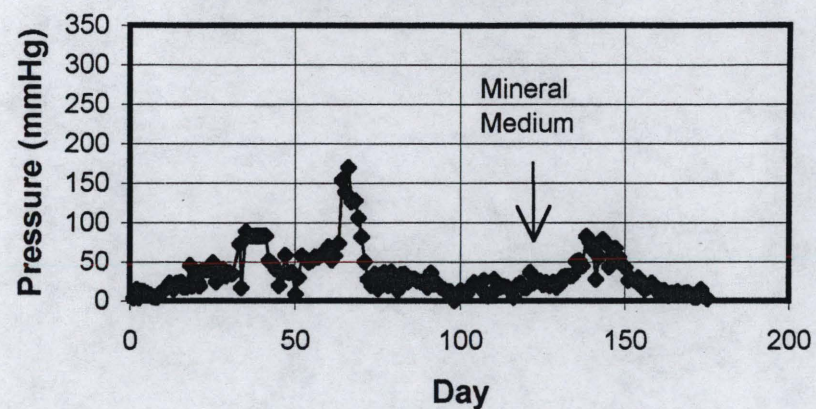
$T_0$	1/10 TSA (organisms/g)	1/50 TSA (organisms/g)	Potato Dextrose with Streptomycin (organisms/g)	Potato Dextrose with Ampicillin (organisms/g)
Reactor 8	TNTC	5.7E+3	18.0E+3	9.8E+3
Reactor 10	TNTC	5.7E+3	18.0E+3	9.8E+3
Reactor 11	TNTC	5.7E+3	18.0E+3	9.8E+3
Reactor 12	TNTC	5.7E+3	18.0E+3	9.8E+3
$T_{12}$				
Reactor 8	2.8E+11	2.7E+11	1E+11	TNTC
Reactor 10	2.4E+11	2E+11	1.4E+11	2.0E+9
Reactor 11	1.8E+11	1.7E+11	1.5E+11	4.1E+9
Reactor 12	1.6E+11	1.3E+11	1.2E+11	3.7E+9
$T_{31}$				
Reactor 8	1.9E+9	2.2E+9	2.90E+08	5.10E+08
Reactor 10	2.4E+9	2.6E+9	2.80E+08	2.10E+08
Reactor 11	2.70E+10	2.50E+10	1.4E+9	9.90E+08
Reactor 12	3.30E+10	3.20E+10	3.5E+9	1.6E+9



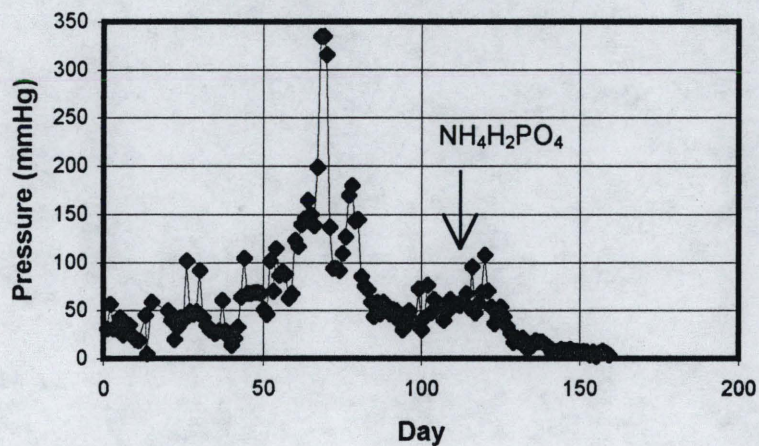
**Reactor A1**  
**Pressure Produced Daily**



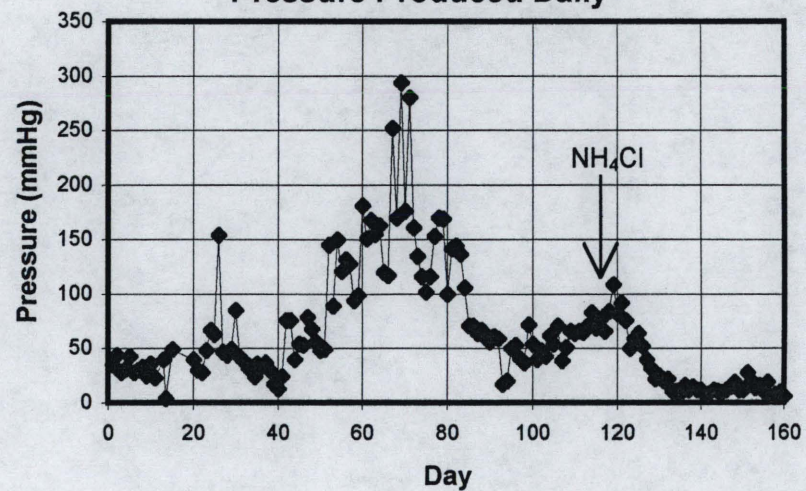
**Reactor A2**  
**Pressure Produced Daily**



**Reactor A3**  
**Pressure Produced Daily**

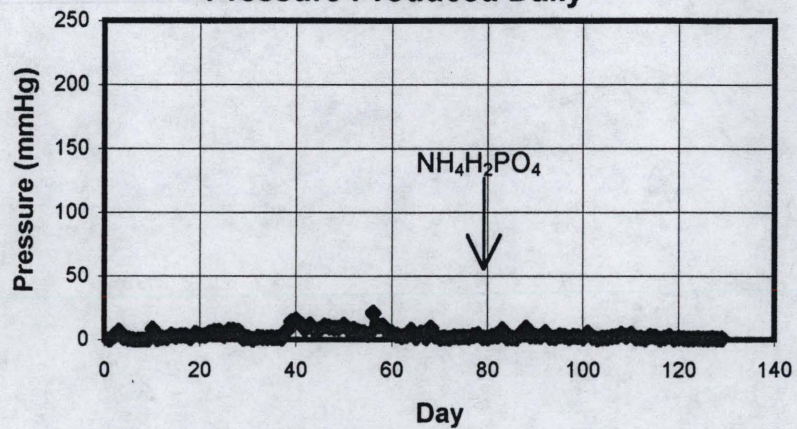


**Reactor A4**  
**Pressure Produced Daily**

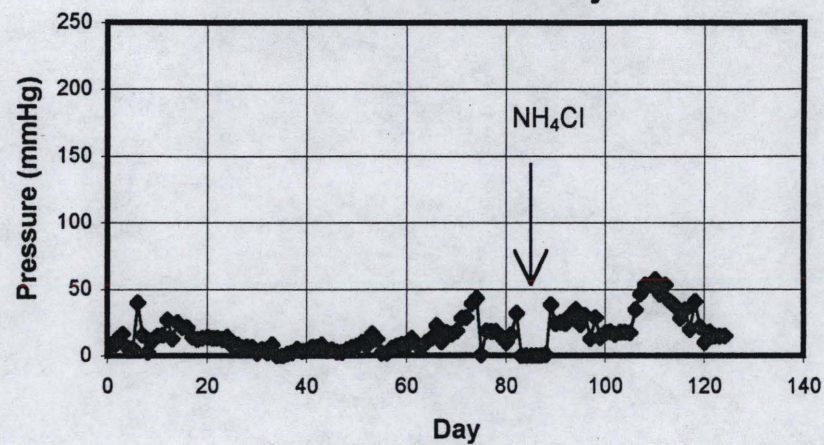




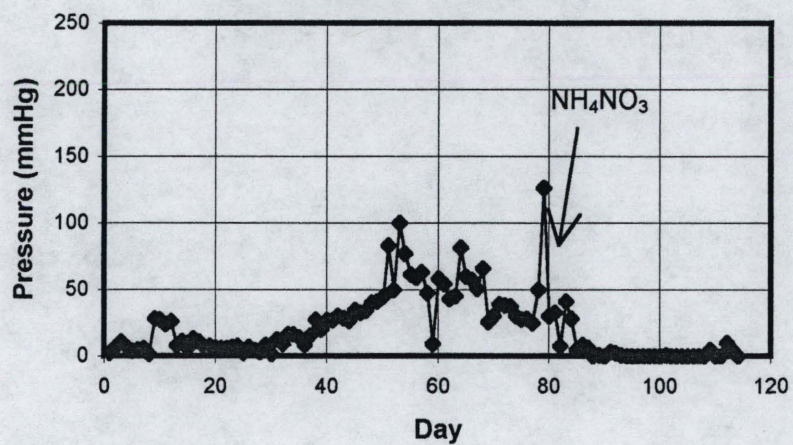
Reactor A5  
Pressure Produced Daily



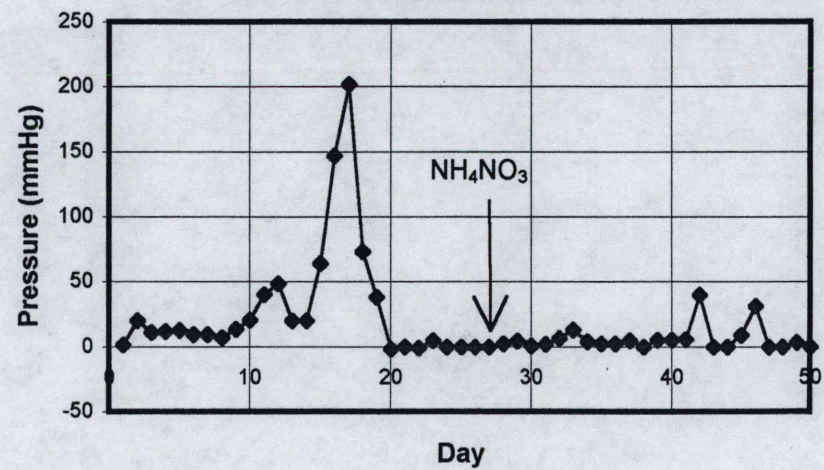
Reactor A6  
Pressure Produced Daily



Reactor 7  
Pressure Produced Daily

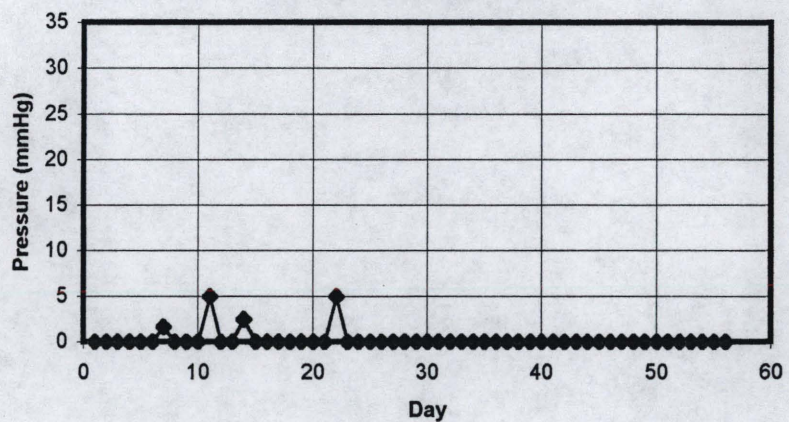


Reactor A8  
Pressure Produced Daily

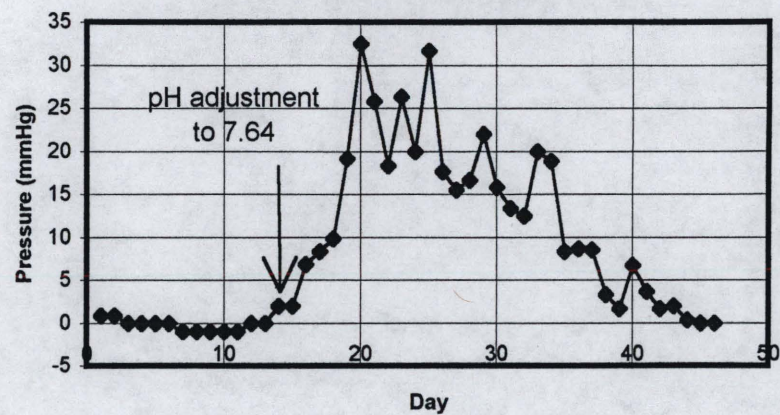




Reactor A9  
Pressure Produced Daily

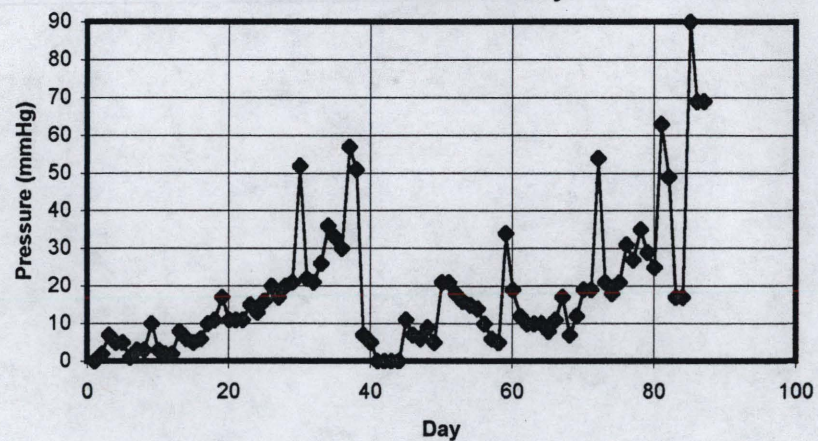


Reactor A10  
Pressure Produced Daily

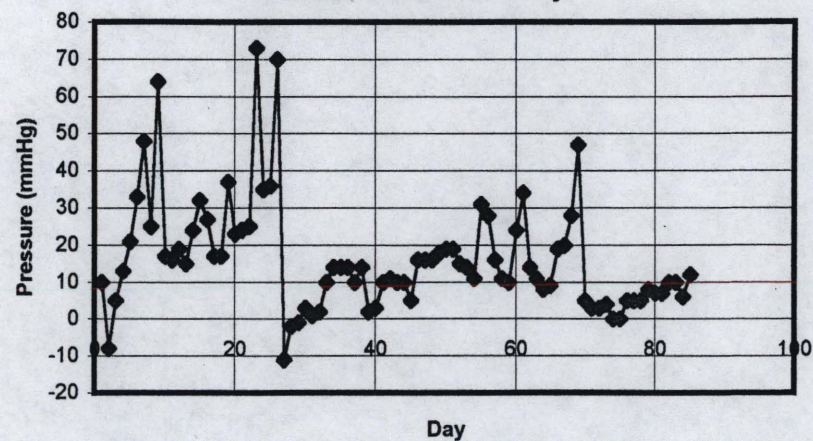




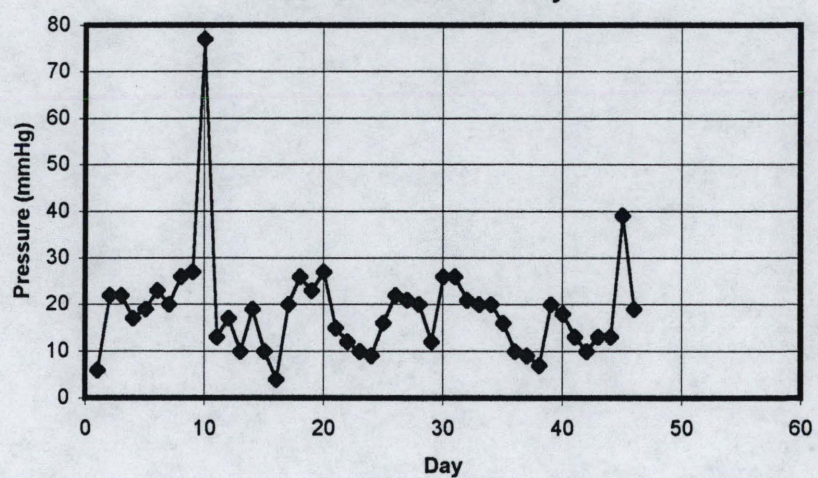
Reactor A11  
Pressure Produced Daily



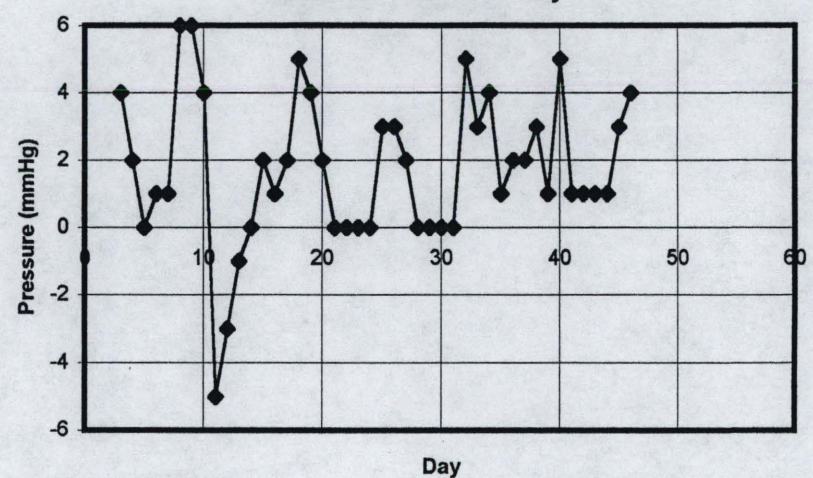
Reactor A12  
Pressure Produced Daily



Reactor A13  
Pressure Produced Daily

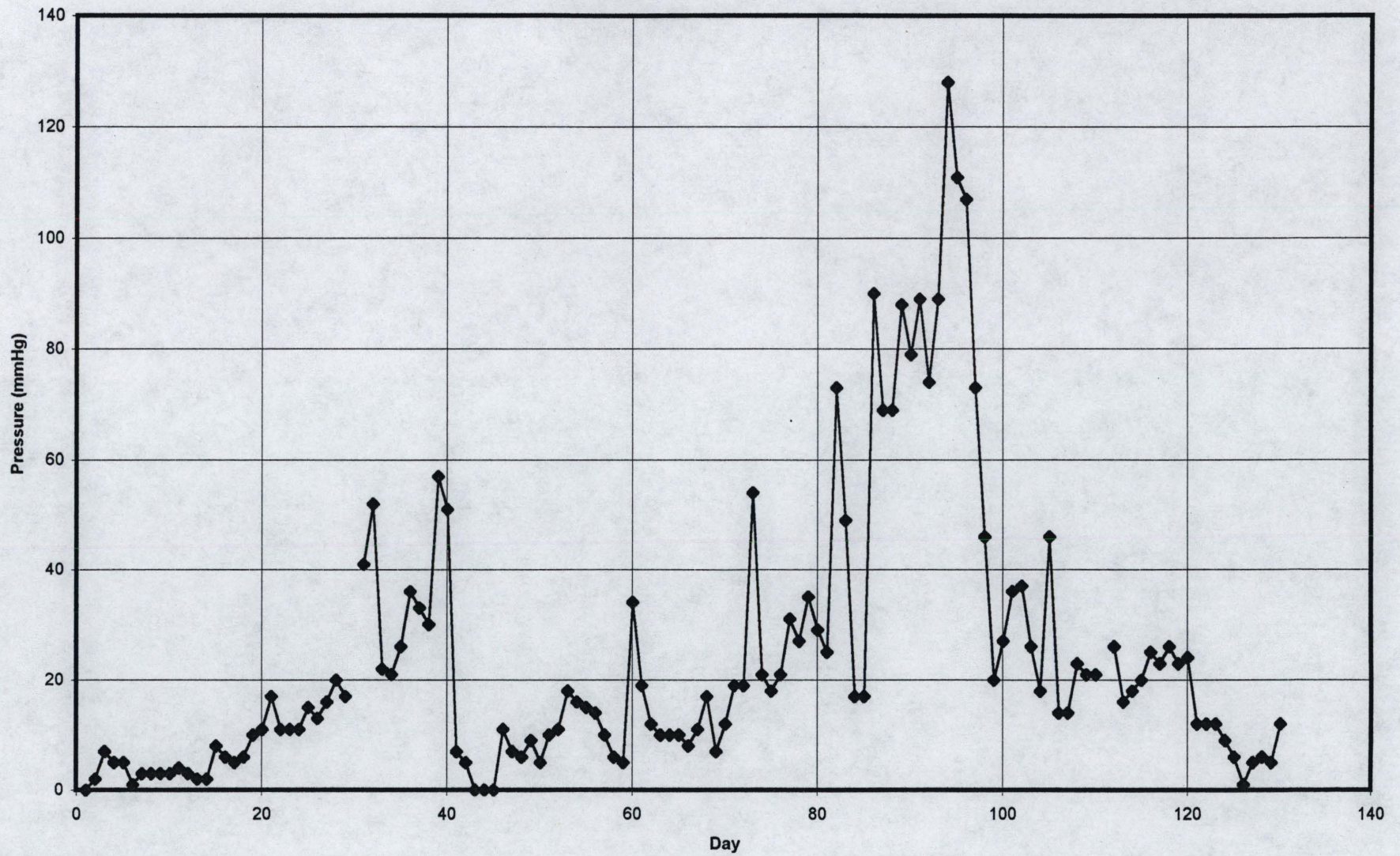


Reactor A14  
Pressure Produced Daily



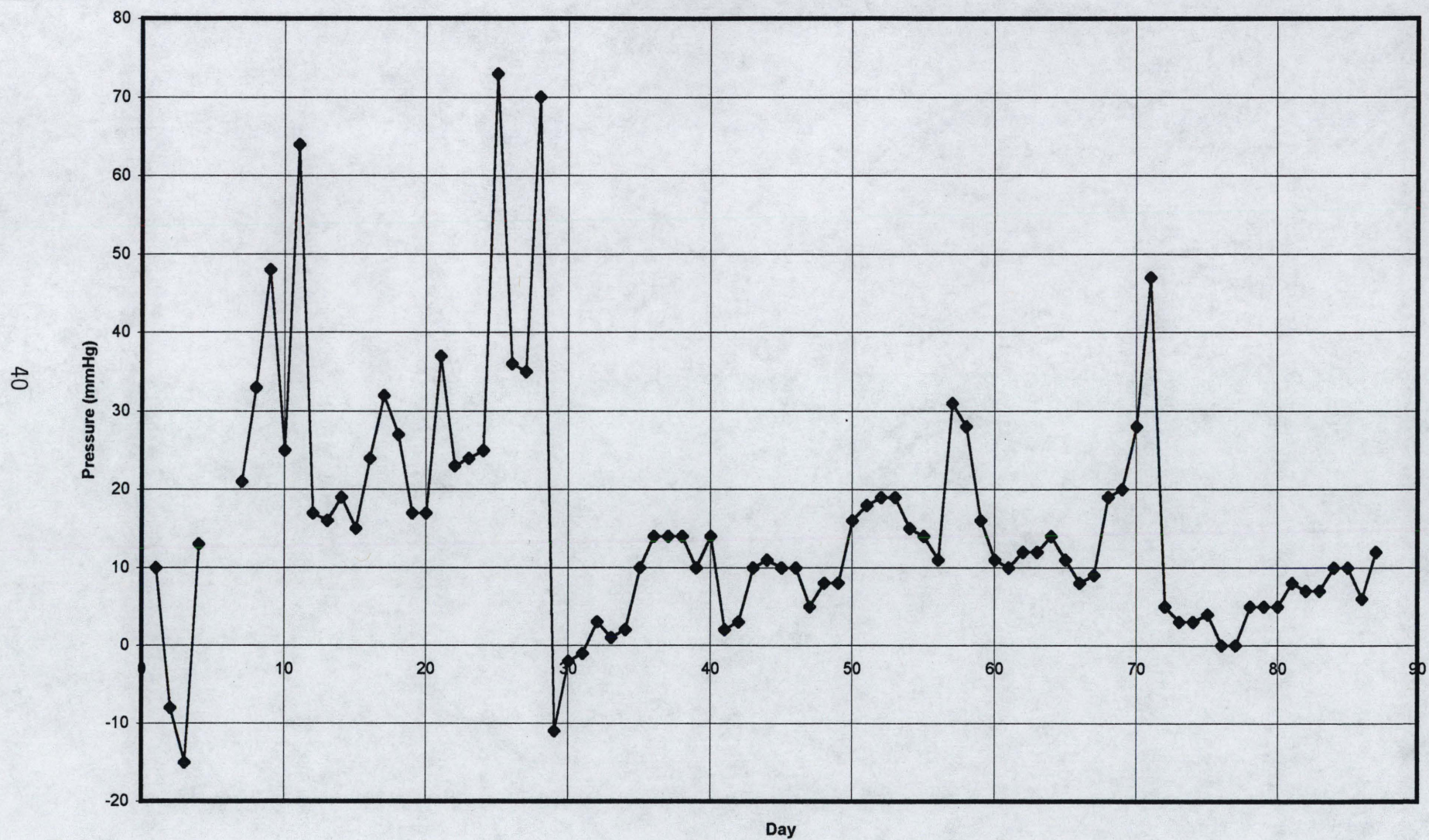


Reactor A11  
Pressure Produced Daily



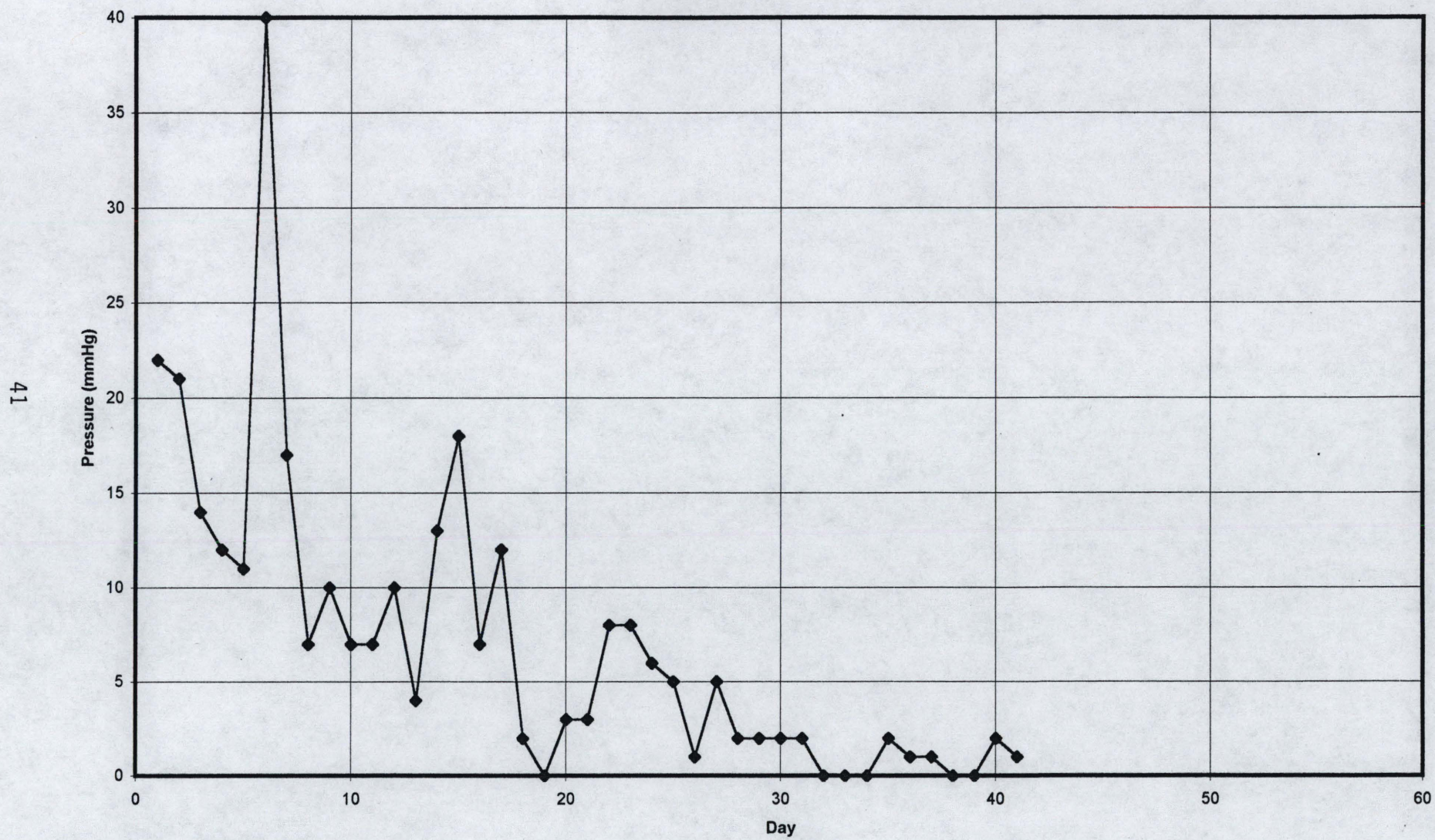


Reactor A12  
Pressure Produced Daily



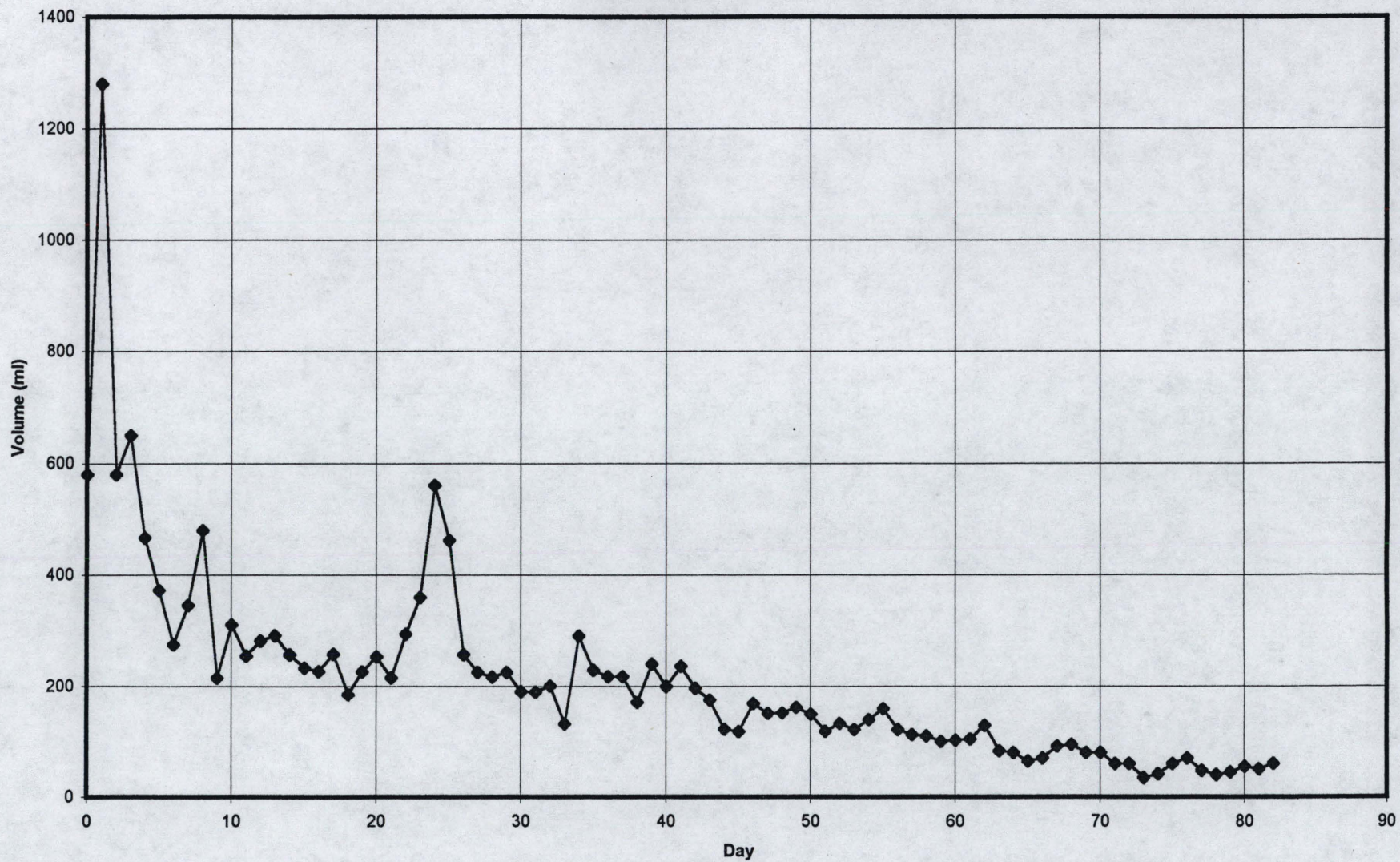


Reactor A15  
Pressure Produced Daily



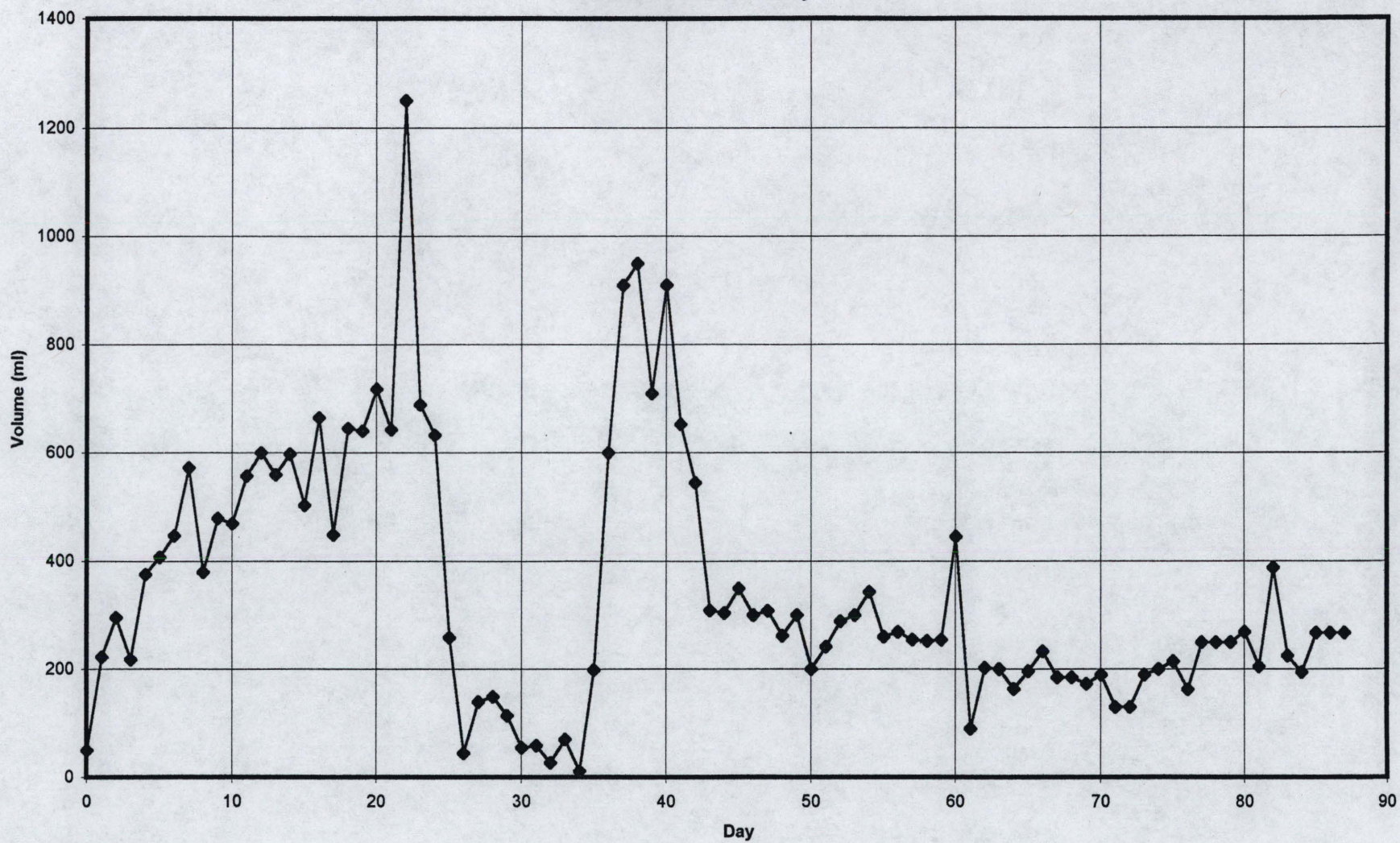


Reactor A16  
Volume Produced Daily





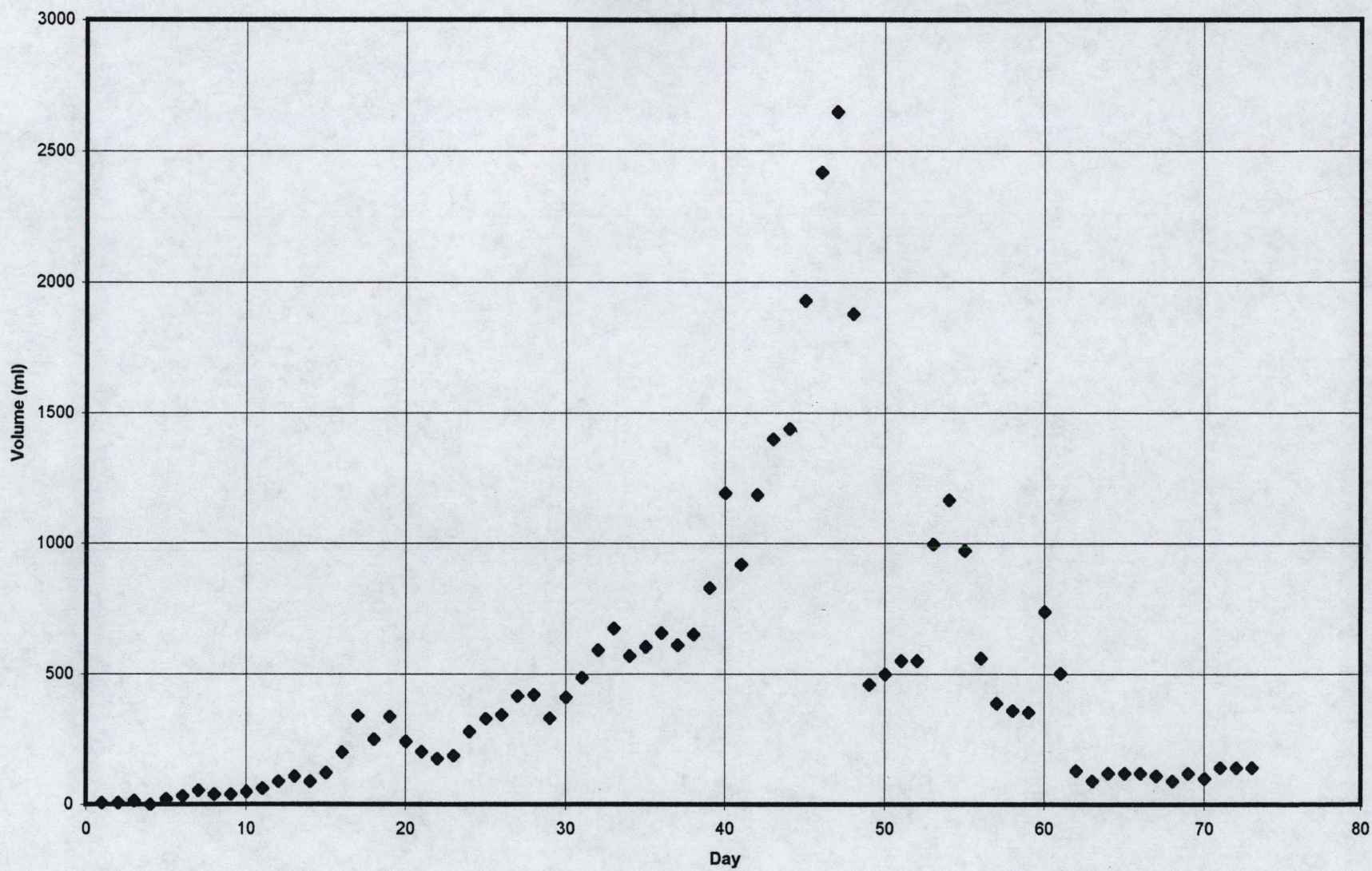
Reactor A18  
Volume Produced Daily





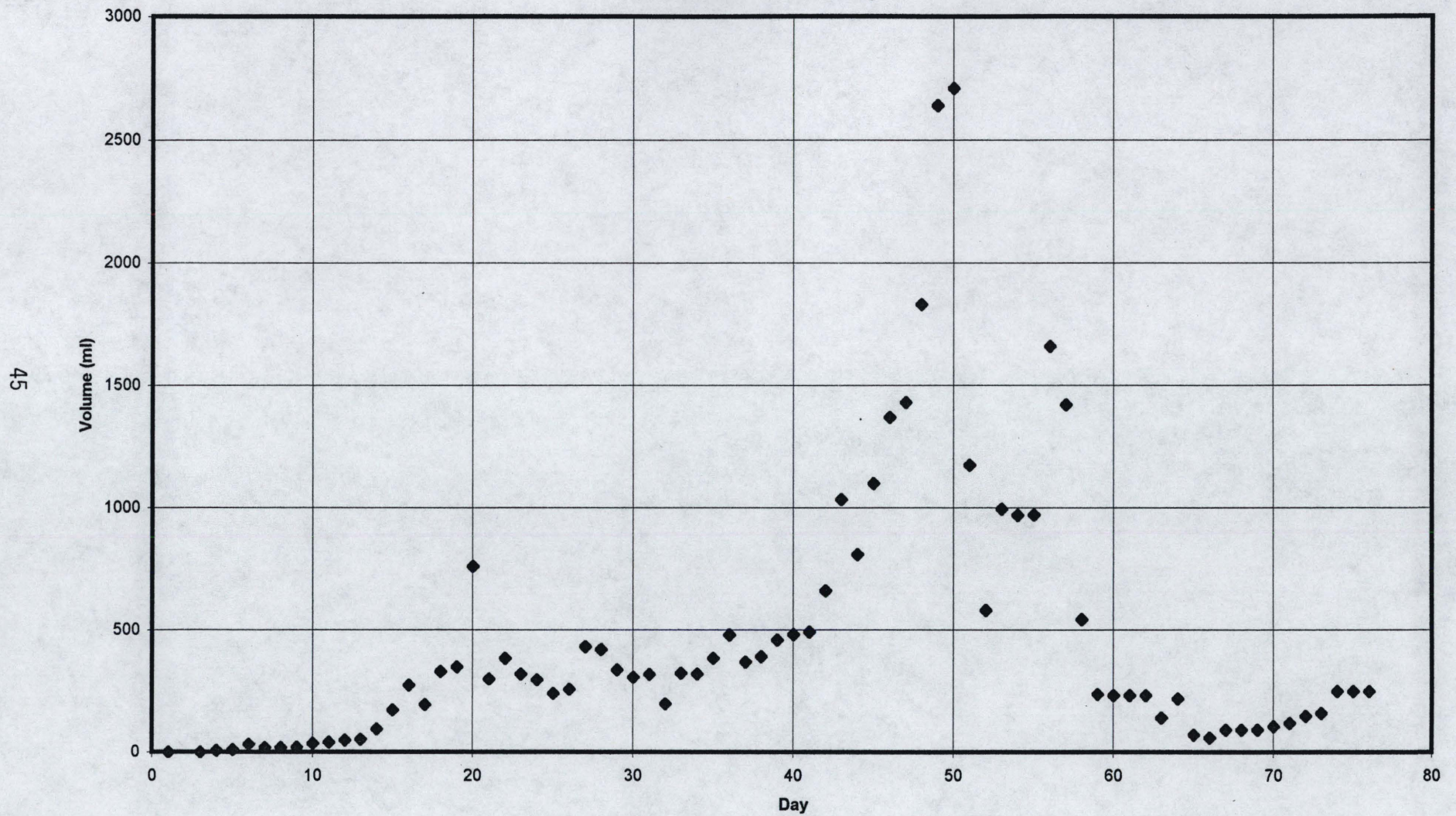
Reactor A19  
Volume Produced Daily

44



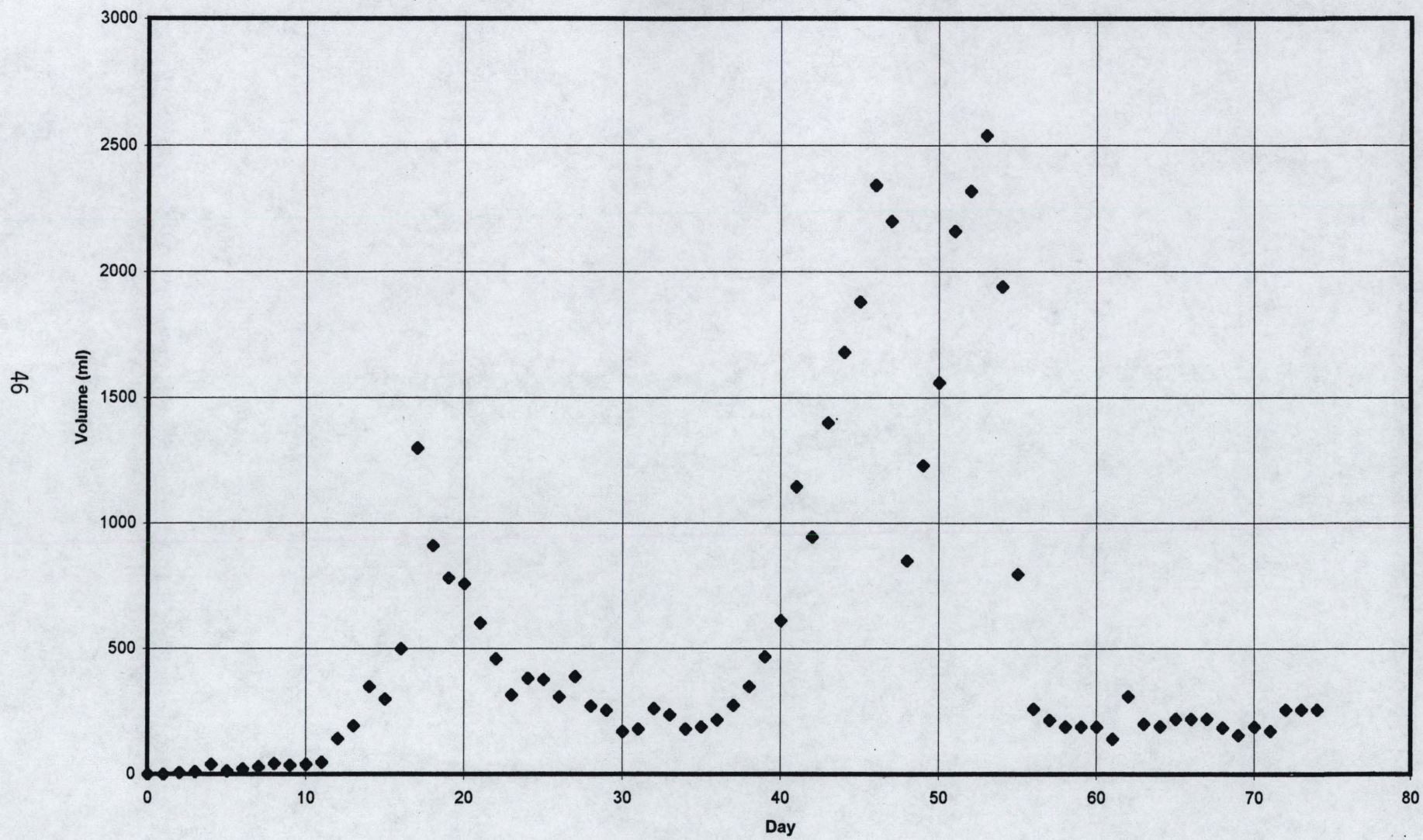


Reactor A20  
Volume Produced Daily



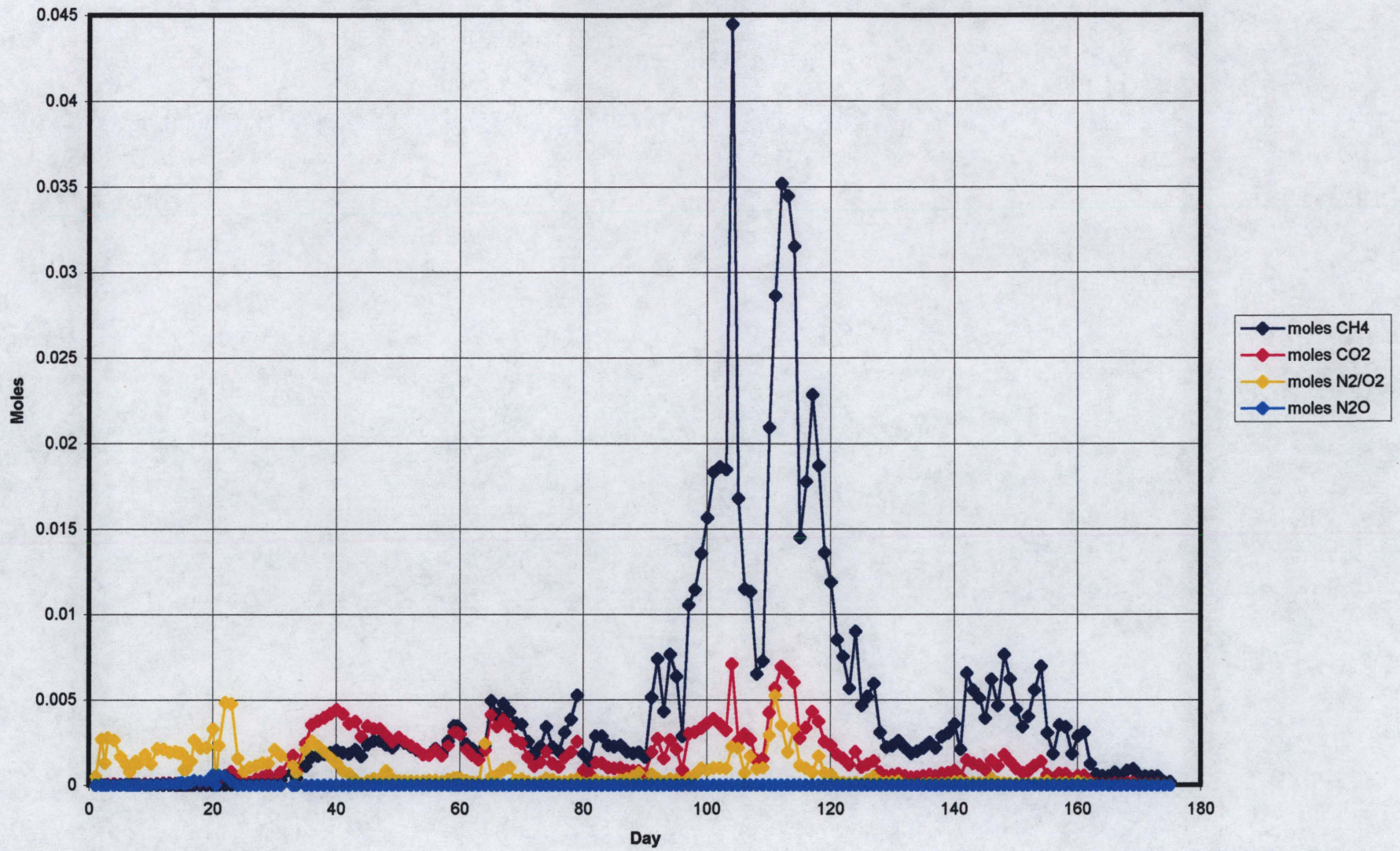


Reactor 21  
Volume Produced Daily





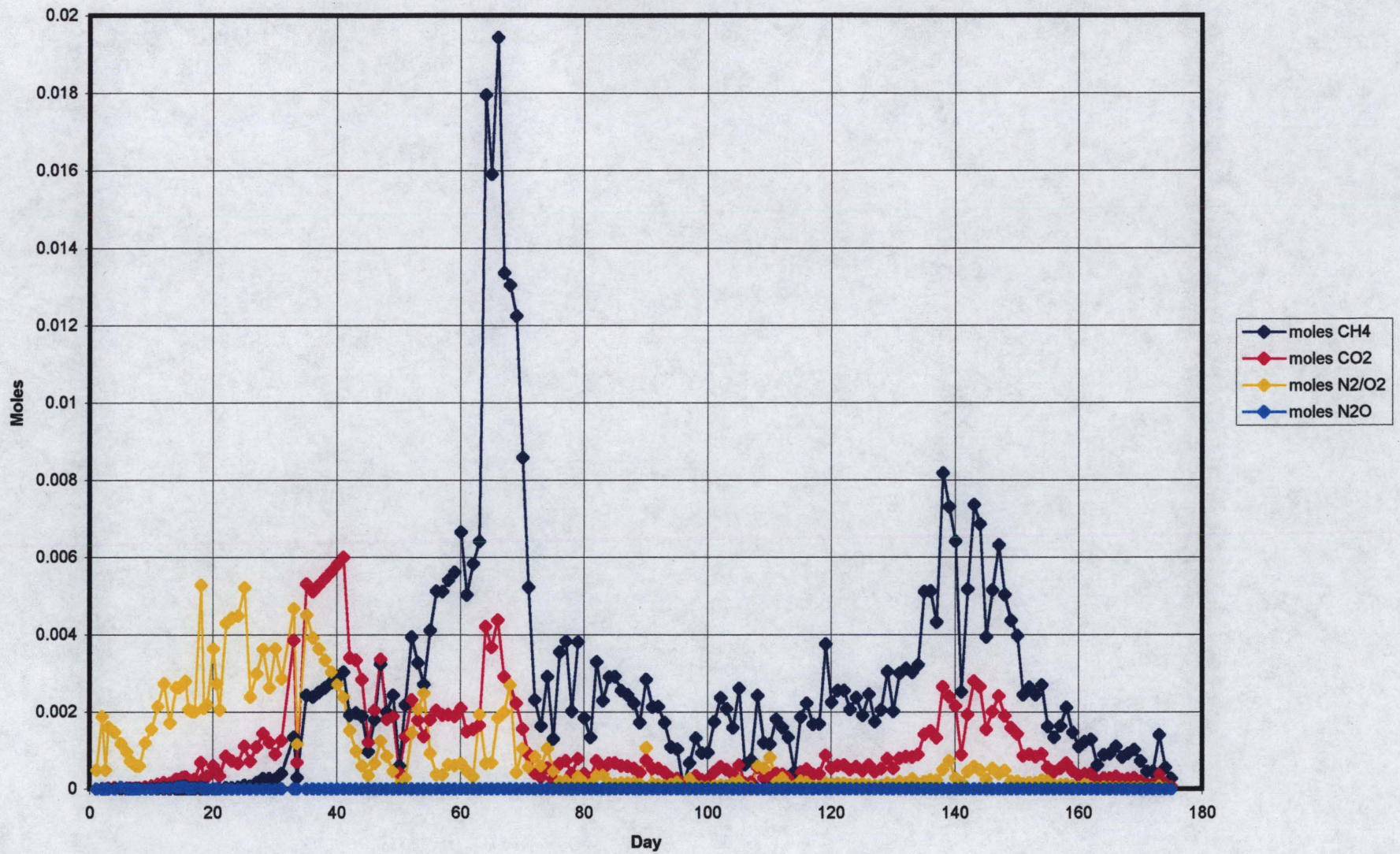
Reactor A1  
Moles Gas Produced Daily



67



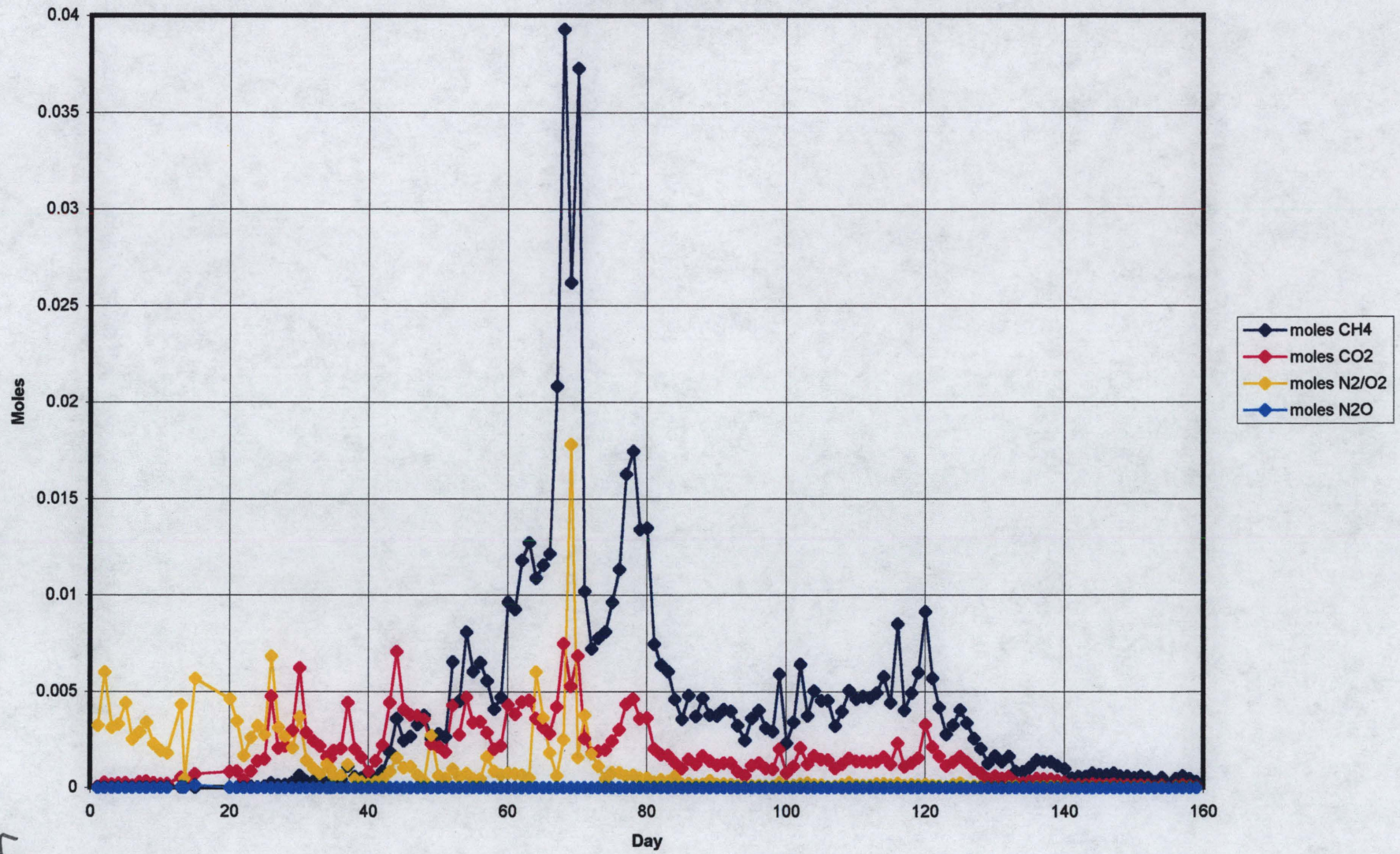
Reactor A2  
Moles Gas Produced Daily



48



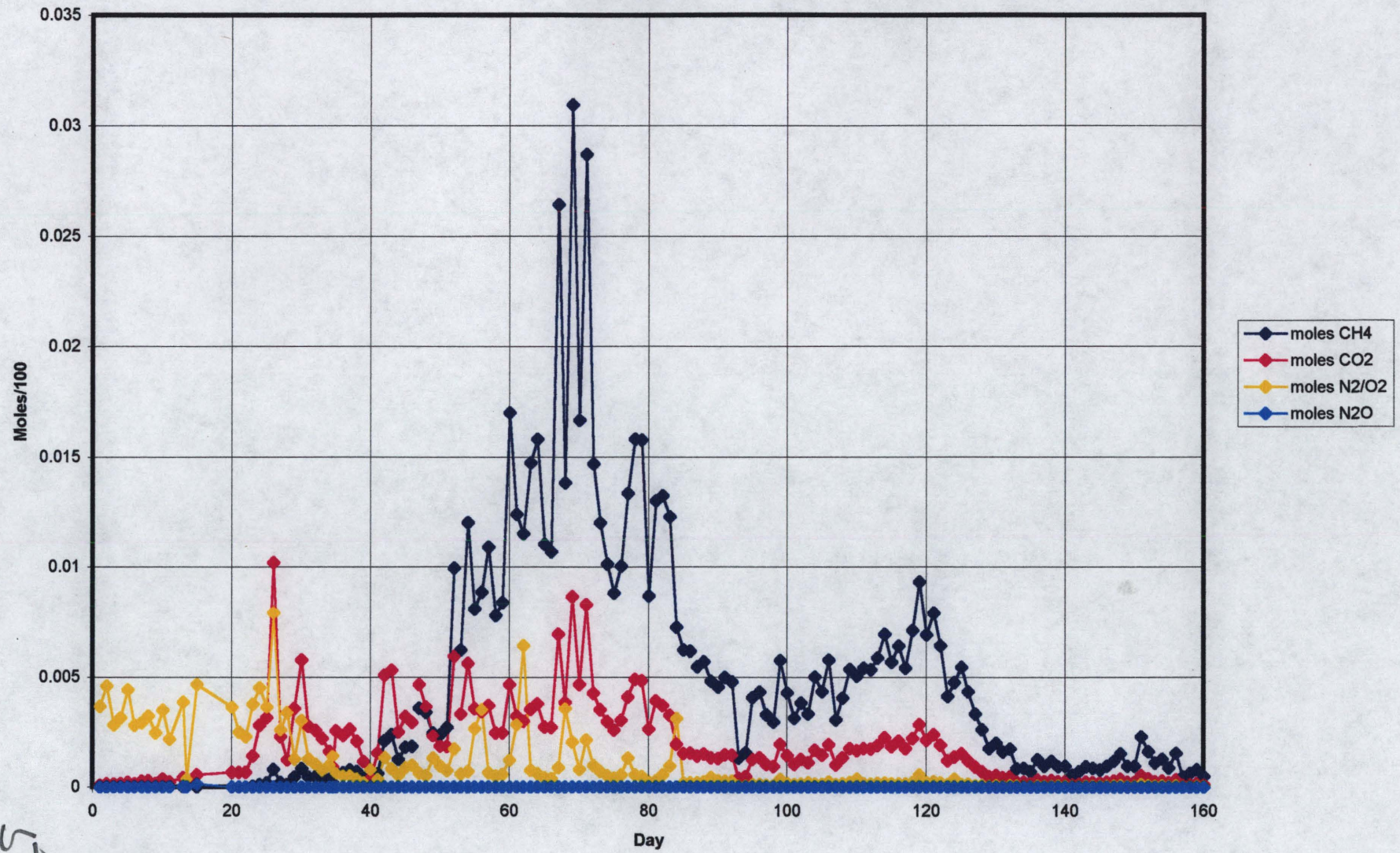
Reactor A3  
Moles Gas Produced Daily



47



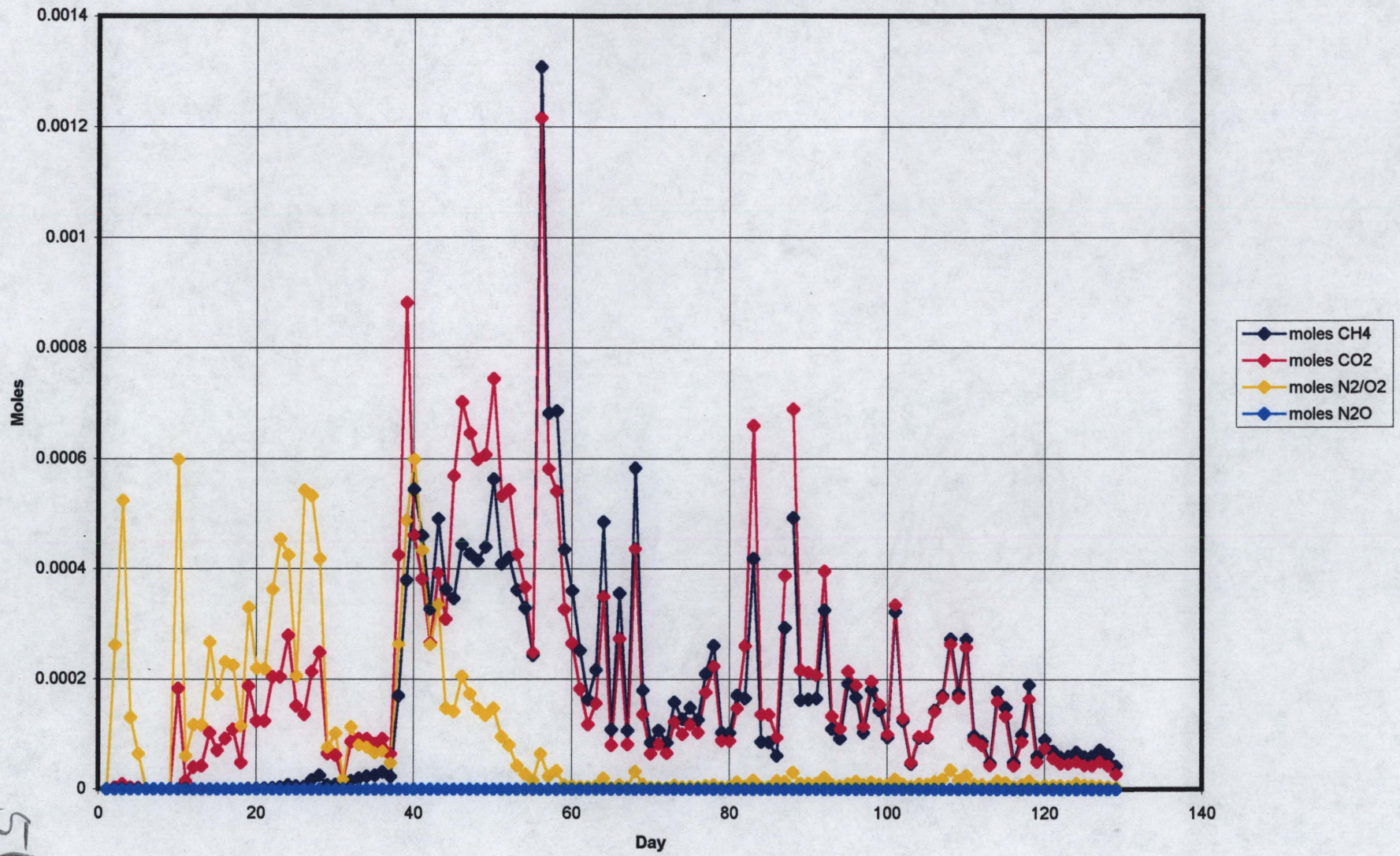
Reactor 4  
Moles Gas Produced Daily



57



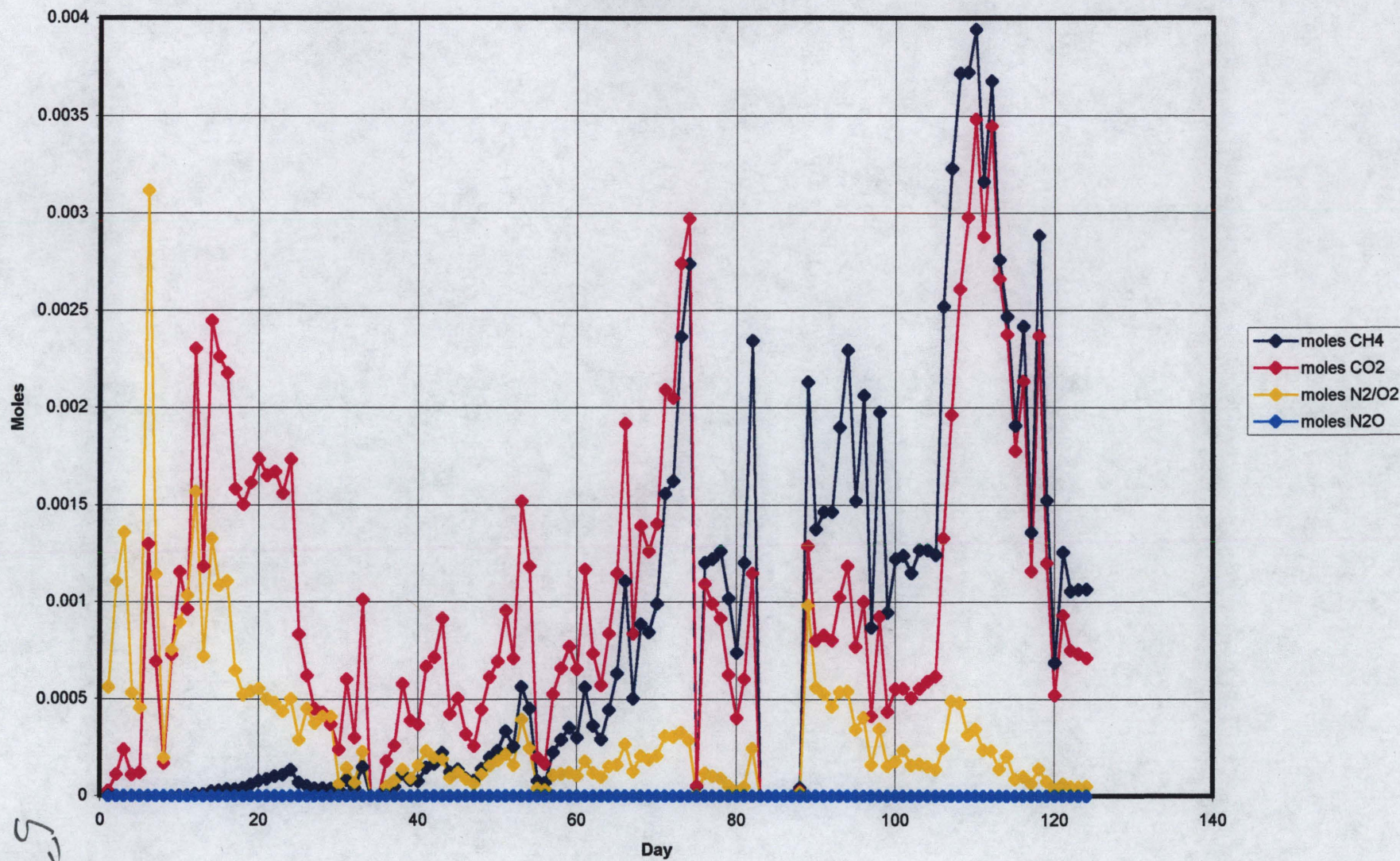
Reactor A5  
Moles Produced Daily



51



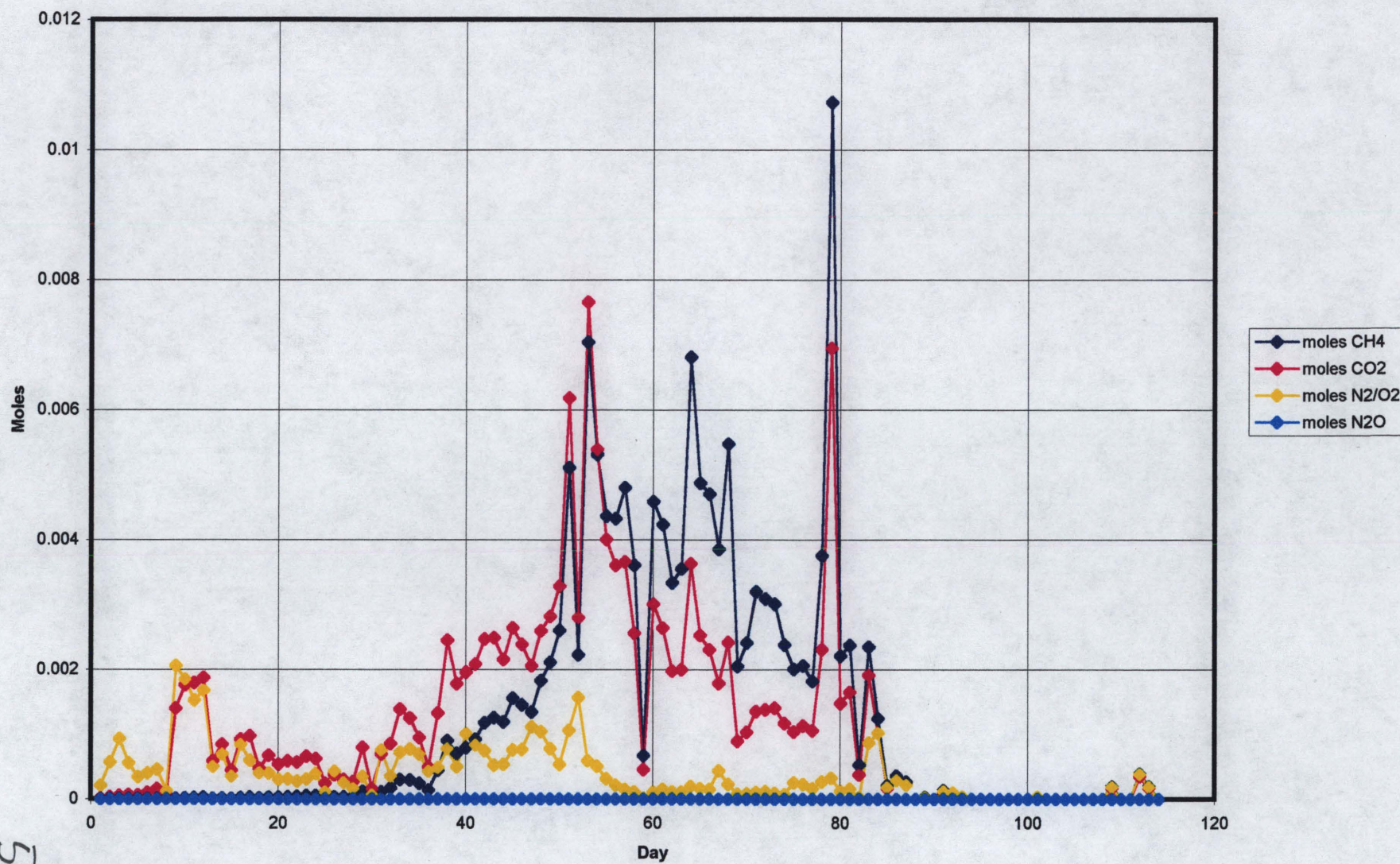
Reactor A6  
Moles Gas Produced Daily



25



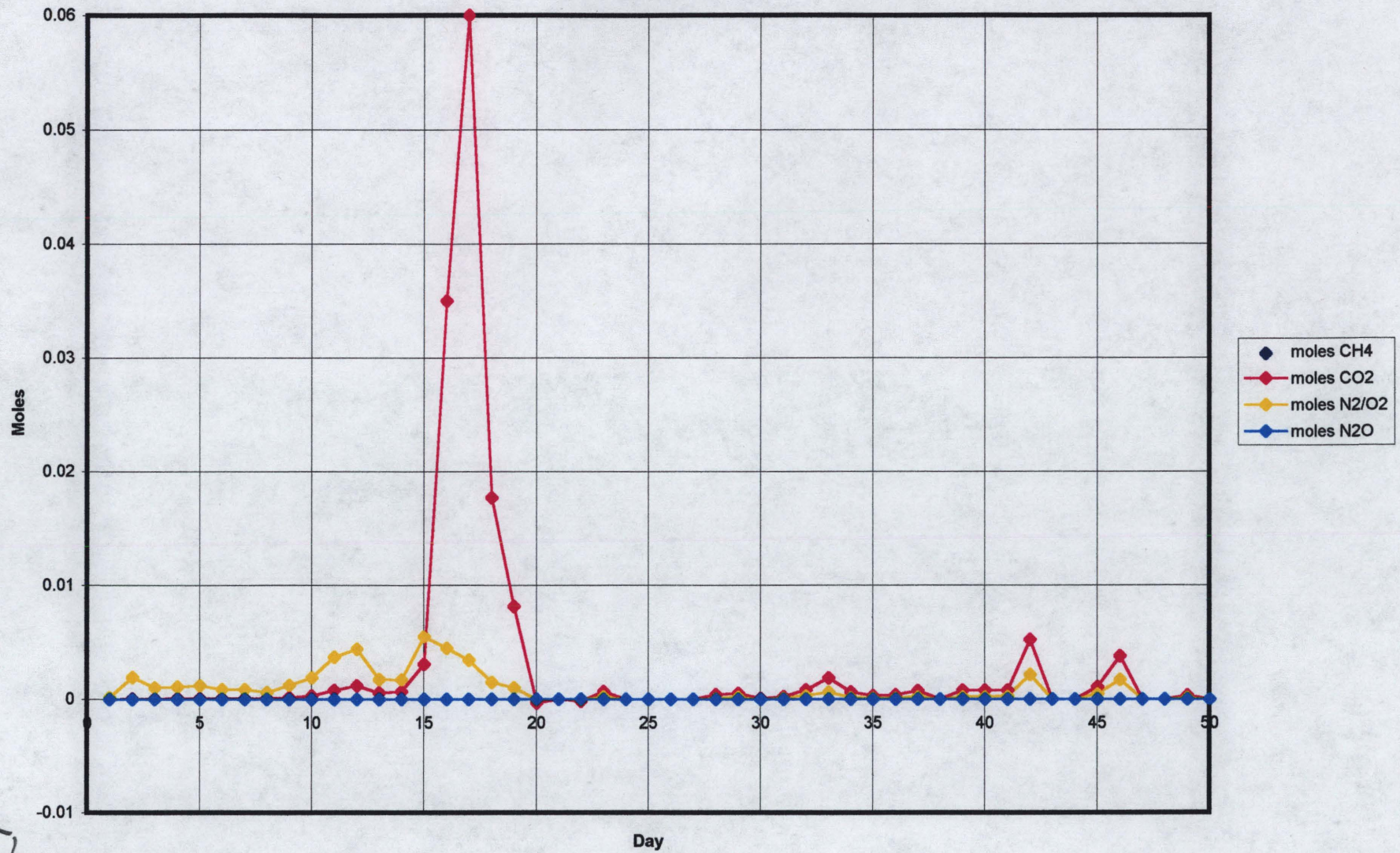
Reactor A7  
Moles Gas Produced Daily



53



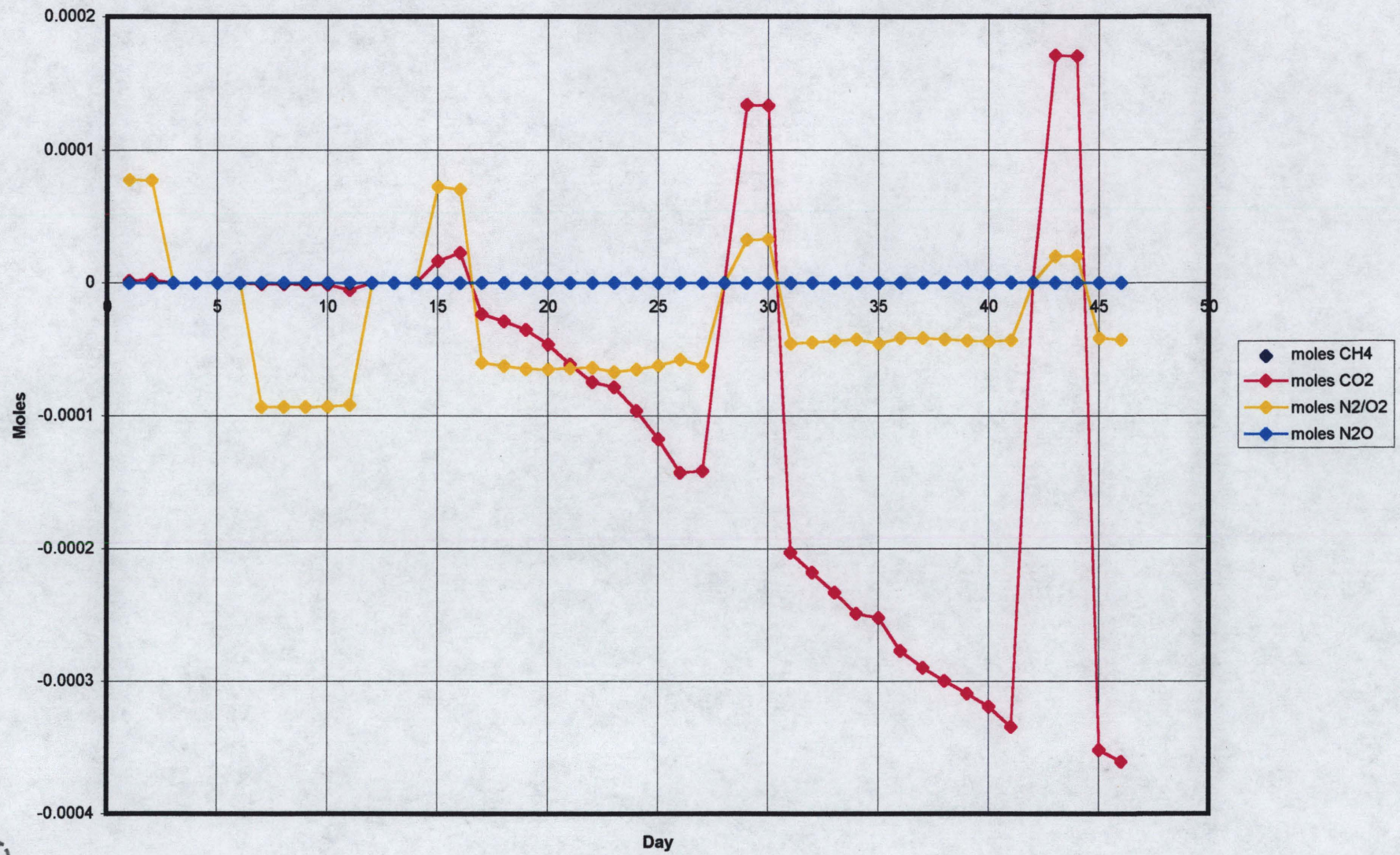
Reactor A8  
Moles Gas Produced Daily



59



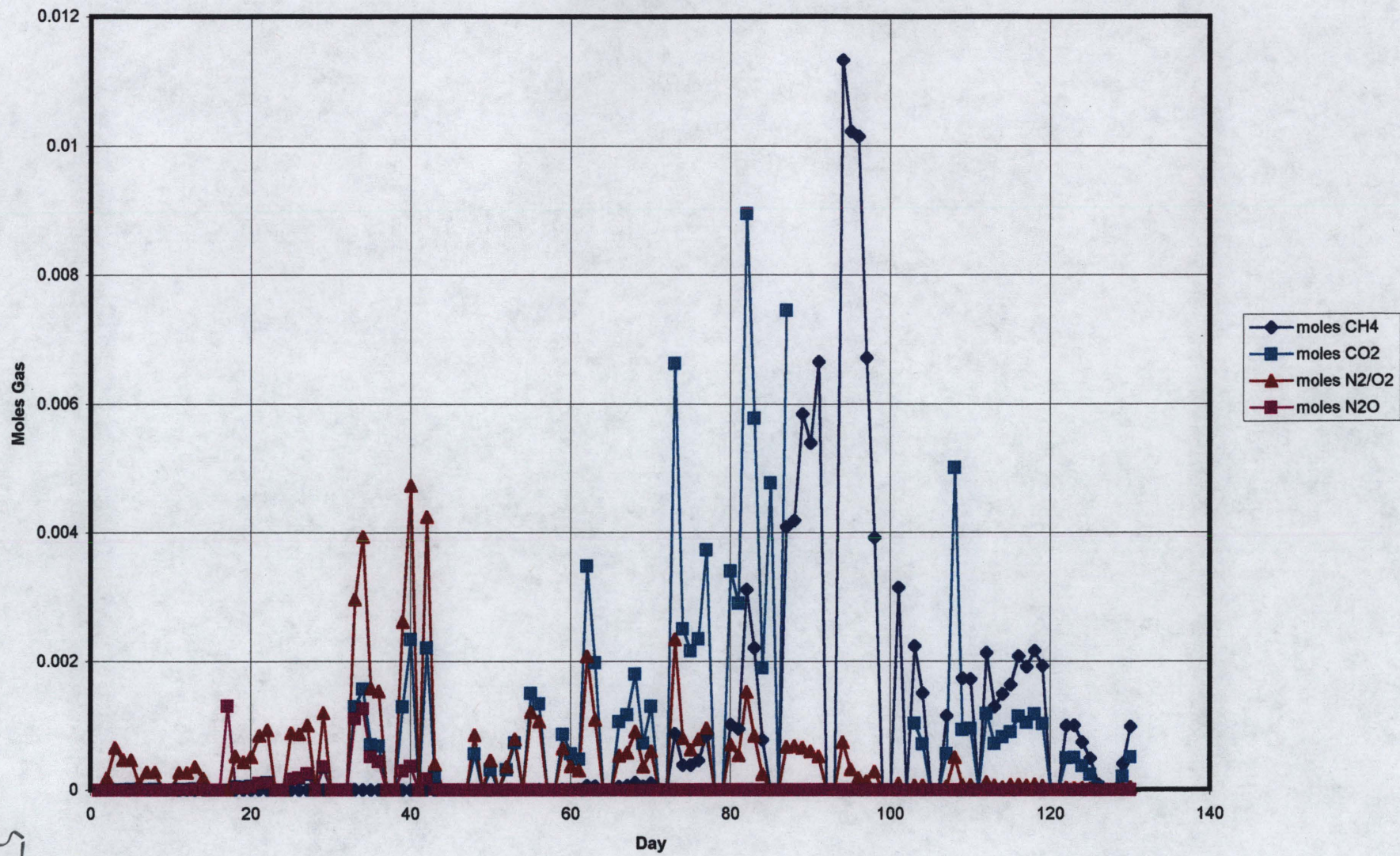
Reactor A10  
Moles Gas Produced Daily



55



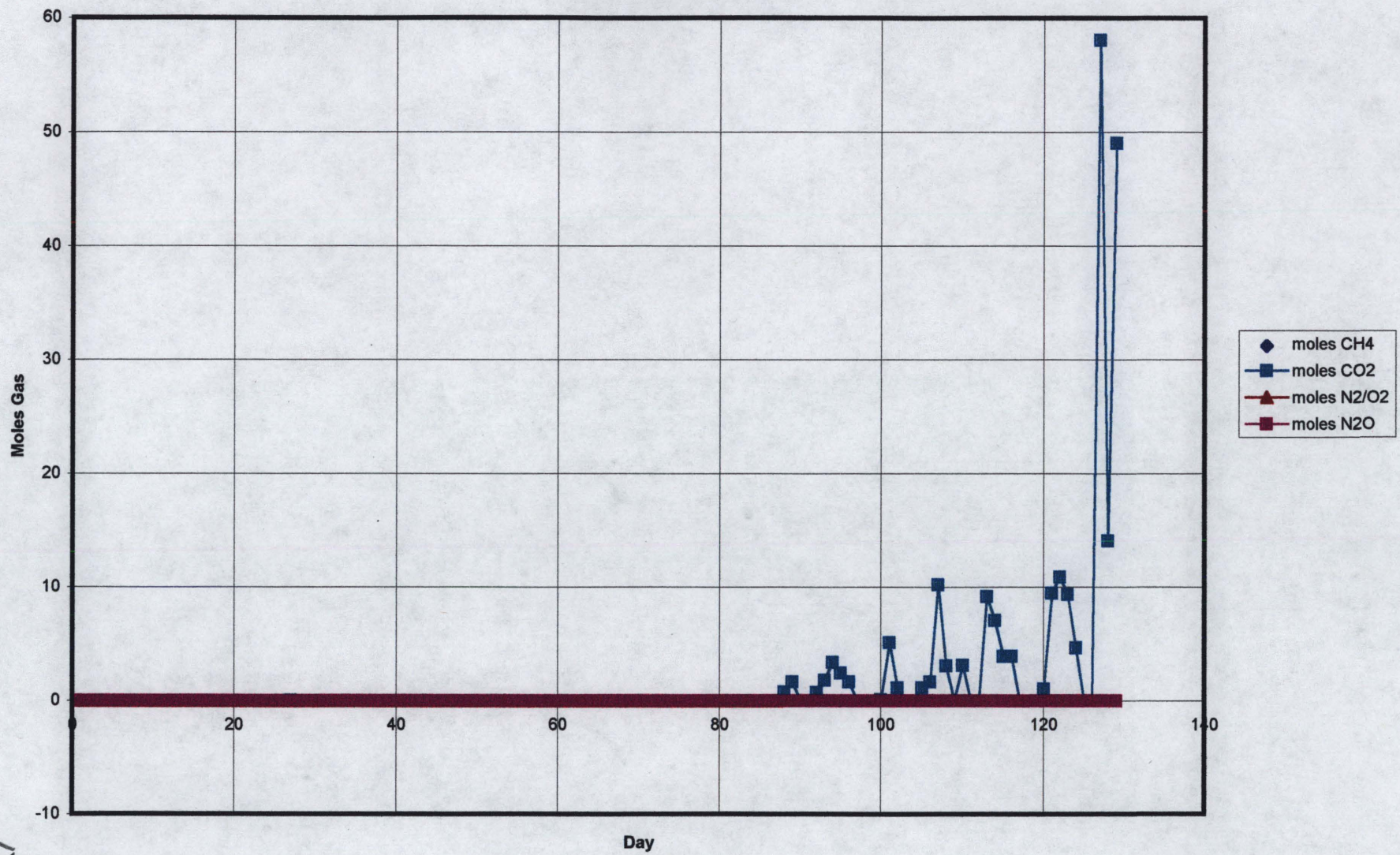
Reactor A11  
Moles Gas Produced Daily



576



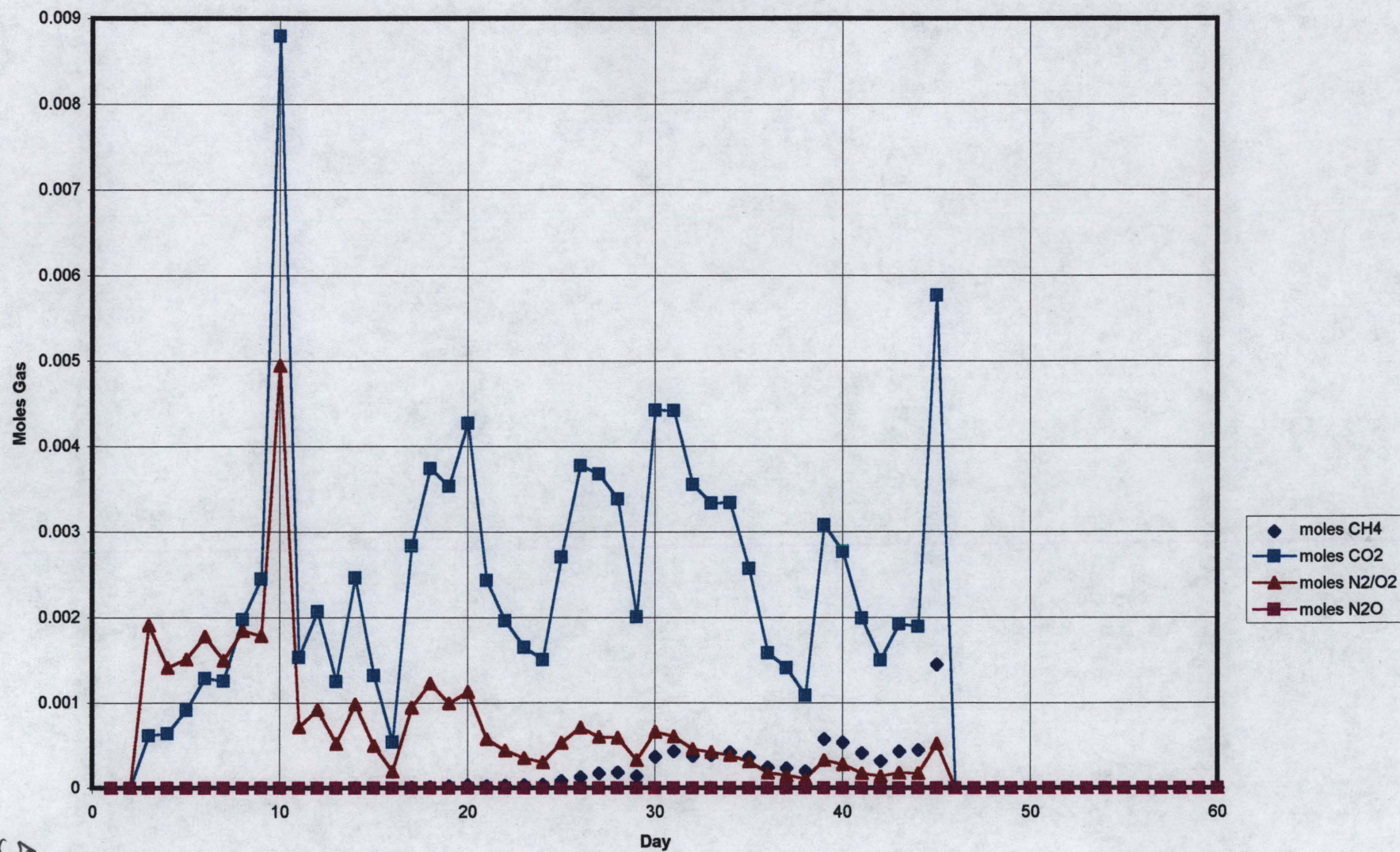
Reactor A12  
Moles Gas Produced Daily



59



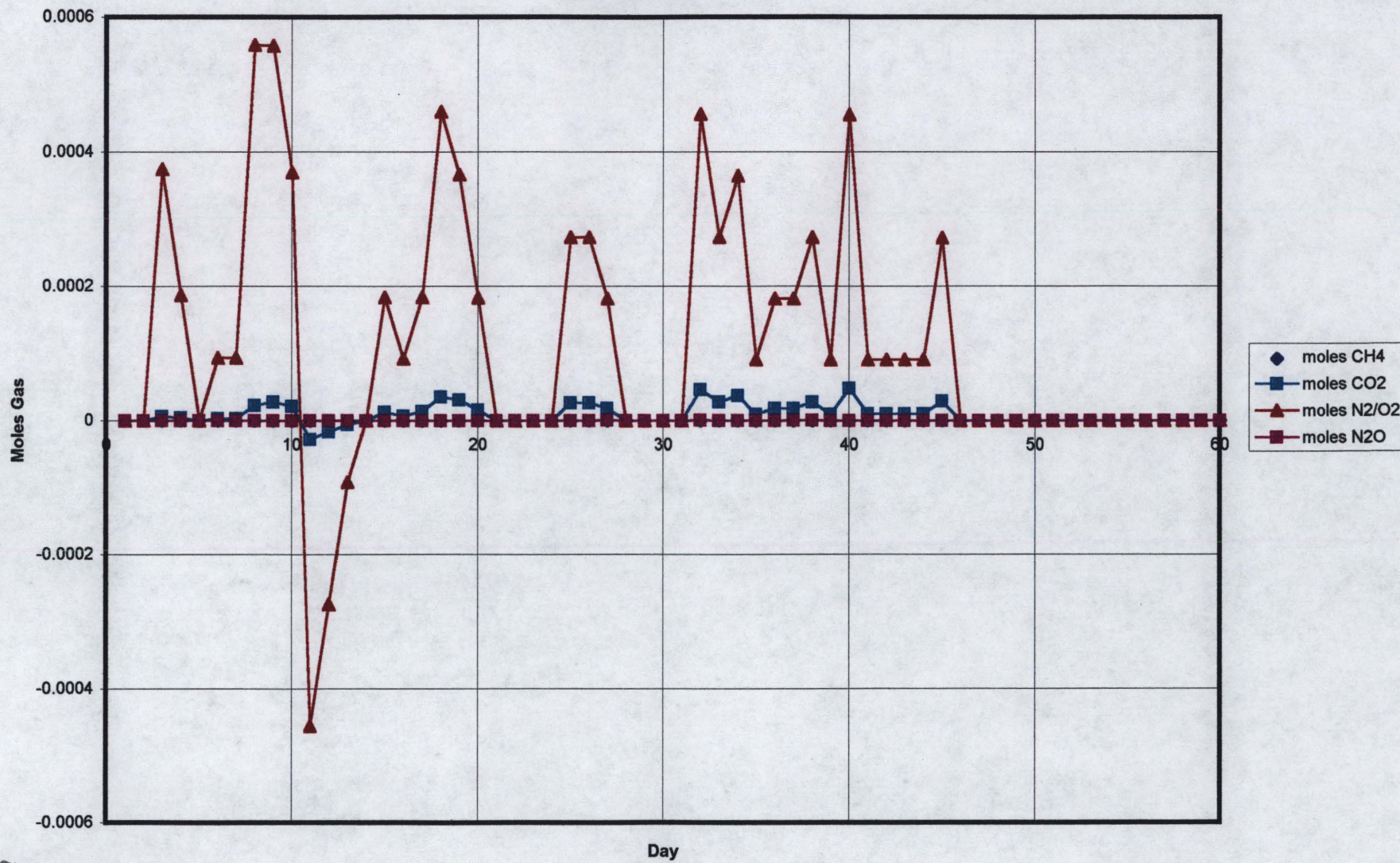
Reactor A13  
Moles Gas Produced Daily



58



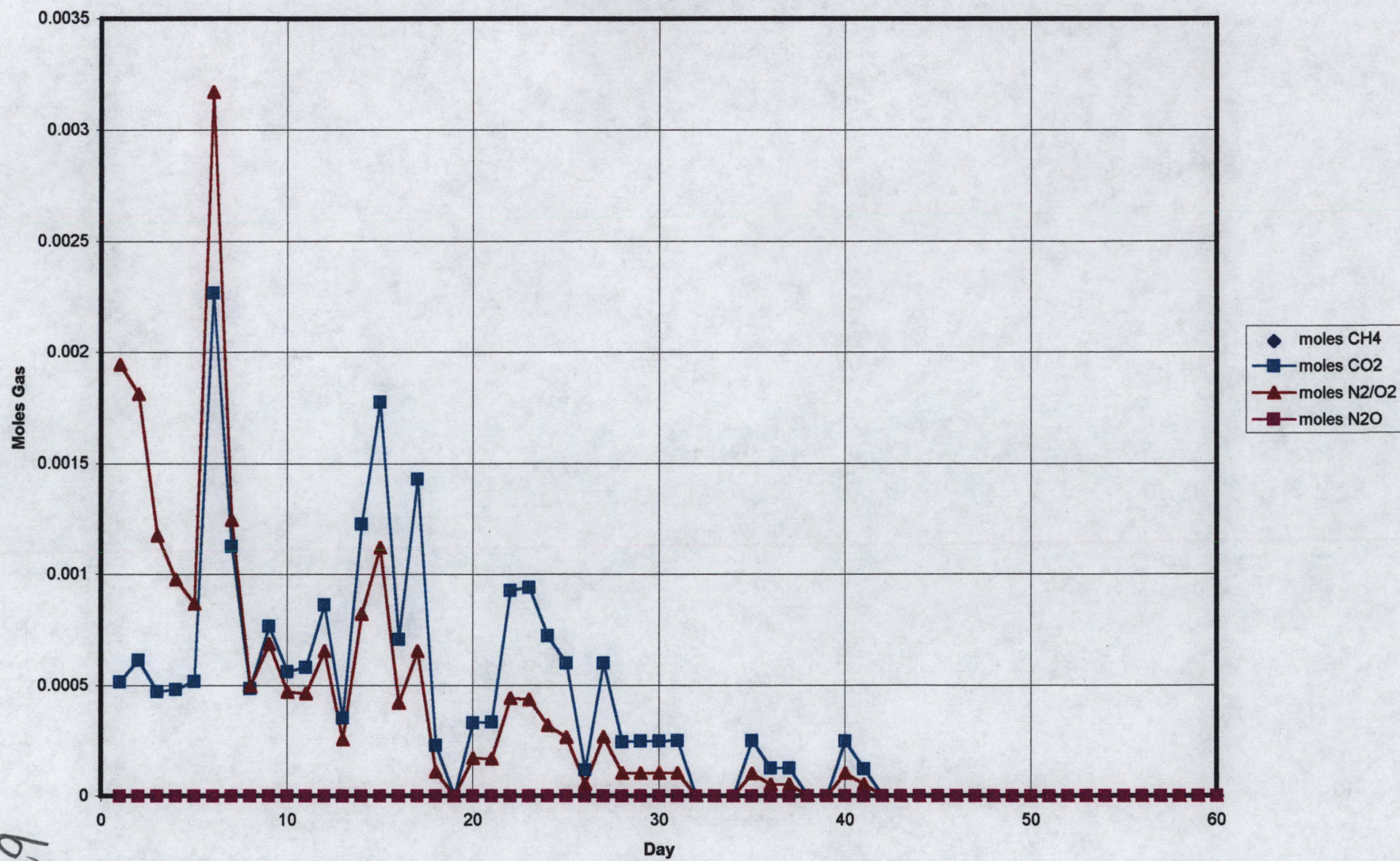
Reactor A14  
Moles Gas Produced Daily



59



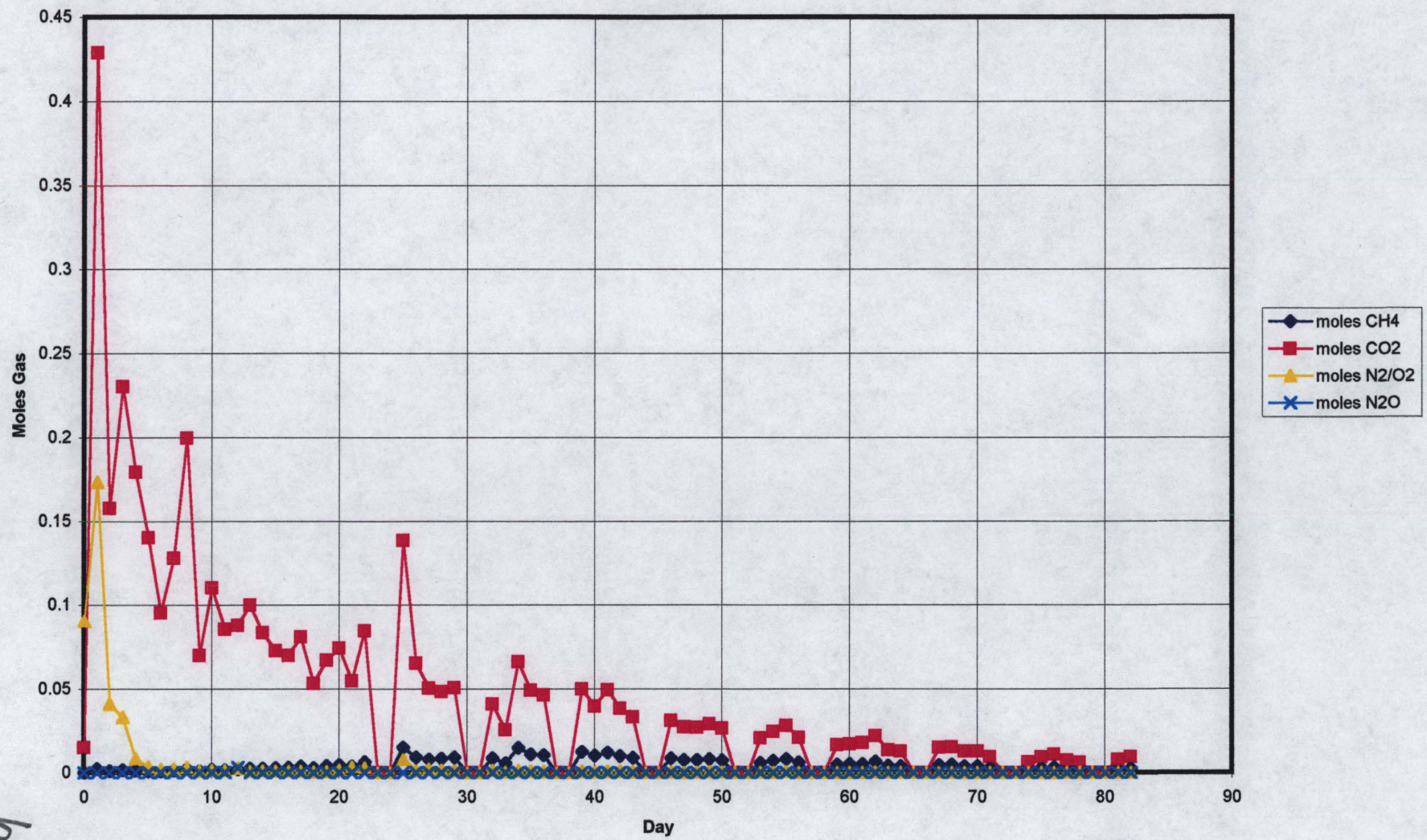
Reactor A15  
Moles Gas Produced Daily



09



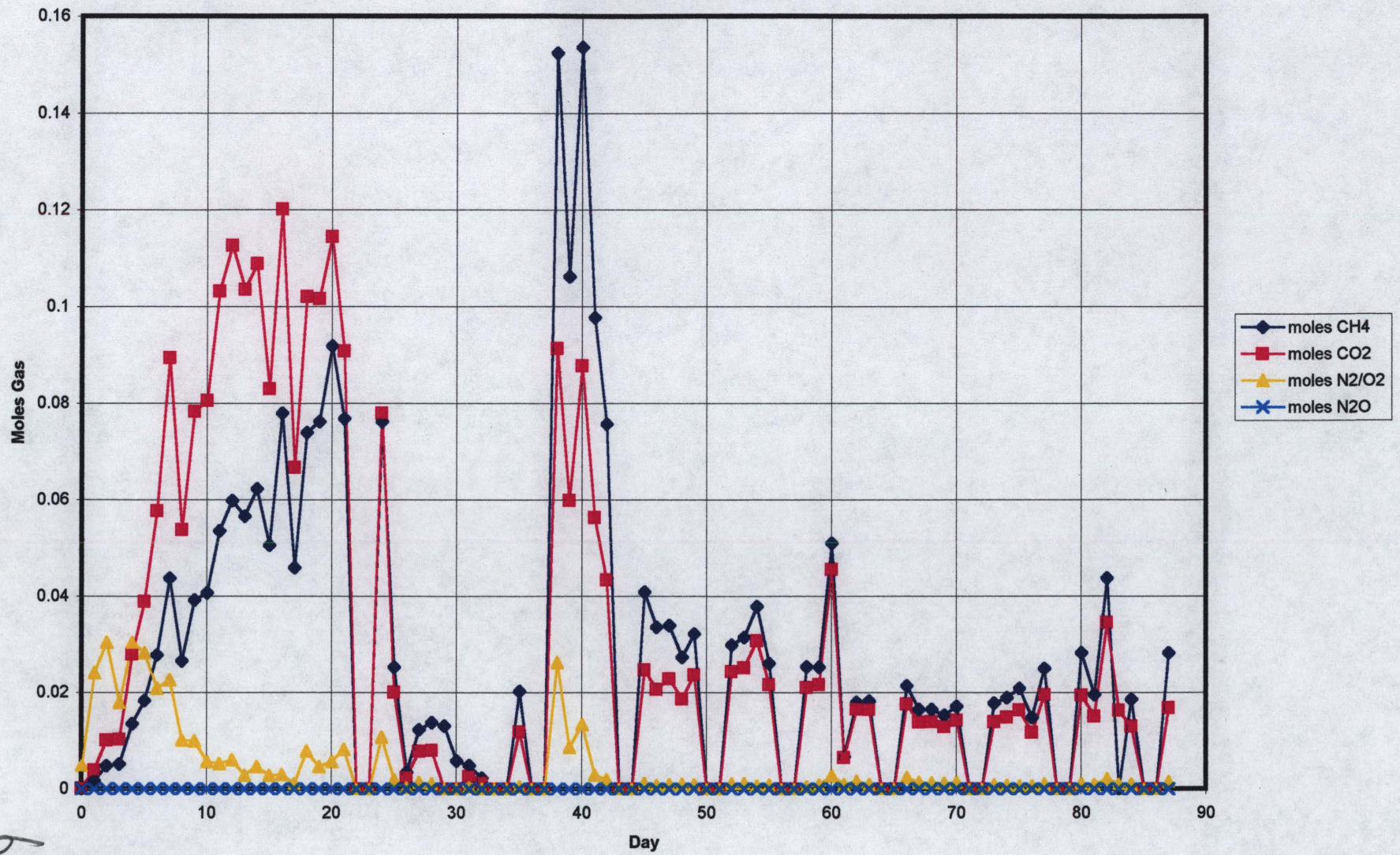
Reactor A16  
Moles Gas Produced Daily



61



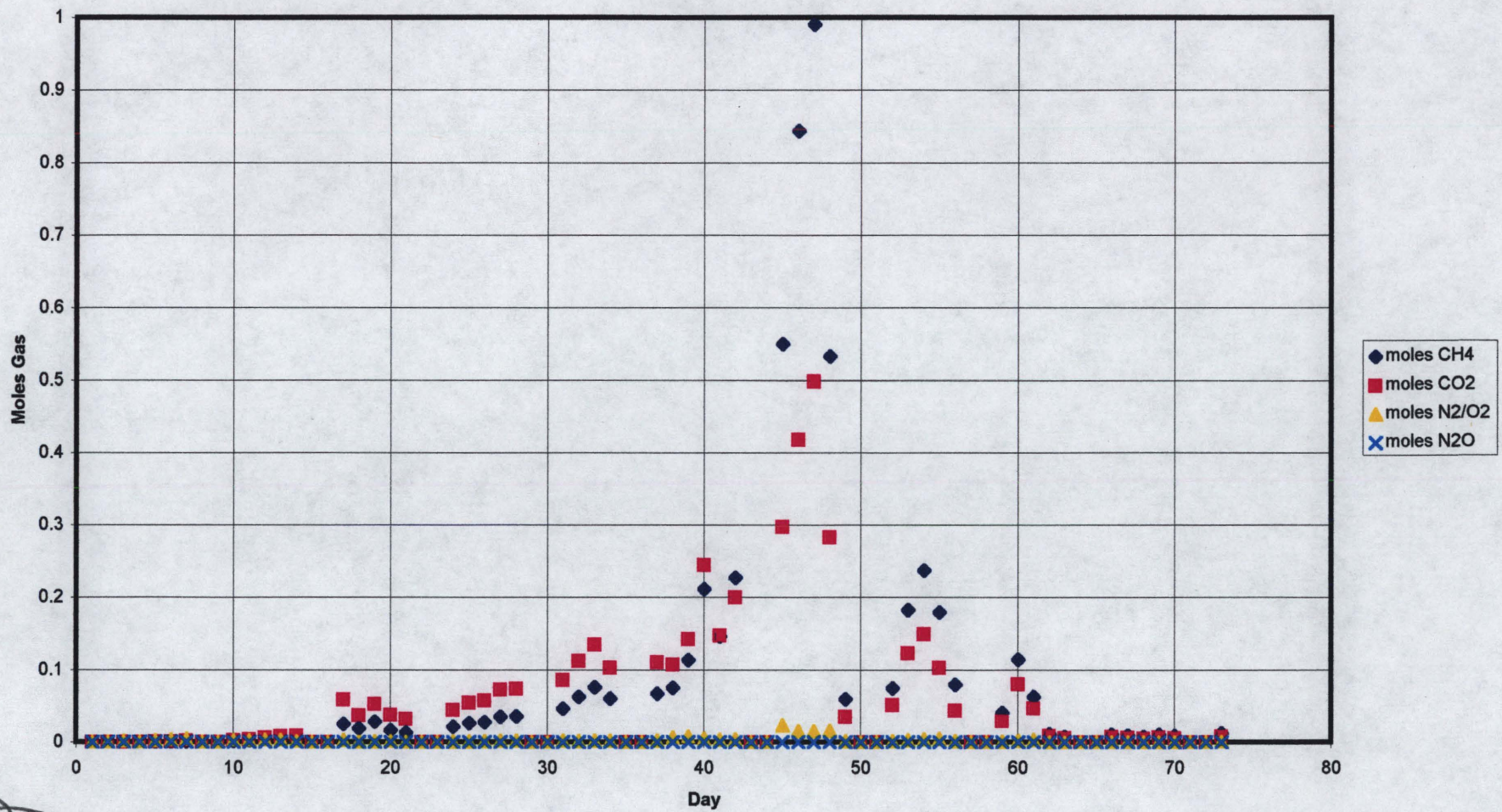
Reactor A18  
Moles Gas Produced Daily



62



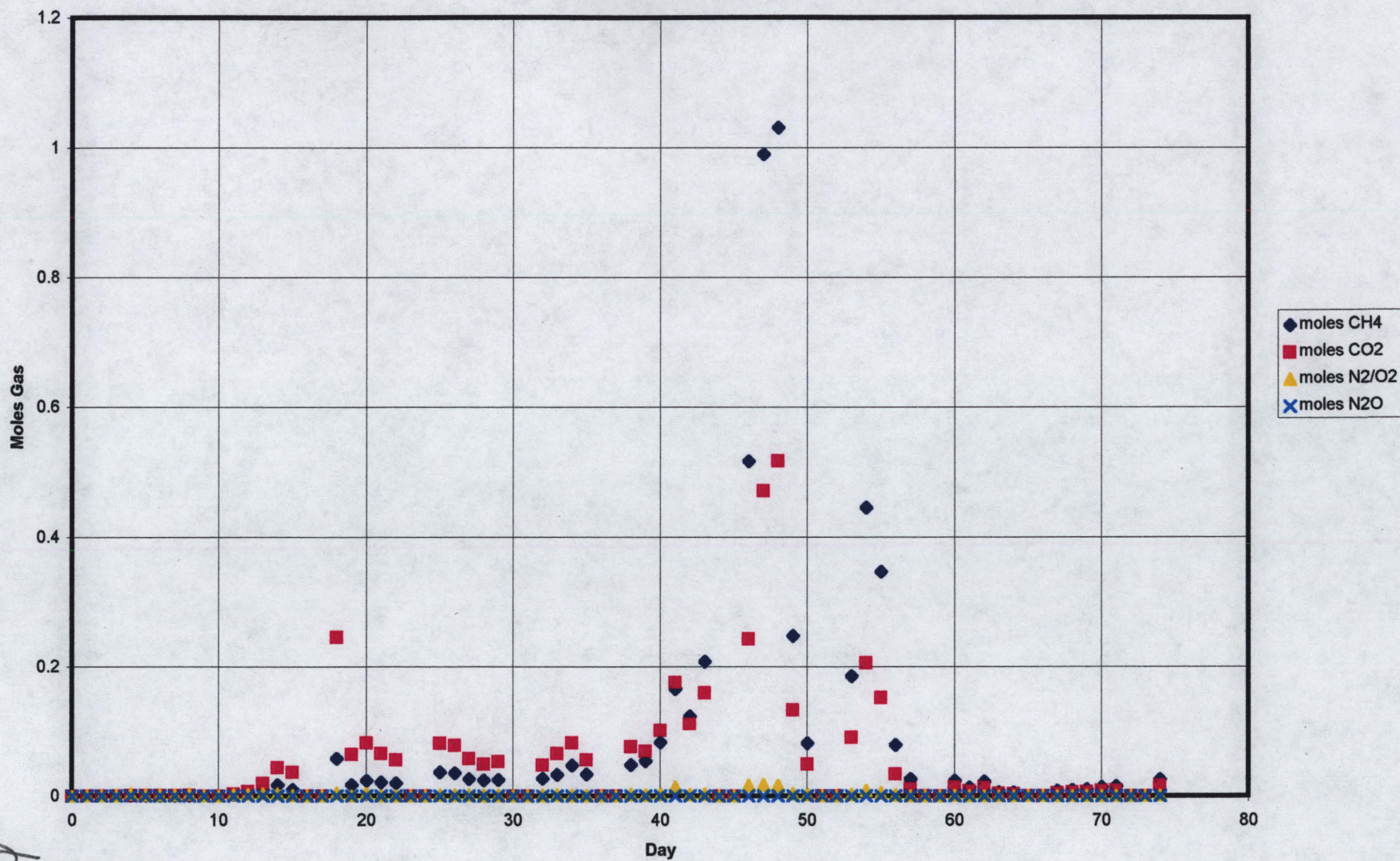
Reactor A19  
Moles Gas Produced Daily



63



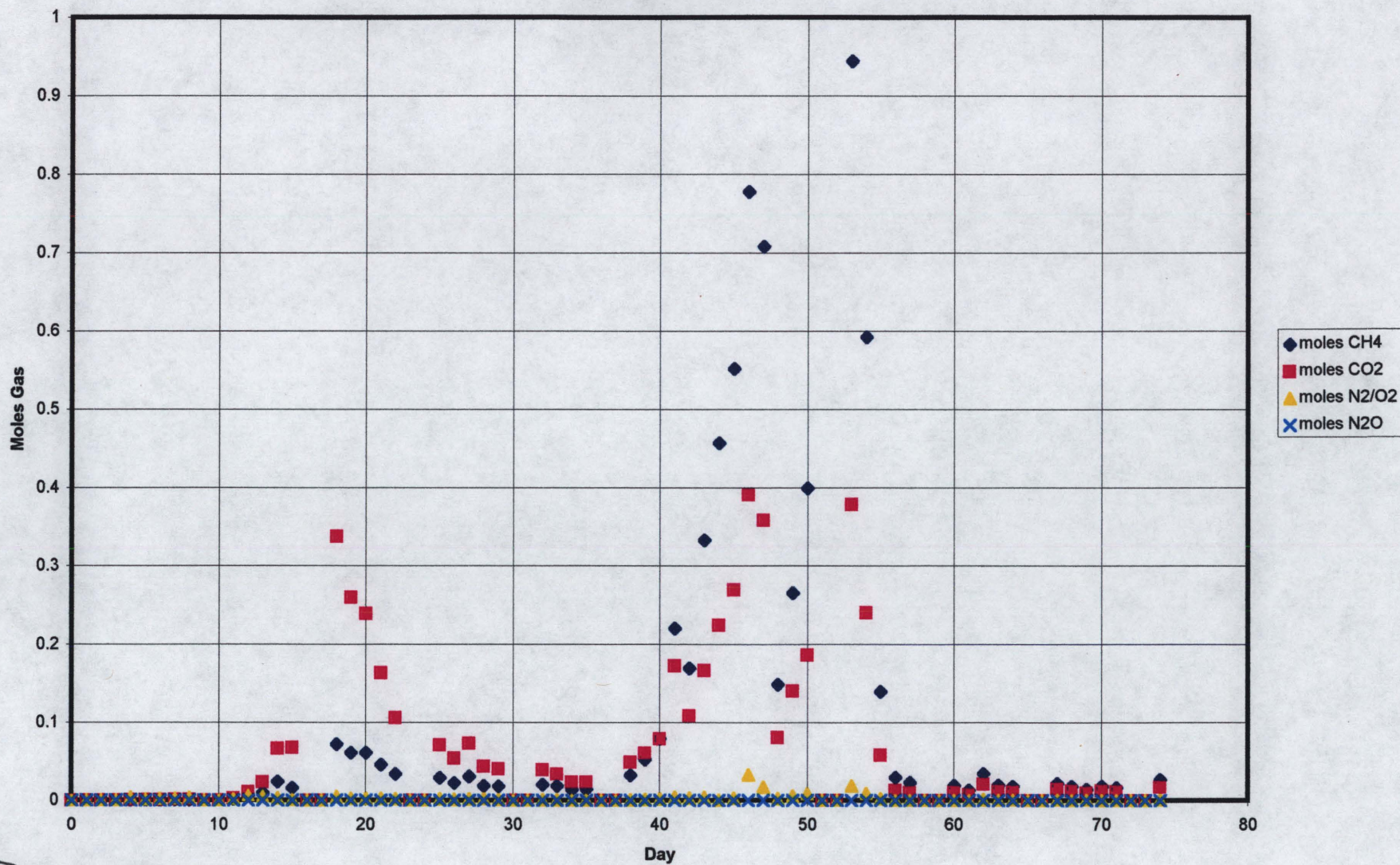
Reactor 20  
Moles Gas Produced Daily



64



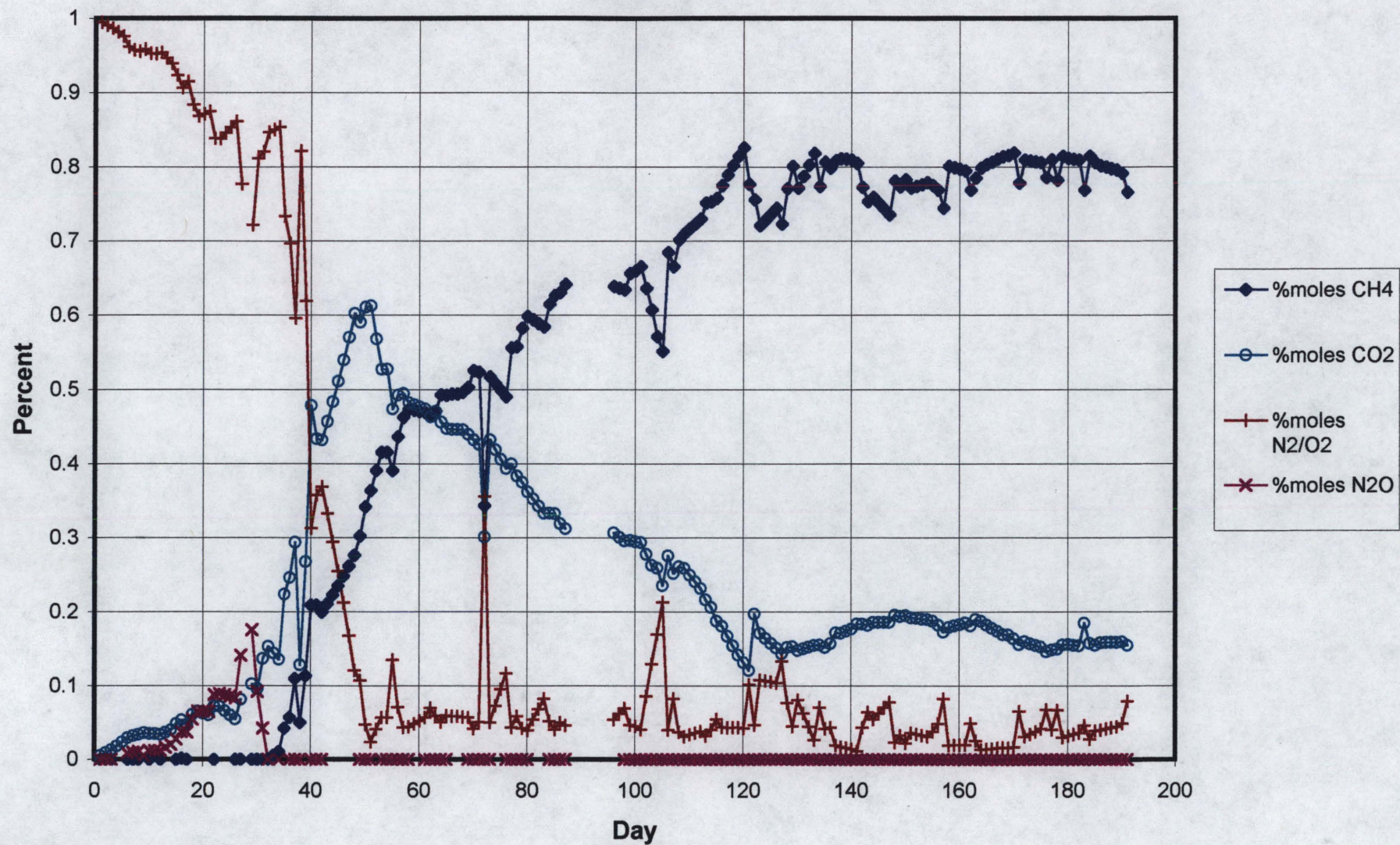
Reactor 21  
Moles Gas Produced Daily



65



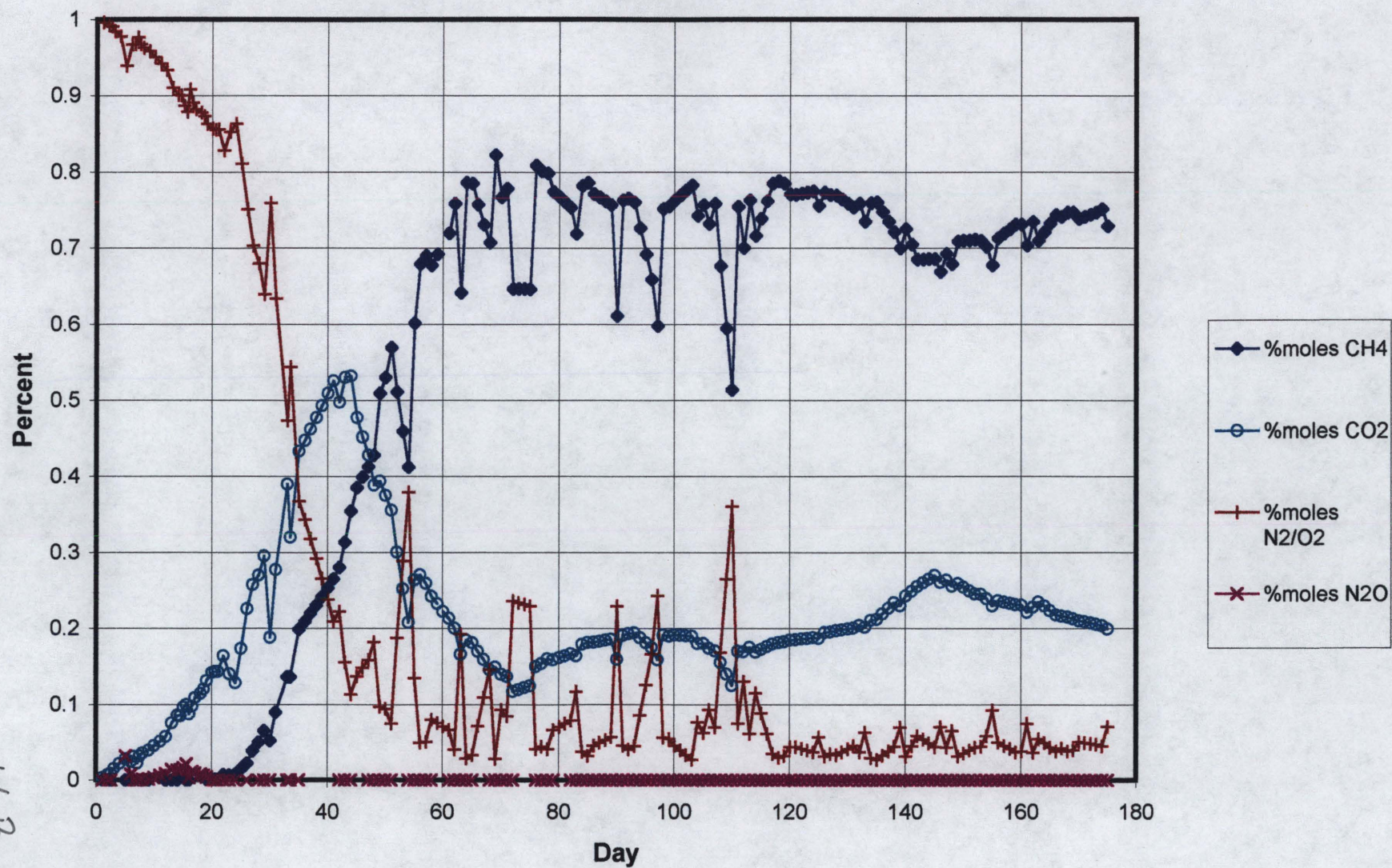
Reactor A1  
Mole Percent/100



66A



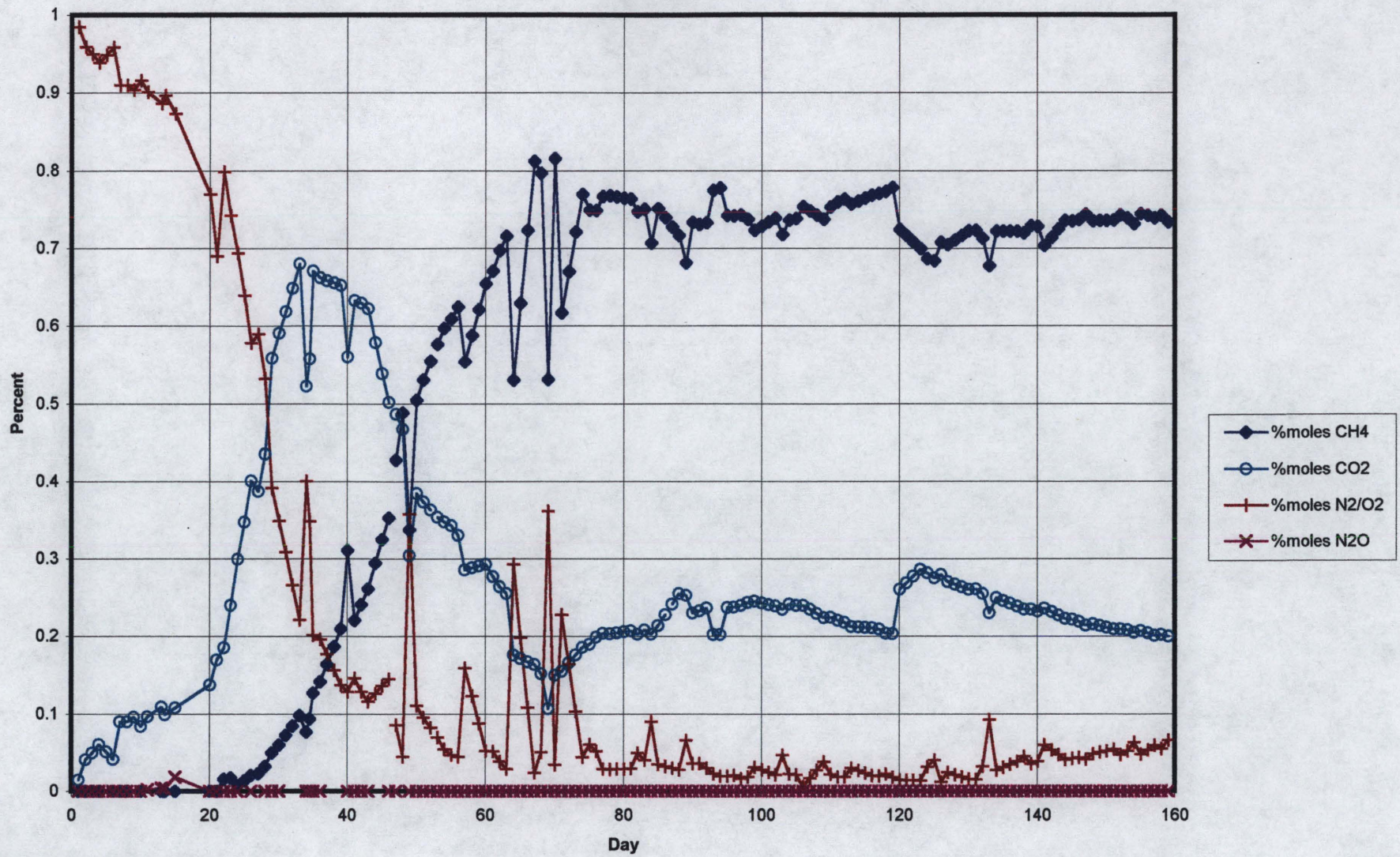
Reactor A2  
Mole Percent/100



66B



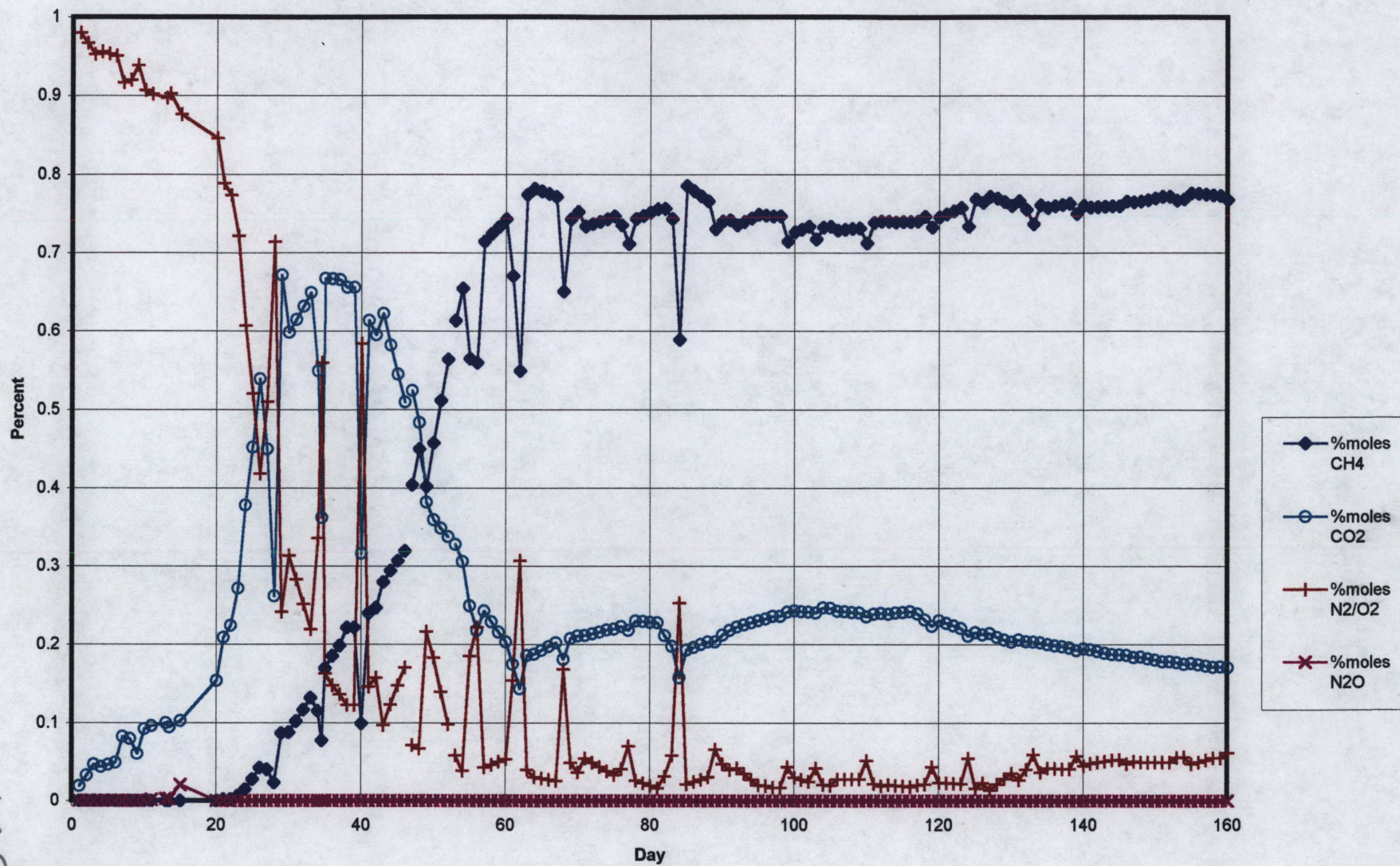
Reactor A3  
Mole Percent/100



67A

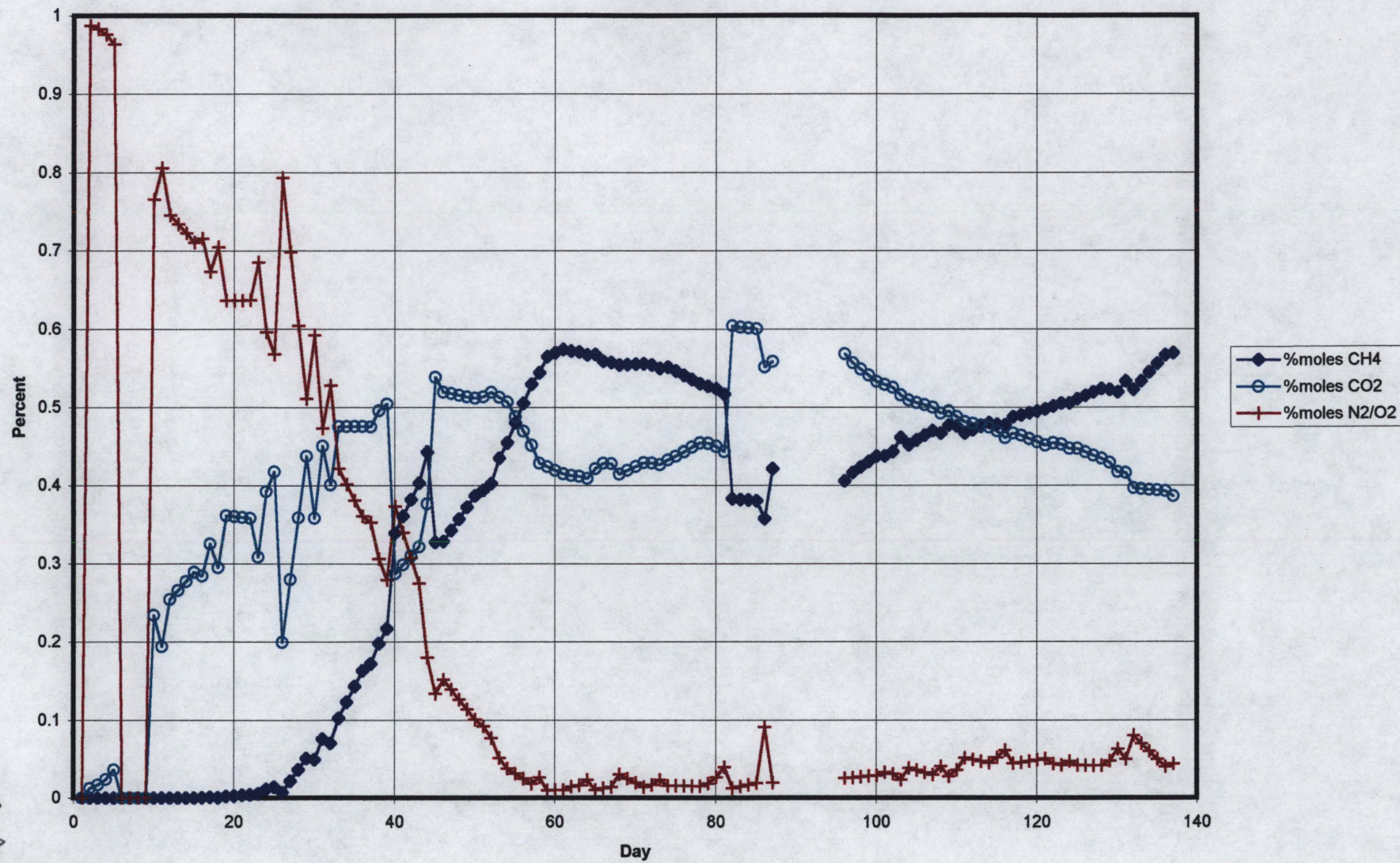


Reactor A4  
Mole Percent/100





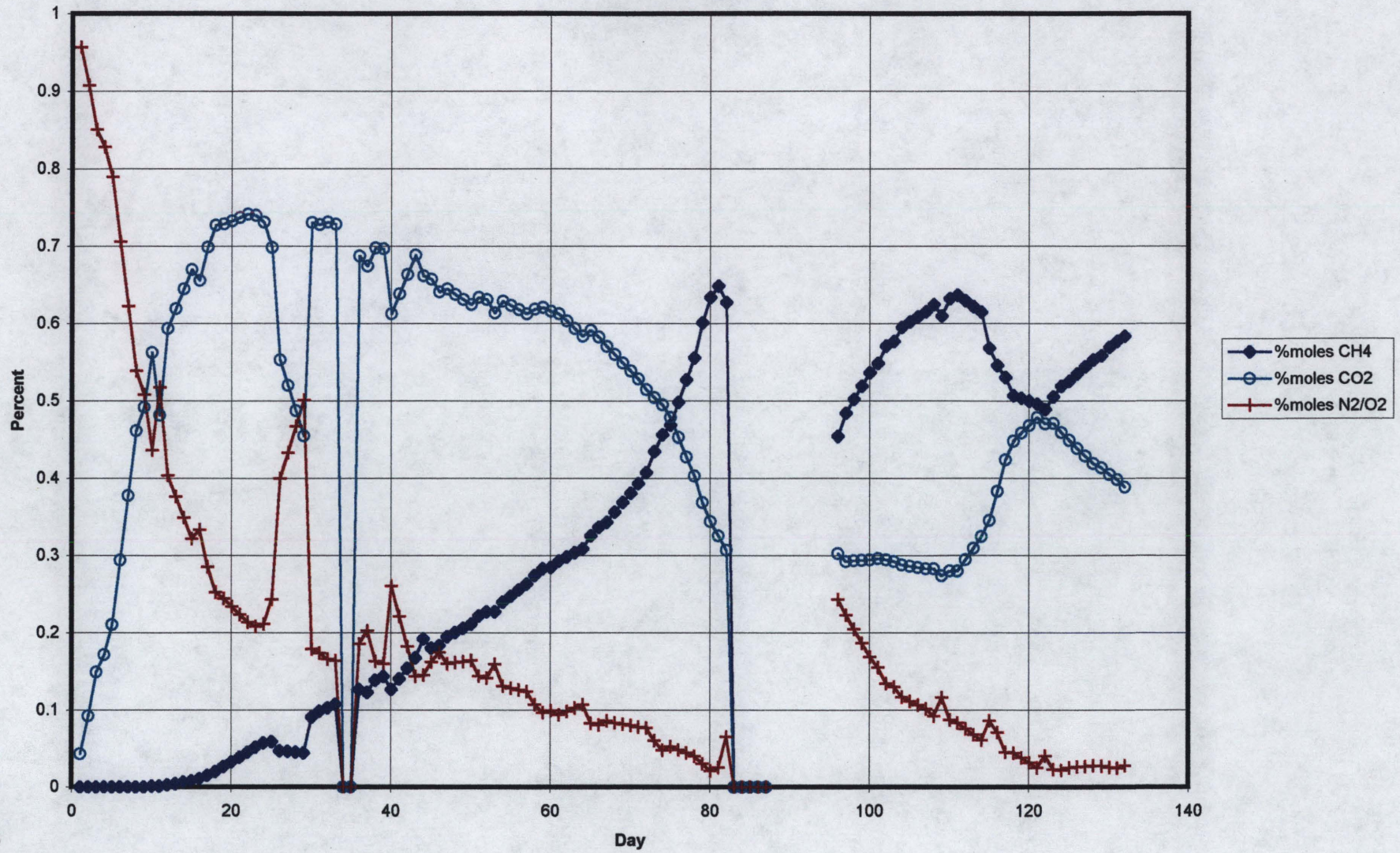
Reactor A5  
Mole Percent/100



68A



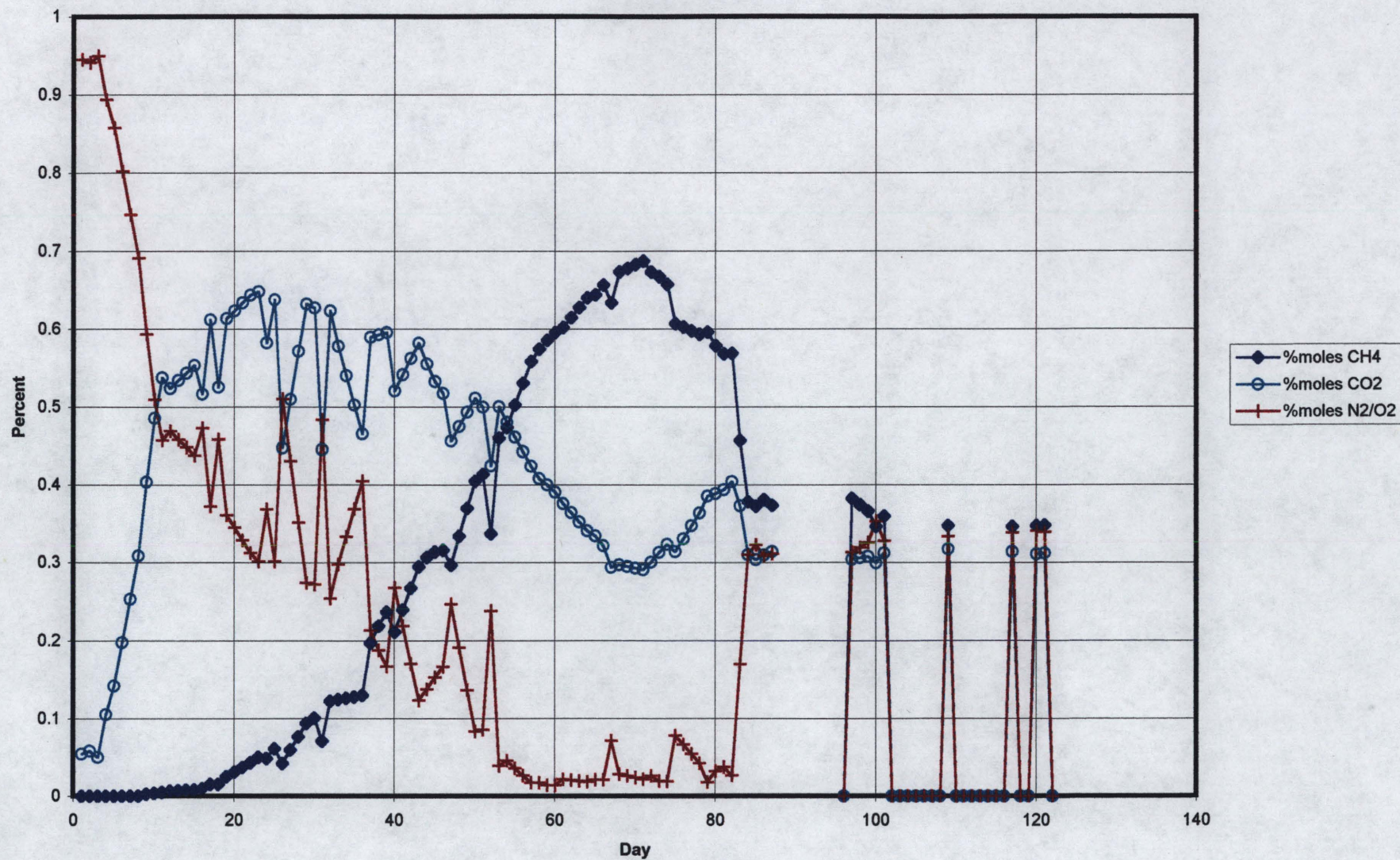
Reactor A6  
Mole Percent/100



68B



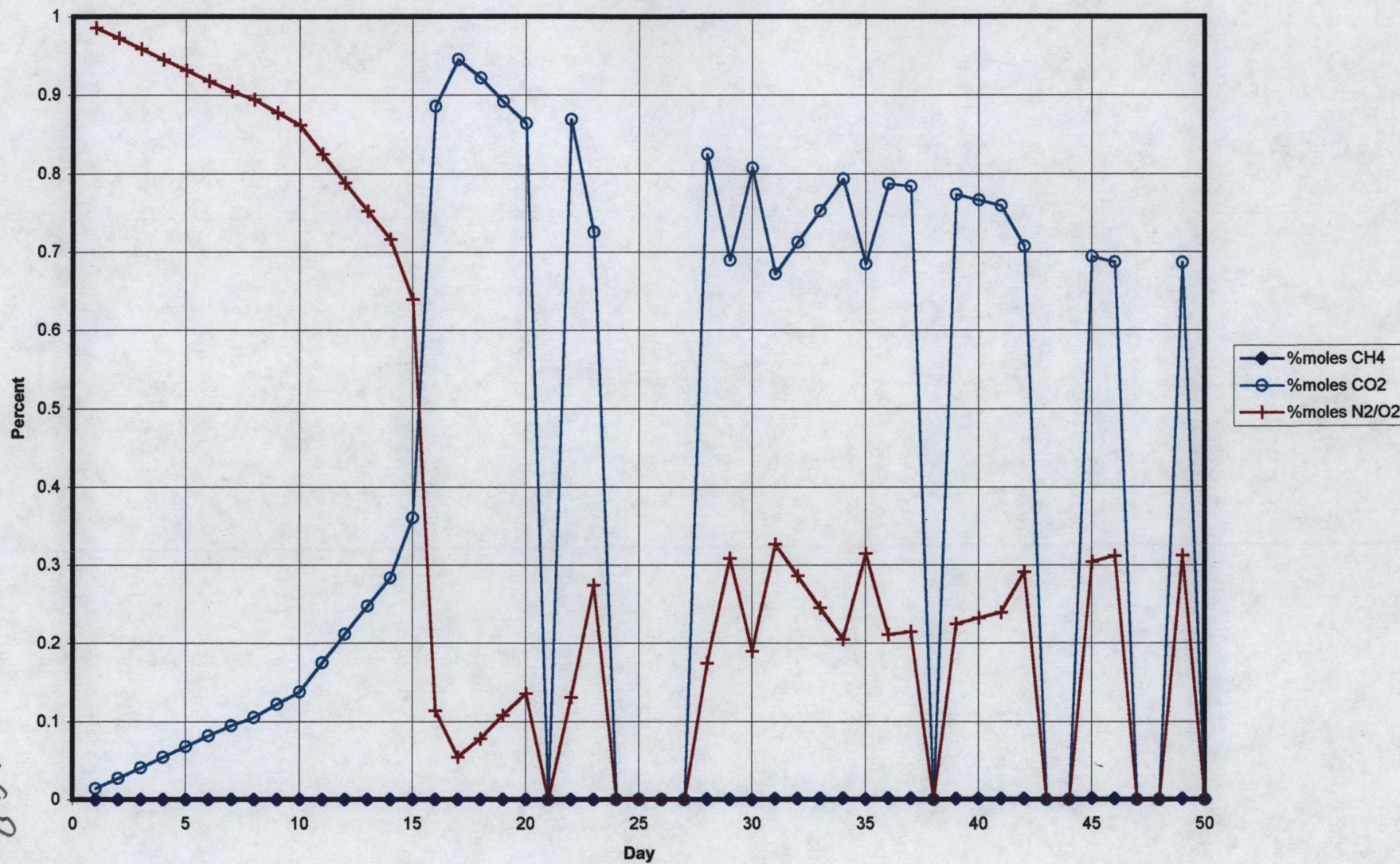
Reactor A7  
Mole Percent/100



69A



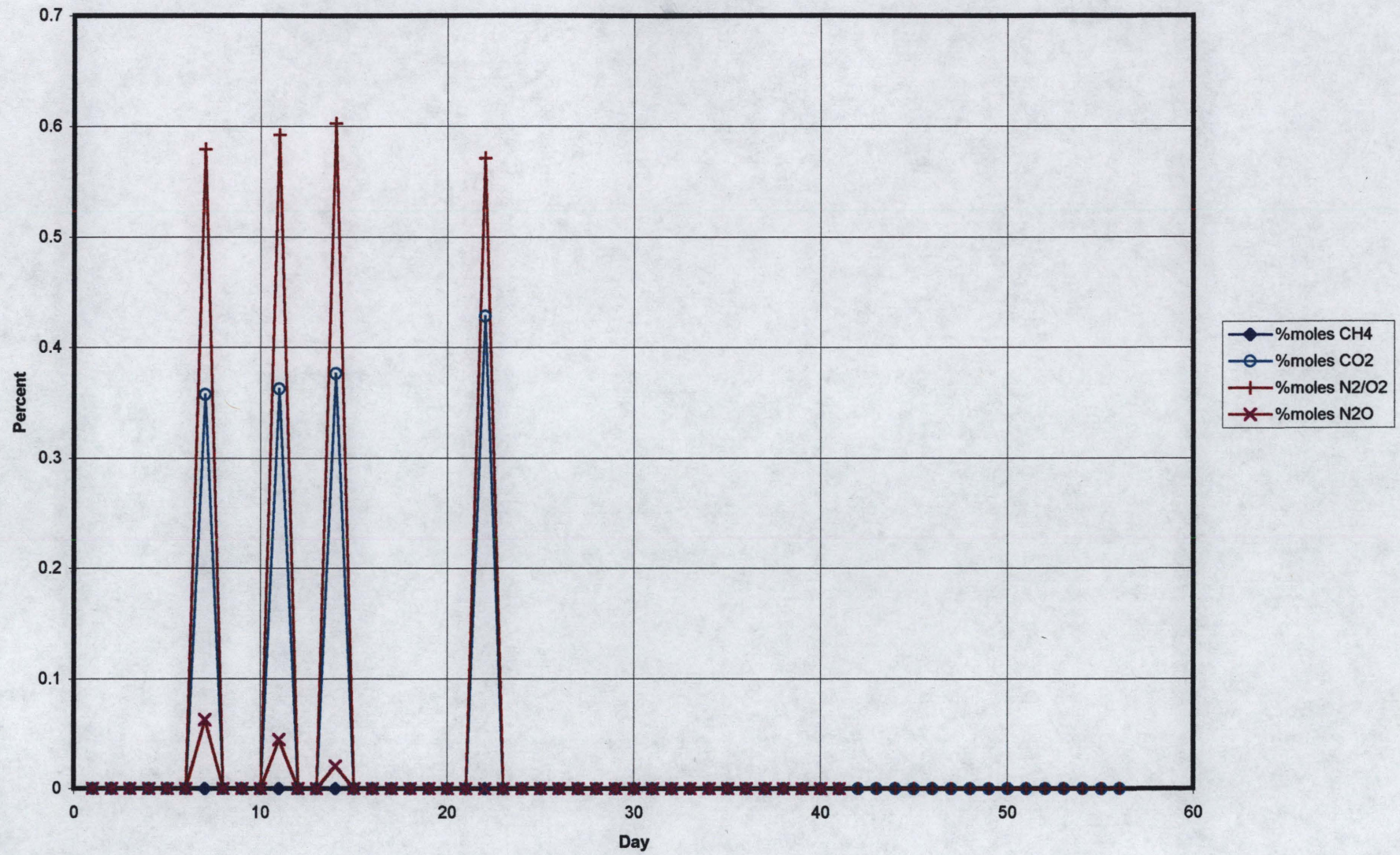
Reactor A8  
Mole Percent/100



69B



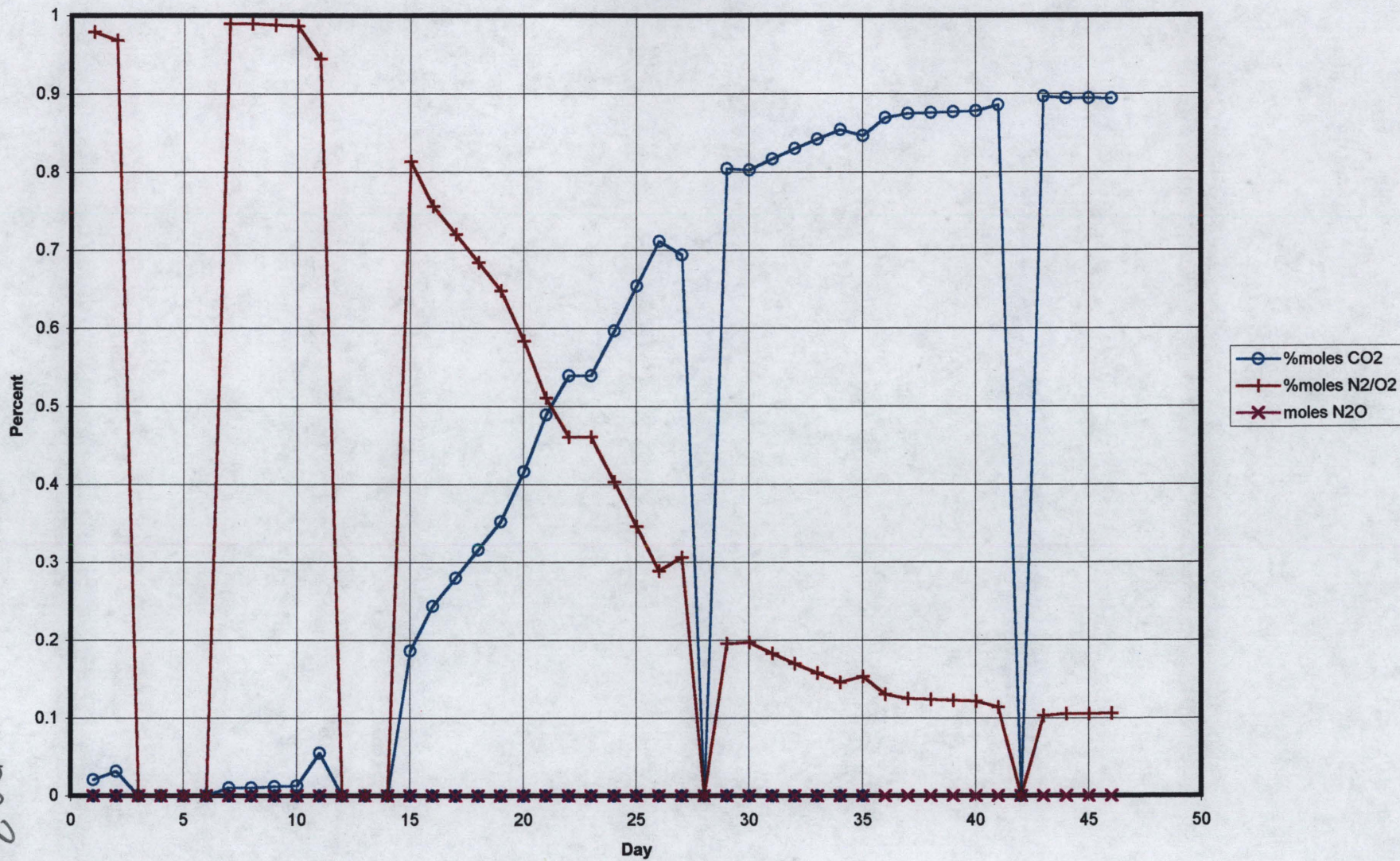
Reactor 9  
Mole Percent/100



176A



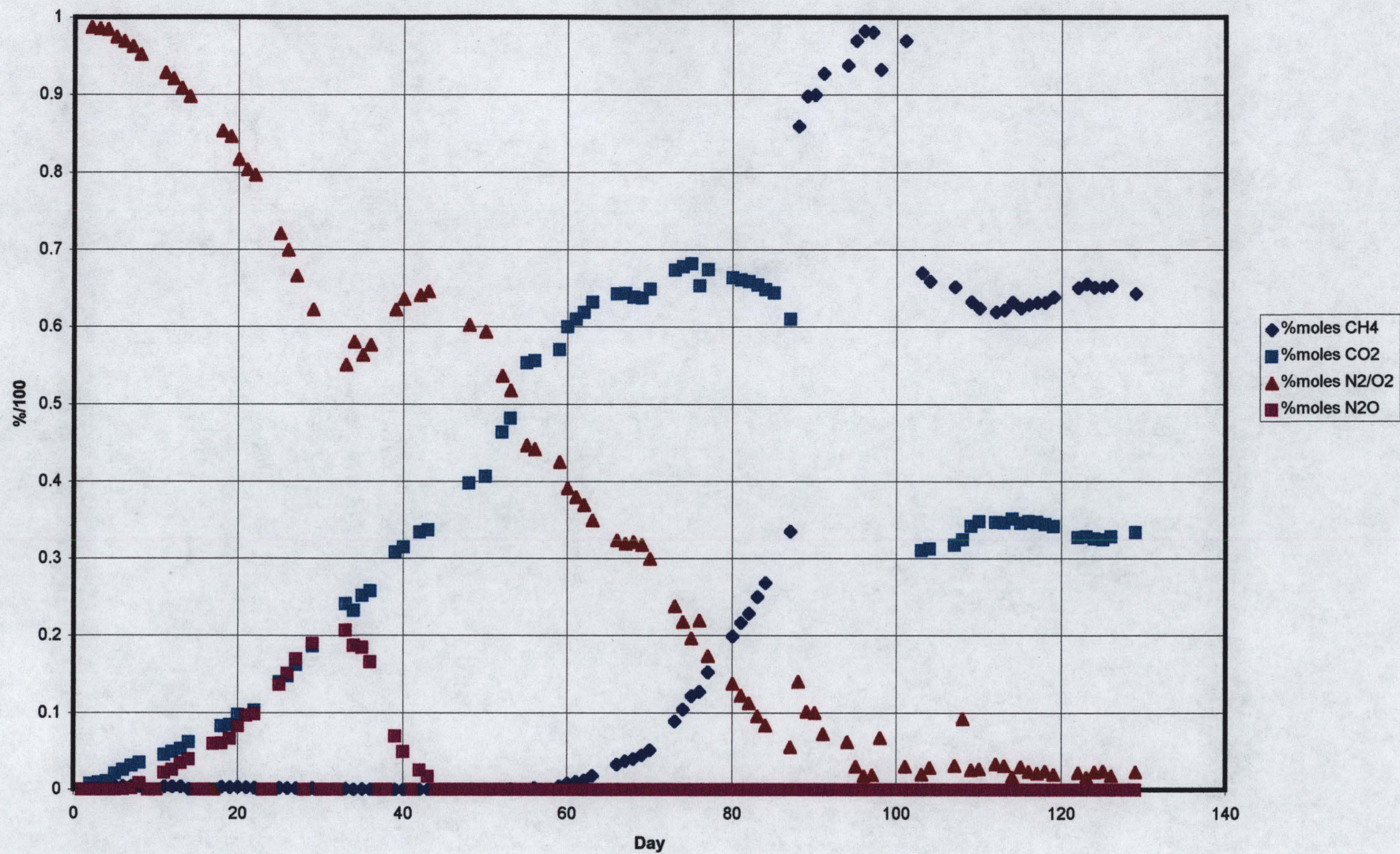
Reactor A10  
Mole Percent/100



706

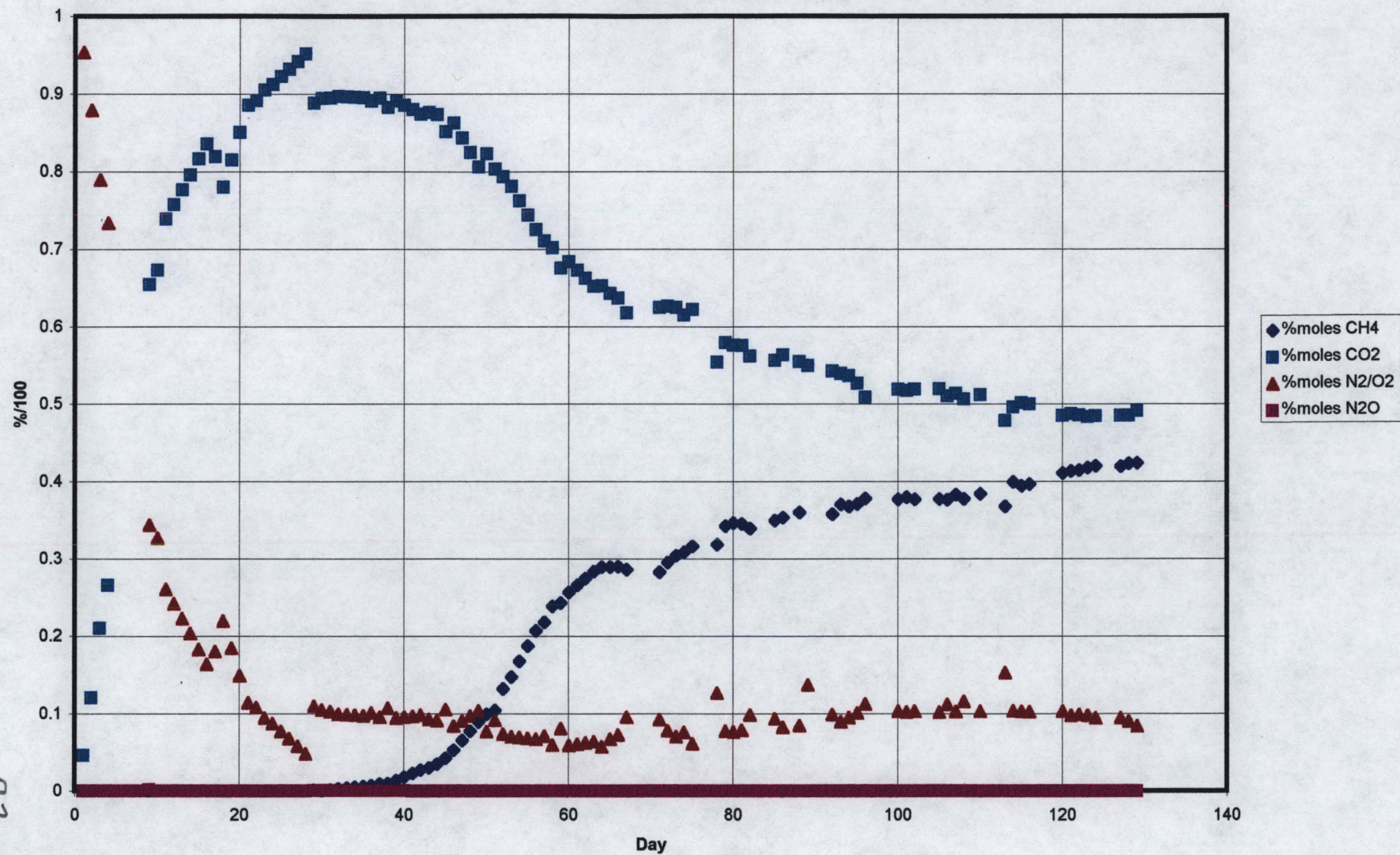


Reactor A11  
Mole Percent/100





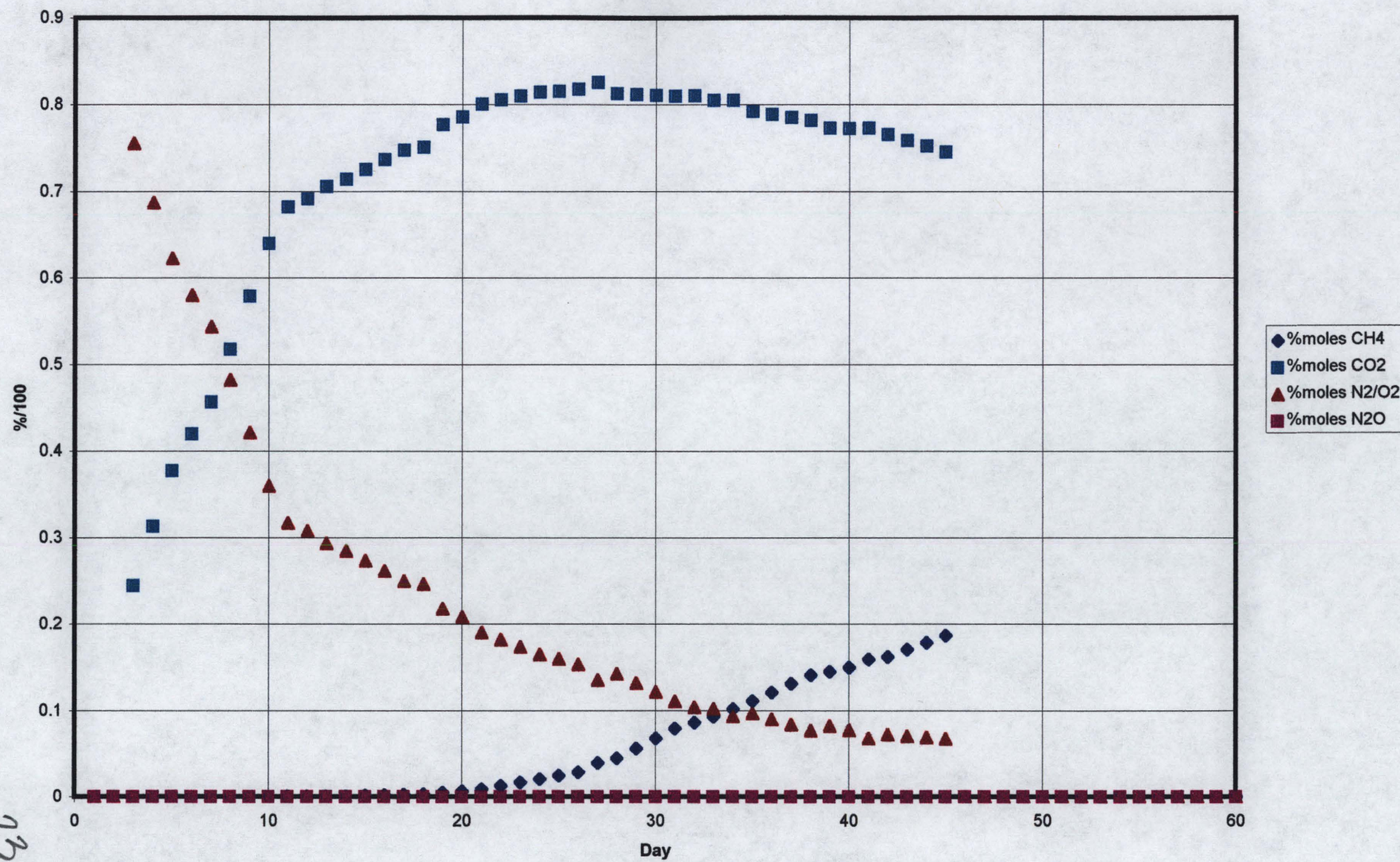
Reactor A12  
Mole Percent/100



26

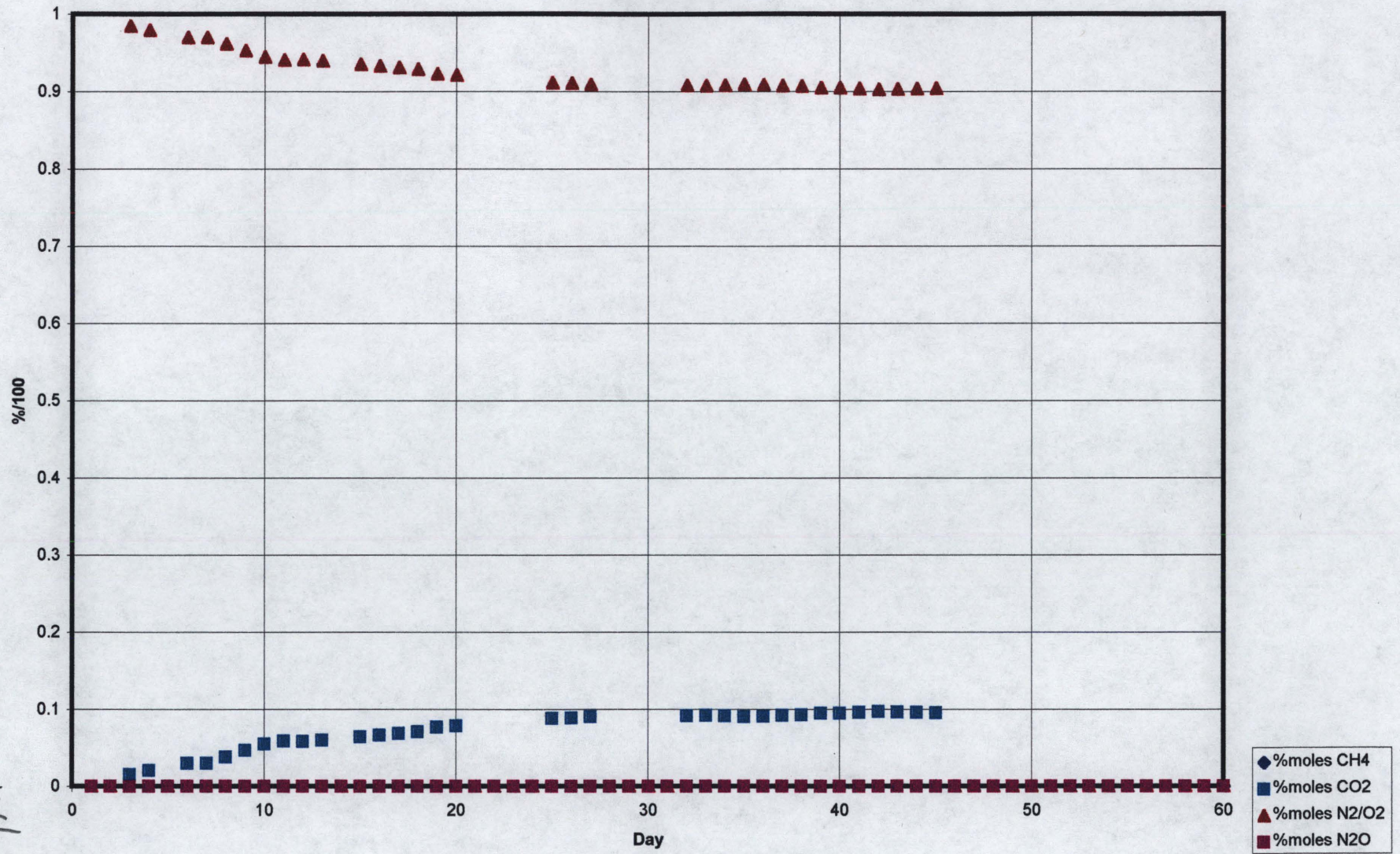


Reactor A13  
Mole Percent/100



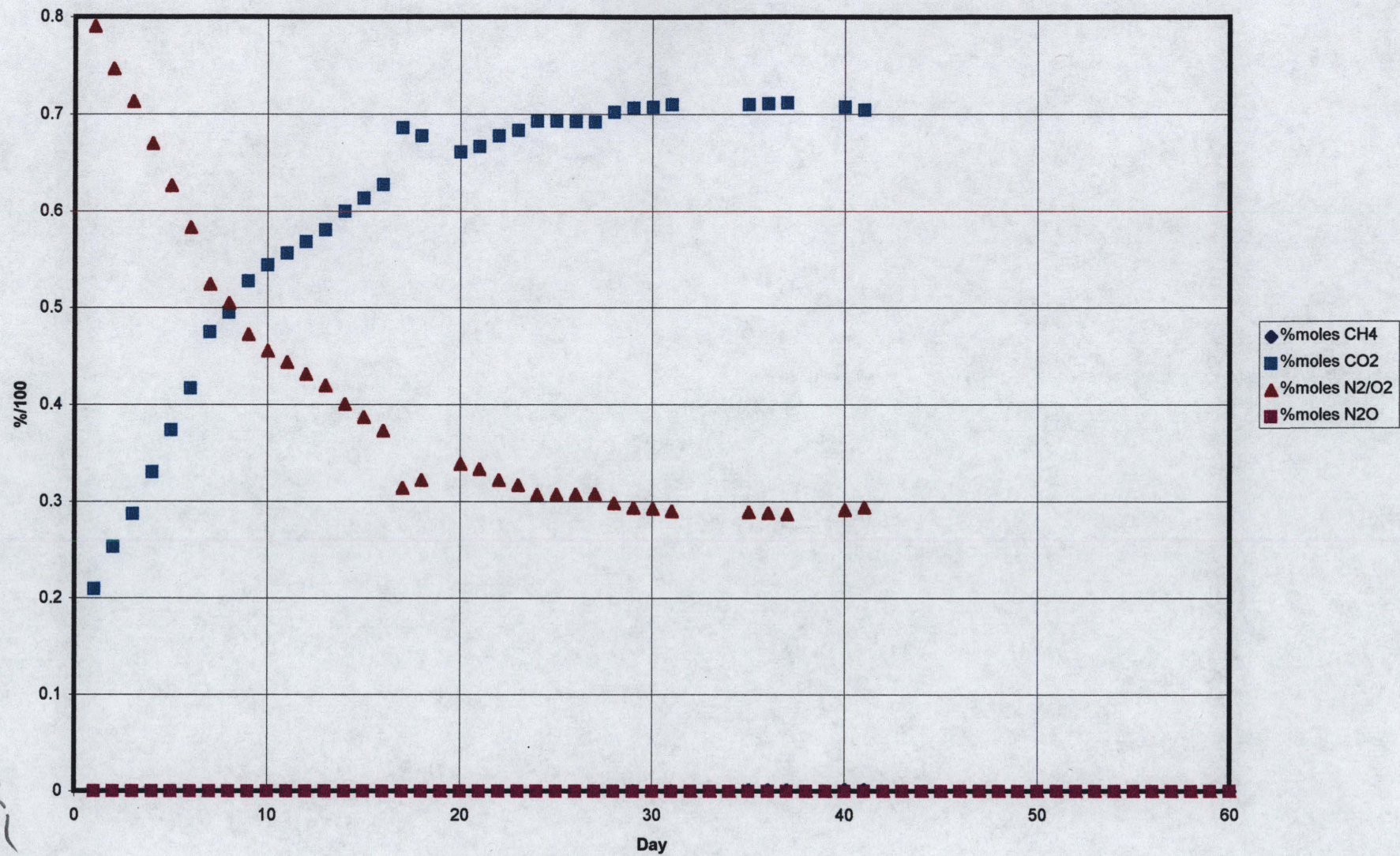


Reactor A14  
Mole Percent/100



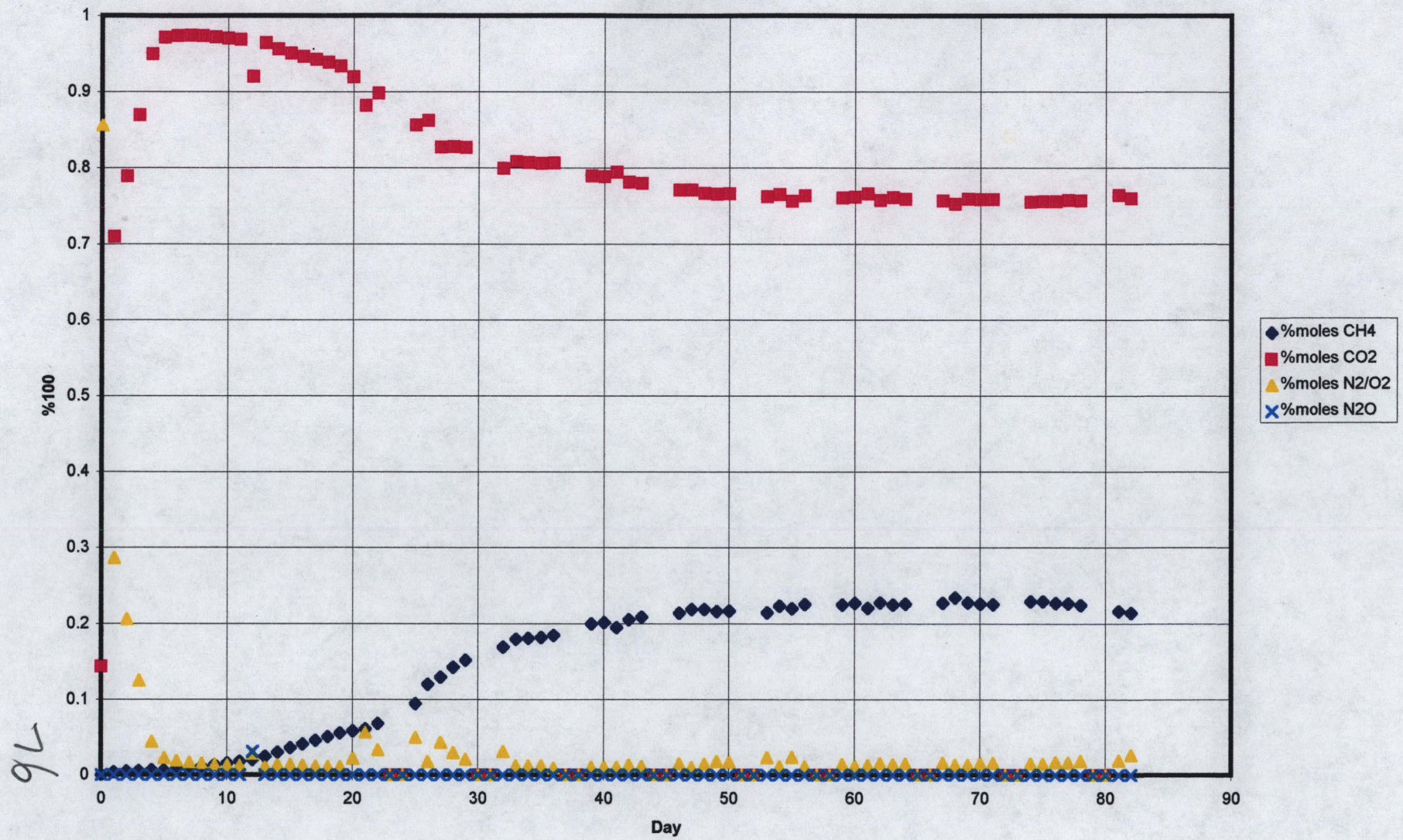


Reactor A15  
Mole Percent/100



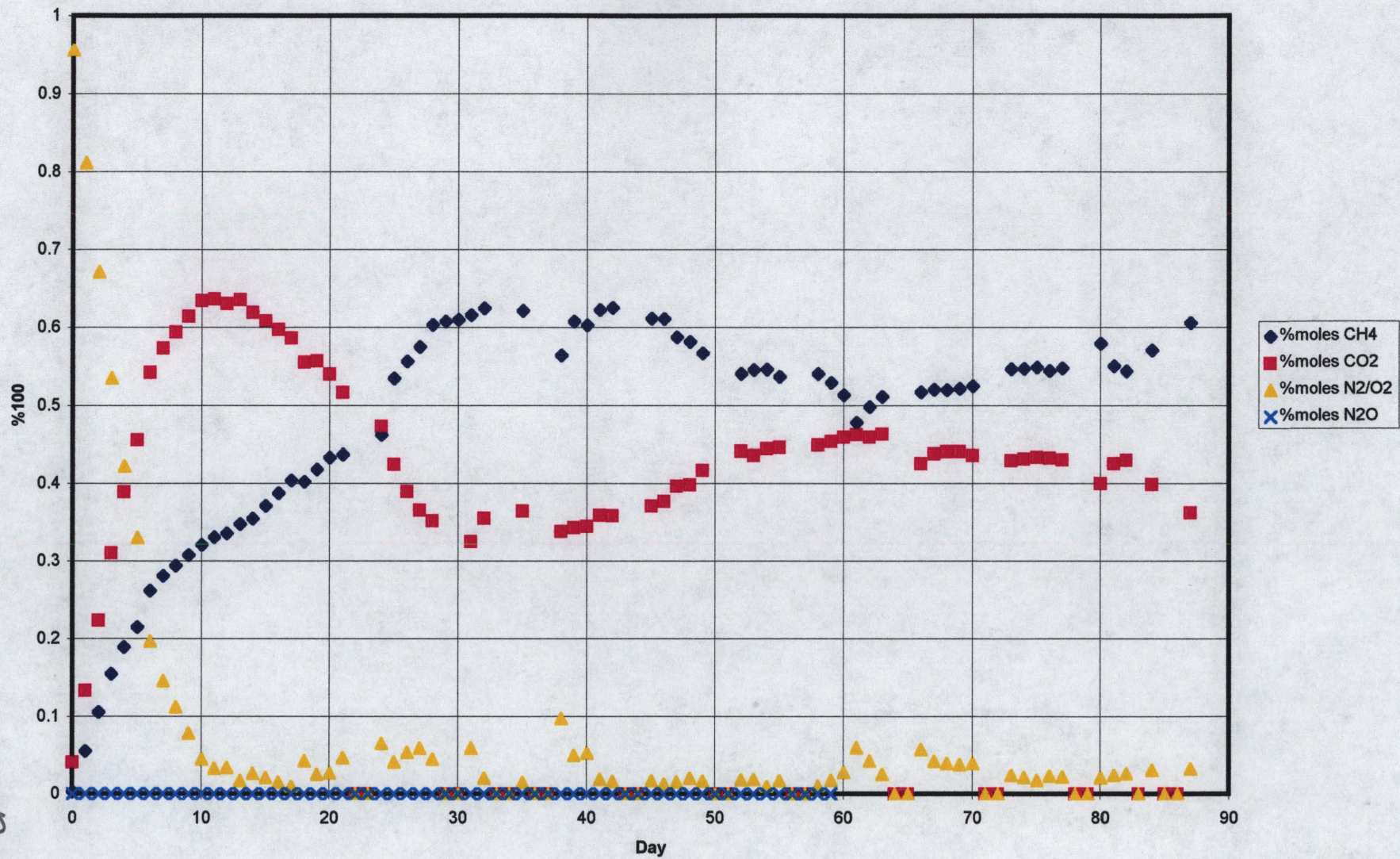


Reactor A16  
Mole Percent/100





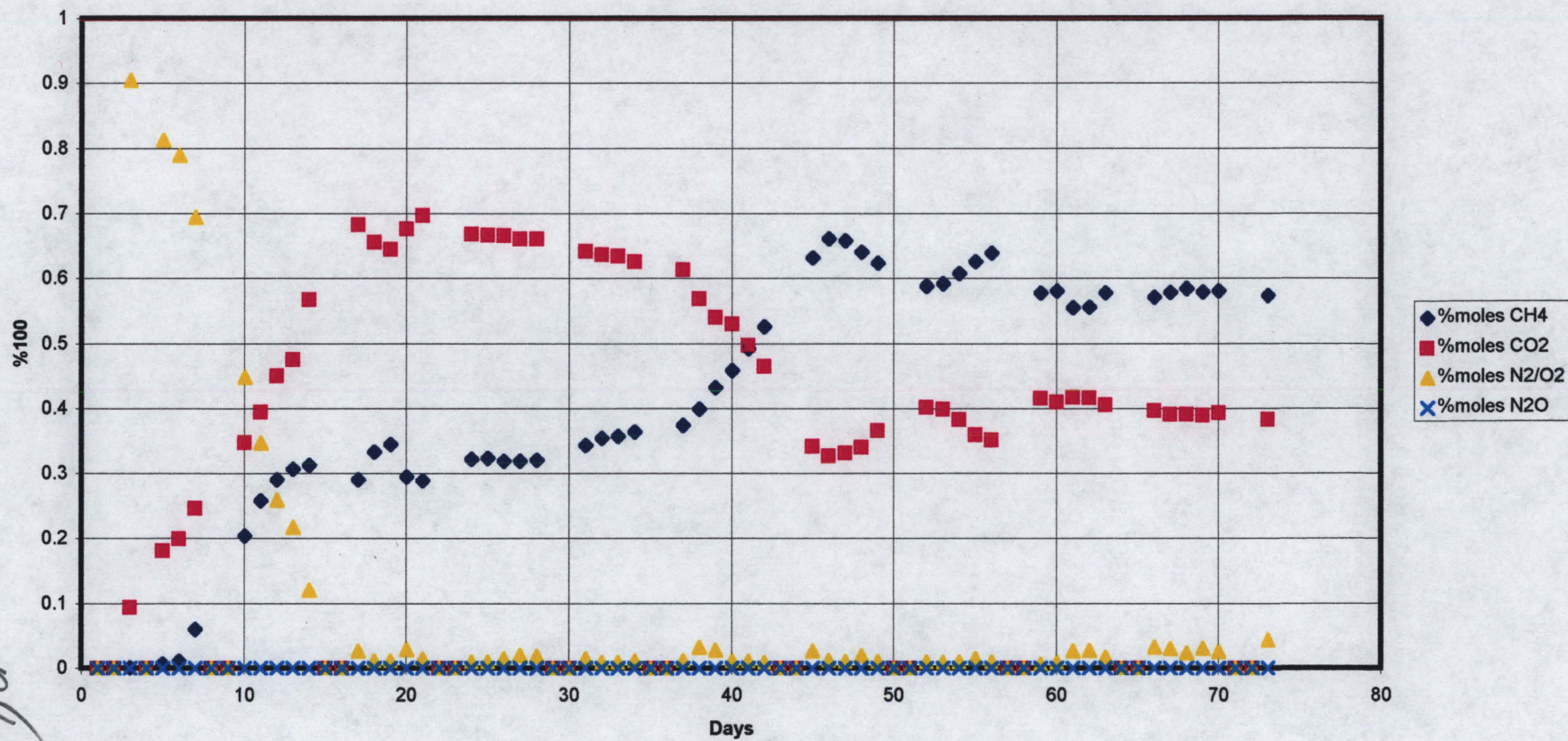
Reactor A18  
Mole Percent/100



66

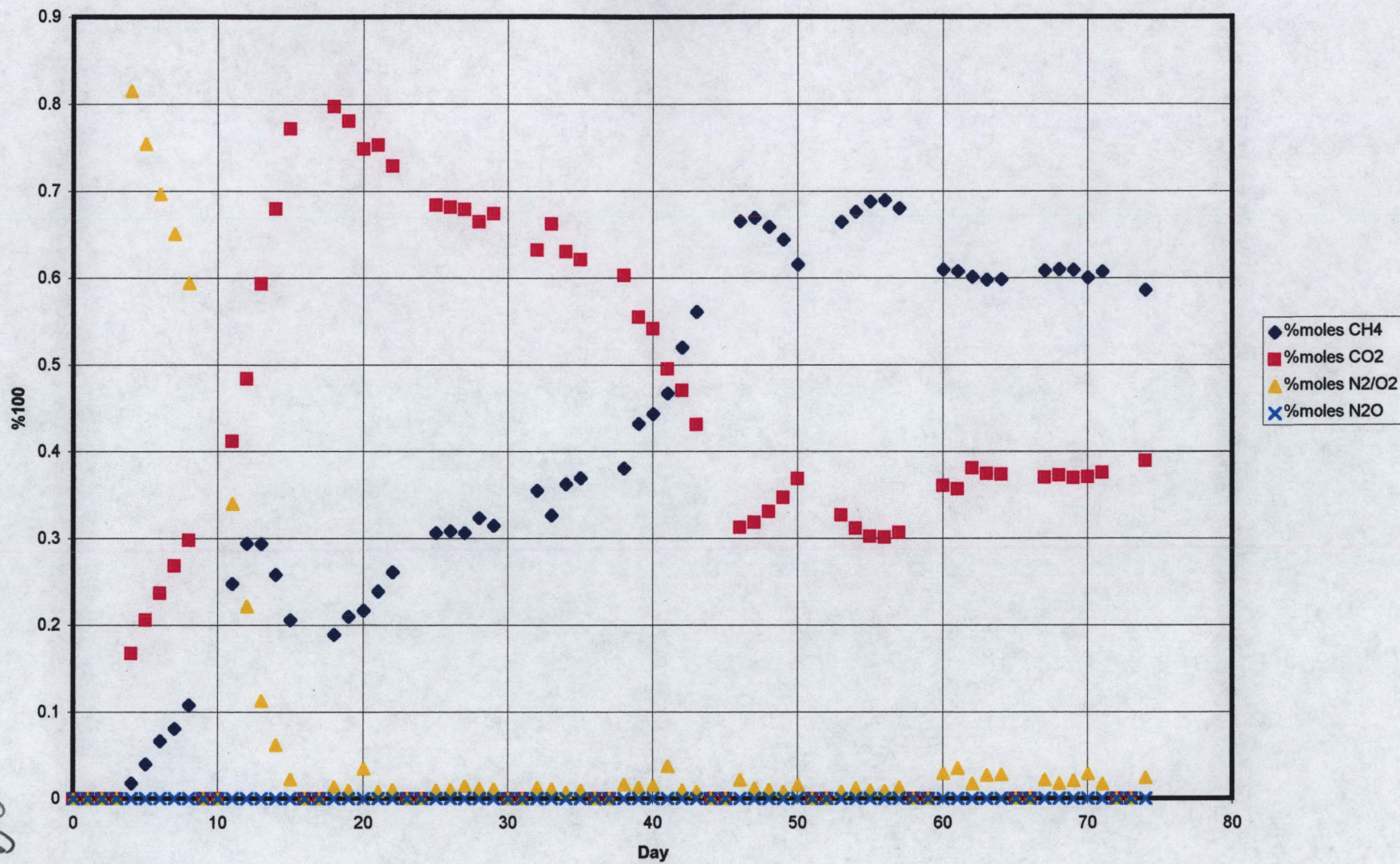


Reactor 19  
Mole Percent/100





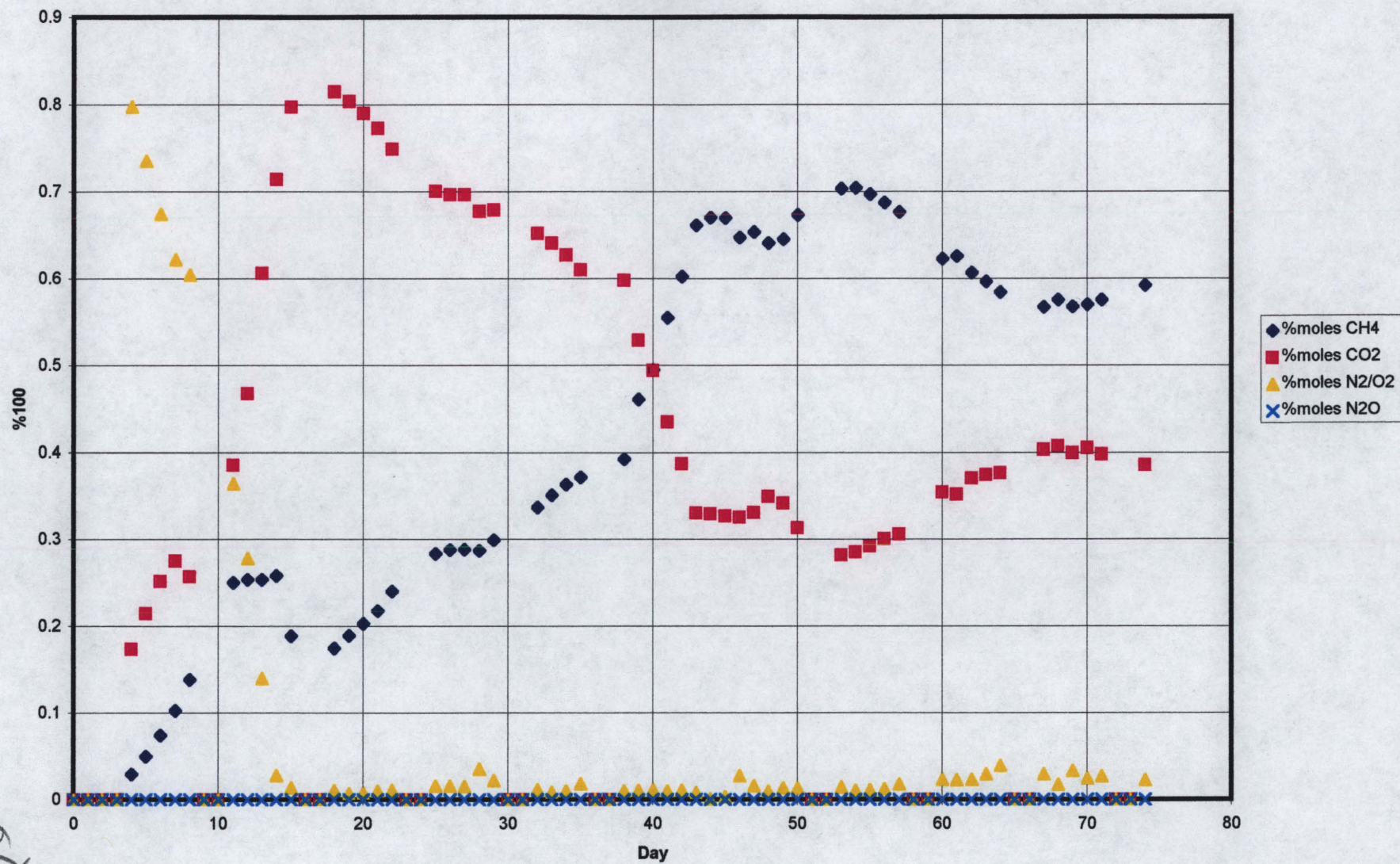
Reactor 20  
Mole Percent/100



66



Reactor 21  
Mole Percent/10



88



Appendix 2

Constituent	Fly Ash	Green Liquor Dregs	Knots	Prim. Sludge	Sec. Sludge	Lime Mud
<u>Macro-Nutrients (%)</u>						
TKN	0.29	0.08	0.10	0.20	0.33	0.10
P	0.08	0.01	0.01	0.07	0.11	0.03
K	0.20	0.77	0.13	0.09	0.12	0.01
Ca	1.50	14.72	0.63	16.85	15.98	35.77
Mg	0.14	0.96	0.07	0.40	0.14	0.82
S	0.03	3.24	0.68	0.30	0.72	0.09
<u>Micro-Nutrients (mg/kg)</u>						
Fe	583.0	10,580.0	504.0	2,886.0	6,347.0	641.0
Mn	511.0	731.0	333.0	1,024.0	1,535.0	82.0
Zn	31.8	1,128.0	59.5	85.5	308.0	33.6
Cu	24.0	177.0	10.6	10.5	37.3	2.7
B	15.3	0.0	3.1	0.7	0.0	0.6
<u>Heavy Metals (mg/kg)</u>						
Ni	2.02	76.25	3.85	5.63	7.27	7.76
Cd	0.00	4.17	0.24	0.00	2.11	0.00
Pb	0.00	48.08	0.00	5.77	18.53	0.00
<u>Other</u>						
% Solids	21.23	45.92	18.92	26.23	36.36	61.04
Sodium (%)	0.02	12.25	2.14	0.36	0.36	0.51
pH	8.34	11.30	10.10	8.95	8.25	9.97
SAR	0.4	83.5	68.2	2.4	2.5	2.3
CaCO <sub>3</sub> eq. (%)	9.75	71.25	7.00	46.5	44.75	97.25

Table 1. NCDA Waste Analysis Laboratory Results (dry weight basis).

Waste Type	TKN (mg/kg)	NH <sub>4</sub> -N (mg/kg)	NO <sub>3</sub> -N (mg/kg)	TOC (mg/kg)	C/N Ratio
Flume Grits	236	15	1	396,000	1,671:1
Wood Yard Debris	1,110	4	2	293,000	263:1

Table 2. Aqua Tech analysis of flume grits and wood yard debris (dry weight basis).



## Calculations to Determine Total Moles Gas Produced

### I. Determine Mass Percent of Individual Gases Using GC

For example,  $\text{CH}_4=20\%$ ,  $\text{N}_2=20\%$ ,  $\text{CO}_2=60\%$

### II. Calculate Mole Percent of Individual Gases

- Divide Mass Percent of Each Gas by Respective Molecular Weight

For example,  $\text{CH}_4 = 0.02 \text{ g} / 16 \text{ g mol}^{-1} = 0.00125 \text{ mol}$

$\text{N}_2 = 0.02 \text{ g} / 28 \text{ g mol}^{-1} = .00071 \text{ mol}$

$\text{CO}_2 = .06 / 44 \text{ g mol}^{-1} = .00136 \text{ mol}$

- Sum Quotients From Above

For example,  $0.00125 + 0.00071 + 0.00136 = 0.00332 \text{ mol}$

- Divide Mass Percent of One Gas by Its Molecular Weight, Divide by Quotient from Above, and Multiply by 100 to Determine Molar Percent

For example,  $\text{CH}_4 = (0.00125 \text{ mol} / 0.00332 \text{ mol}) \times 100 = 37.6\%$

$\text{N}_2 = (0.00071 \text{ mol} / 0.00332 \text{ mol}) \times 100 = 21.4\%$

$\text{CO}_2 = (.00136 \text{ mol} / .00332 \text{ mol}) \times 100 = 41.0\%$

### III. Determine Total Moles Gas Produced Using Formula $n = PV/RT$

- $P$  = pressure generated in the reactor (in atm)

Pressure measured with U-tube of mercury, read as change in mmHg

Measured pressure change divided by 760 mmHg to standardize units (1 mmHg = 1 atm), and corrected for atmospheric pressure by adding 1 atm

For example,  $\Delta \text{mmHg} = 20 \text{ mmHg}$

$P = 20 \text{ mmHg} / 760 \text{ mmHg atm}^{-1} + 1 \text{ atm} = 1.026 \text{ atm}$

- $V$  = volume of gas generated in the reactor (in L)

- Volume was calculated using the formula  $V_1 = (P_2 V_2) / P_1$ , where  $V_1$  = volume generated in the reactor (in L),  $P_2$  = pressure generated in the reactor (in atm),  $V_2$  = volume of reactor (in L), and  $P_1 = 1 \text{ atm}$

- $V = V_1 - V_2 + \text{Change in volume in mercury tube}$

For example,  $V_1 = (1.026 \text{ atm} \times 2.41 \text{ L}) / 1 \text{ atm} = 2.472 \text{ L}$

- Add change in volume in mercury tube, which is  $\pi r^2 \Delta \text{mmHg}$

For example,  $\pi \times (4 \text{ mm})^2 \times 20 \text{ mm} = 1004.8 \text{ mm}^3 = .0010 \text{ L}$

$V = 2.472 \text{ L} - 2.41 \text{ L} + 0.0010 \text{ L} = 0.063 \text{ L}$

- $R$  = Gas Constant =  $0.08205 \text{ L atm deg}^{-1} \text{ mol}^{-1}$

- $T$  = Kelvin Temperature =  $295 \text{ K}$

- Solve for  $n$

For example,  $n = (1.026 \text{ atm} \times 0.063 \text{ L}) / (0.08205 \text{ L atm deg}^{-1} \text{ mol}^{-1} \times 295 \text{ K})$   
 $n = 0.00267 \text{ mol}$

### IV. Multiply Mole Percent of Individual Gas By Total Moles Gas Produced

For example,  $\text{CH}_4 = 37.6\% \times 0.00267 \text{ mol} = 0.00100 \text{ mol}$

$\text{N}_2 = 21.4\% \times 0.00267 \text{ mol} = .00057 \text{ mol}$

$\text{CO}_2 = 41.0\% \times 0.00267 \text{ mol} = .00109 \text{ mol}$



# Volume of Gas Generated in the Reactor

$$P_1V_1 = P_2V_2$$

$$V = V_1 - V_2 + \text{Volume Difference in Mercury Tube}$$

